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Whistler Ecosystems Monitoring Program

2019

Palmer Project #

1602504

Prepared For

Resort Municipality of Whistler

May 20, 2020

May 20, 2020

Ms. Heather Beresford
Environmental Stewardship Manager
Resort Municipality of Whistler
4325 Blackcomb Way
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Dear Ms. Beresford,

Re: Whistler Ecosystems Monitoring Program
Project #: 160254

Enclosed you will find the final Whistler Ecosystems Monitoring Program 2019 report. This report has been authored by Palmer and Snowline.

We hope our team-based approach will not only fulfil the conservation goals for the Resort Municipality of Whistler (RMOW) but will maintain the connection to residents and produce a report that is scientifically defensible.

Thank you for this opportunity to support you on this interesting project. Should you have any questions or require additional information, please feel free to contact Irene Tuite (778) 772 7728 or via email at irene.tuite@pecg.ca

Yours truly,

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Acknowledgements

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Executive Summary

The Resort Municipality of Whistler (RMOW) is located in the southern Coast Mountains of British Columbia, approximately 100 km north of the city of Vancouver. The study area for Ecosystems Monitoring Program (the Program) contains aquatic and terrestrial ecosystems interspersed among areas of urban development.

The RMOW initiated the Ecosystems Monitoring Program in 2013. The program design was based on the use of species, habitat, and climate indicators, to identify temporal and spatial trends in the overall health of ecosystems in the Whistler area. Cascade Environmental Resource Group Ltd conducted the first three years of the Ecosystem Monitoring Program. In 2016, Palmer Environmental Consulting Group Inc. partnered with Snowline Ecological Research and began the next phase of the program. Changes made to the program were designed to maintain comparability and consistency with previous years to the greatest extent possible. The Program continues to evolve as results from past years help direct future monitoring.

The 2019 Ecosystems Monitoring included surveys of benthic invertebrates, fish, Coastal Tailed Frogs (*Ascaphus truei*) and beavers (*Castor canadensis*) as major components. Additional monitoring efforts were directed towards Western Toads (*Anaxyrus boreas*), Northern Goshawks (*Accipiter gentilis laingi*), and black cottonwoods (*Populus trichocarpa*). Complementary monitoring components included water quality, stream temperatures, and climate.

Six stream sites have been established to monitor the aquatic health of streams in the Resort Municipality of Whistler. The 2019 data collection included benthic invertebrate sampling, fish sampling, and *in situ* measurement of water chemistry parameters. In addition, results from RMOW's water and sediment sampling program were screened against the BC water and sediment quality guidelines for protection of aquatic life. *In situ* water quality readings for temperature, dissolved oxygen (DO), pH, turbidity, and conductivity taken concurrently with benthic invertebrate sampling were above minimum guideline values (where available), and readings were within the expected ranges.

Jordan Creek was sampled for both water and sediment in 2019. An exceedance of the BC water quality guideline for calcium was detected at both the upstream and downstream site. Calcium concentrations were 10.7 and 10.6 mg/L, which exceeds the guideline of 8 mg/L. At the upstream site, sediment copper concentration was 35.9 mg/kg (dry weight), which marginally exceeds the BC Interim Sediment Quality Guideline (ISQG) of 35.7 mg/kg.

Crabapple Creek was sampled for both water and sediment quality in 2019. Aluminum concentrations of 0.174 and 0.207 mg/kg were detected, which exceed the BC guideline of 0.1000 mg/kg. Sediment arsenic concentration was 6.45 mg/kg (dry weight) which exceeds the ISQG is 5.9 mg/kg. Sediment copper concentration was 42.9 mg/kg (dry weight) which exceeds the ISQG of 35.7 mg/kg. The benthic invertebrate communities in these streams did not shown impairment that could be attributable to poor water quality. Further monitoring of sediment and water quality in the streams is required to detect if there are any trends in parameter concentrations that could point to deteriorating water or sediment quality.

High proportions (>70%) of pollution sensitive organisms occurred at the upstream site on the River of Golden Dreams, 21-Mile Creek, and Crabapple Creek, which indicates healthy benthic invertebrate communities. The downstream site on the River of Golden Dreams had the lowest proportion of Ephemeroptera, Plecoptera, and Trichoptera (EPT) in 2019 (31%). Jordan Creek, as in previous years, had a relatively low proportion of EPT organisms (44%). The benthic invertebrate data (all years) were analysed using the CABIN Reference Condition Approach. The data were compared to the Fraser Basin 2014 Reference Model. The Bray-Curtis analysis indicated that, in 2019, the downstream site on the River of Golden Dreams was most similar in community structure to reference condition, and the site on 21-Mile Creek was the most dissimilar (i.e. furthest from reference condition).

Three species of fish were captured during 2019 sampling: Threespine Stickleback (*Gasterosteus aculeatus*), undifferentiated trout fry from resident populations of Rainbow Trout (*Oncorhynchus mykiss*) and Coastal Cutthroat Trout (*O. clarkii clarkii*) and sculpin (*Cottus* sp.). Relative condition of fish was generally good in 2019, in keeping with previous years.

The 2019 program continued to expand upon past years' surveys of Coastal Tailed Frogs (*Ascaphus truei*) and. Fifteen sites were again surveyed with a continued emphasis on previously unsurveyed creeks on the west side of Whistler Valley (Van West Creek, Sproatt Creek, and "FJ West Creek"). No evidence of negative impacts was detected at any creek, including the two on Whistler Mountain (Whistler Creek and Archibald Creek).

Mapping irregularities were discovered at lower elevations of the west-side creeks during 2018 tailed frog surveys. Further investigation in 2019 located where Sproatt Creek is diverted underground upstream of the CN Rail tracks (which is why it is dry downstream in low flows). It also confirmed that a branch mapped southeast to Alpha Lake is currently dry since there is no connection with the main stream. Evidence of extensive flooding and stream diversion on FJ West Creek was traced to a storm in November 2017. As a result, the RMOW's stream mapping in that area is no longer correct. Increased flows caused by rain precluded locating any underground diversion of Van West Creek in Function Junction so will require additional field work in 2020.

The 2019 beaver (*Castor canadensis*) survey was the most comprehensive yet. It located 27 active lodges, more than ever before, mainly due to determined efforts to fully access the Miller Creek Wetlands. These additional efforts found eight lodges (compared to two in 2018) and confirmed for the first time that the Miller Creek Wetlands provide beaver habitat as significant as the River of Golden Dreams (where seven lodges were found). Based on number of beavers per lodge, the beaver population is now estimated to be 157 in Whistler Valley (low to high estimates range from 113 to 173 beavers). A field survey of the beaver-affected area in the Miller Creek Wetlands showed that almost double the area visible from air photos was actually flooded. This correction brings the total area of beaver-affected wetlands to 100.3 hectares, approximately two-thirds the total area of wetlands mapped in Whistler Valley, and further demonstrates the importance of beavers in creating habitat that is also critical for many other species. Based on these 2019 results, Whistler's beaver population and the area of beaver-affected wetlands appear to be stable, though larger than previously known.

Three exploratory indicators added to the program in 2018 were again included in 2019: Northern Goshawk (*Accipiter gentilis laingi*), Western Toad (*Anaxyrus boreas*), and black cottonwood (*Populus trichocarpa*). Recent breeding of Northern Goshawks has been sporadically documented since 2011 in Whistler's

unlogged forests at low elevations. There were 11 records of goshawks, including evidence of breeding near Comfortably Numb Trail. Based on a concentration of visual records between Whistler Creekside and Kadenwood, it is possible there was a second pair in 2019 there but no evidence of breeding was reported. No evidence of breeding of Western Toads has been recently found south of Lost Lake, including in 2019. Several new ponds adjacent to Highway 99 in the Callaghan Creek area were confirmed to be suitable for inclusion in 2020 surveys. A plan to improve the mapping of black cottonwoods in Whistler Valley using the RMOW orthophotography and new LIDAR layer was unsuccessful which means field surveys will be needed instead.

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1. Introduction

1.1 Overview

This report describes monitoring studies conducted in 2018 by Palmer and Snowline Ecological Research (Snowline) on aquatic and terrestrial environments in Whistler, British Columbia. The 2019 study was the seventh year of the Ecosystem Monitoring Program and the fourth conducted by this team. The purpose of the program is to monitor the health of ecosystems over time through ecological indicators (proxies) to help guide the conservation of species and ecosystems and inform sustainable land use planning and development in Whistler.

Monitoring is a vital component of ecosystem management. It is therefore important that methods used to establish long-term data collection be done in a scientifically defensible manner. Appropriate selection of indicator species and monitoring methods will provide valuable insight into ecosystem health and functioning.

1.2 Background

The Whistler Biodiversity Project (WBP), funded in significant part by the Resort Municipality of Whistler (RMOW) from 2006 through 2012, began surveys in late 2004. This work led to the first publicly documented record of several important and/or at-risk species. Including Coastal Tailed Frog (*Ascaphus truei*), and Red-legged Frog (*Rana aurora*), initiated the first beaver census, and greatly enhanced the inventory of species documented within Whistler. Early results were summarized in a report (Brett 2007) that recommended further inventory work, as well as the identification and monitoring of indicator species. This work was the precursor to a report the RMOW commissioned that in turn proposed a framework for the establishment and application of ecological monitoring in Whistler (Askey *et al.* 2008).

The Ecosystem Monitoring Program was initiated by the RMOW in 2013. The program design was based on the use of species, habitat, and climate indicators to identify temporal and spatial trends in the overall condition of ecosystems. The initial study design and selection of indicators (Cascade 2014) was based on information from:

- Askey *et al.* (2008) proposed framework;
- Species data collected through the Whistler Biodiversity Project (Brett 2007 and online lists¹); and
- Local data held by Cascade Environmental Resource Group Inc (Cascade).

Cascade was contracted to conduct the first three years from 2013 through 2015 of the Ecosystem Monitoring Program (Cascade 2014-2016). In 2016, PEGC and Snowline were contracted to conduct the program for the following three years. Several changes were made to the study design in 2016 to make it more scientifically robust (e.g. adopting data collection methods which allow for statistical analysis) while maintaining comparability and consistency with previous years to the greatest extent possible. The changes implemented in 2016 included:

- The addition of benthic invertebrates as an indicator for aquatic ecosystem health;
- The use of multiple pass depletion electrofishing methods for fish;

¹ www.whistlerbiodiversity.ca

- Alterations to previously defined species thresholds;
- Changing the methodology for Coastal Tailed Frog surveys from area-constrained to time constrained and increasing the elevational range of study sites on each creek;
- Moving Pileated Woodpecker (*Dryocopus pileatus*) surveys to breeding season and expanding the scope of the cavity tree survey;
- Removal/replacement of some study sites; and
- A return to a full beaver census throughout Whistler Valley.

The work plan has continued to evolve since 2016 as results are evaluated and priorities re-assessed. Some of the main changes made for 2017 and 2018 included:

- The installation of two additional temperature loggers at aquatic sampling sites in Crabapple Creek and 21 Mile Creek;
- Use of the single-pass electrofishing method with no stop nets for fish sampling;
- An increase in the number of tailed frog survey sites, especially on the west side of the valley.
- The first mapping to calculate the area affected by beaver flooding and other activities (“beaver-affected wetlands”).
- Expanded efforts to census the beaver population, notably on the River of Golden Dreams and in the Miller Creek Wetlands.
- The addition of new species of conservation and monitoring value (Northern Goshawks, Western Toads, and black cottonwoods).

Brett (2018) identified monitoring priorities for species and habitats most important to conserving biodiversity within the RMOW’s Development Footprint. Recommendations for the future of the Ecosystem Monitoring Program will build on past results within that context and propose methods to effectively monitor priority species and habitats in the future.

1.3 Study Area

The RMOW is located in the southern Coast Mountains of British Columbia, approximately 100 km north of Vancouver. The study area, defined by the extent of the RMOW municipal boundaries (Figure 1-1), contains a range of aquatic and terrestrial ecosystems at montane to alpine elevations. Most development (within the municipal “Development Footprint”²) is in the valley bottom, from Function Junction to Green Lake.

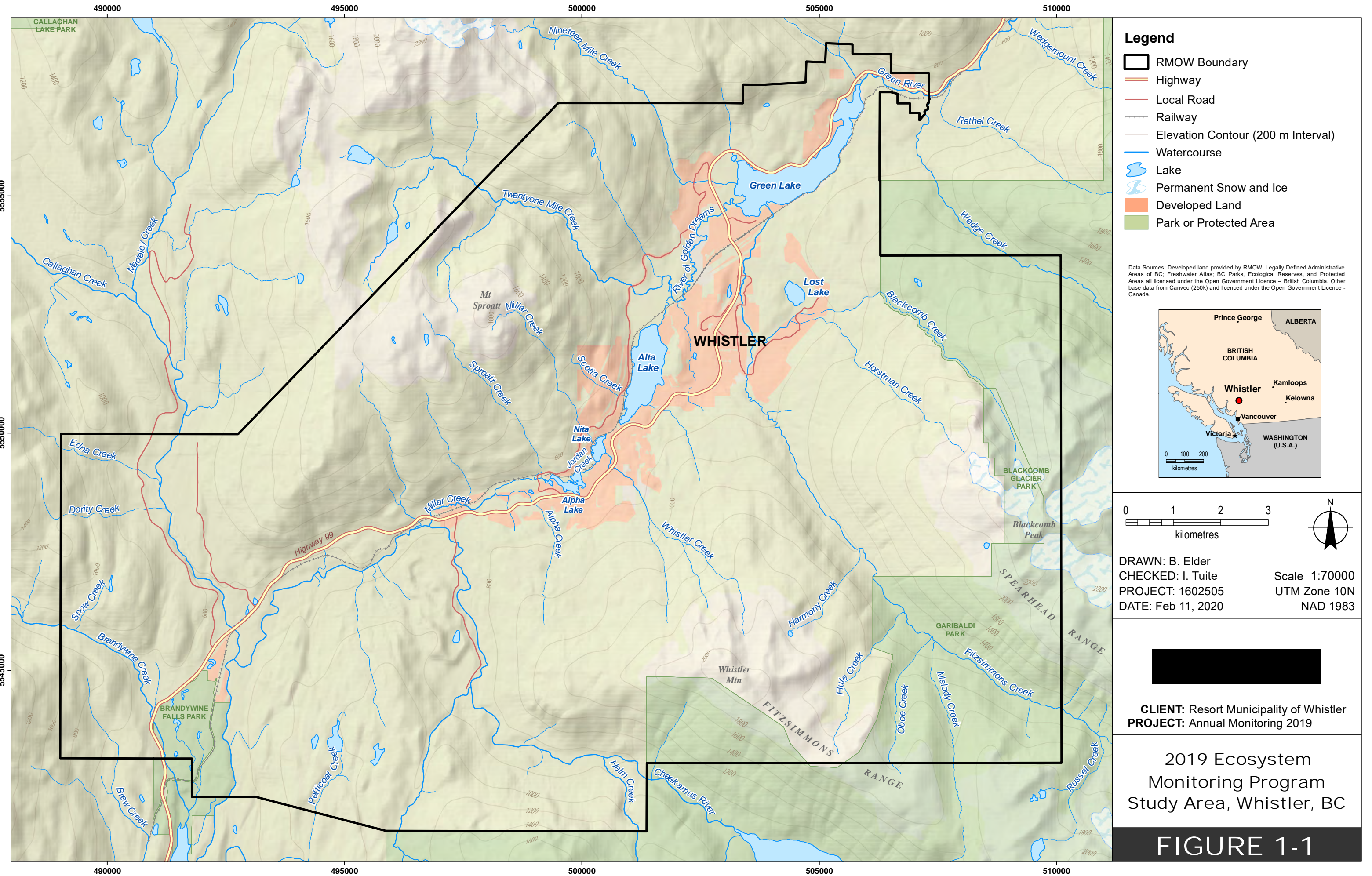
1.4 Study Design

The Ecosystems Monitoring Program is based on the use of indicators that reflect the health of a broader range of populations, taxa, and/or overall ecosystem health. Table 1-1 shows the indicators, field methodologies, and metrics for each program component; detailed study designs are provided in the associated component sections of this report.

² Now termed “Urban Development Containment Area” in the latest draft Official Community Plan (<https://www.whistler.ca/ocp>).

Table 1-1. 2019 Ecosystems Monitoring Program.

| Study Component | Indicator(s) | Methodology/ Equipment | Metrics/Parameters |
|---------------------------|---|---|---|
| Aquatic Habitat | Water Quality | <i>In Situ</i> measurements using a digital meter | <ul style="list-style-type: none"> <i>In Situ</i> parameters: pH, conductivity, dissolved oxygen |
| | Stream Temperature | Temperature loggers set to hourly logging, installed in the study streams | <ul style="list-style-type: none"> Daily and monthly summary statistics for the open water period |
| Aquatic Species | Benthic macroinvertebrate community | CABIN protocol | <ul style="list-style-type: none"> Abundance Taxa richness EPT taxa richness Percentage EPT Diversity indices |
| | Fish | One-pass electrofishing and minnow traps | <ul style="list-style-type: none"> Species identification Fish length to weight relationships Fish Health (Condition) |
| Riparian Species | Coastal Tailed Frog (<i>Ascaphus truei</i>) | Time constrained surveys Malt et al 2014a,b) | <ul style="list-style-type: none"> Tadpole abundance and density Counts of tadpoles by development stage Water temperature and habitat descriptors |
| | Beaver (<i>Castor canadensis</i>) | Field inventories of beaver lodges and activity | <ul style="list-style-type: none"> Number and distribution of active lodges Area of beaver-affected wetland |
| Additional Species | Northern Goshawk (<i>Accipiter gentilis laingi</i>) | Compilation of existing data | <ul style="list-style-type: none"> Documented observations and nest locations |
| | Black cottonwood (<i>Populus trichocarpa</i>) | Previous RMOW mapping | <ul style="list-style-type: none"> Preliminary analysis by area, age, and abundance |
| | Western Toad (<i>Anaxyrus boreas</i>) | Field surveys for presence | <ul style="list-style-type: none"> Presence of tadpoles or metamorphs |
| Climate | Alta Lake freeze and thaw dates | Annual observations by The Point Artists Centre | <ul style="list-style-type: none"> Ice-on and Ice-off dates Days frozen |



2. Stream Water Quality

Lead Biologists: Palmer

2.1 Introduction

The objectives of stream water quality monitoring as a component of the Aquatic Ecosystems Monitoring are to allow consideration for the influence of water chemistry on biological communities and assess water quality by comparing parameter readings to provincial guidelines for the protection of aquatic life. *In situ* water quality measurements included temperature, dissolved oxygen (DO), pH, turbidity, and conductivity. These five analytes describe fundamental characteristics of a water body. Temperature loggers installed in the RMOW streams record hourly readings, which provides a continuous temperature record during the open water season.

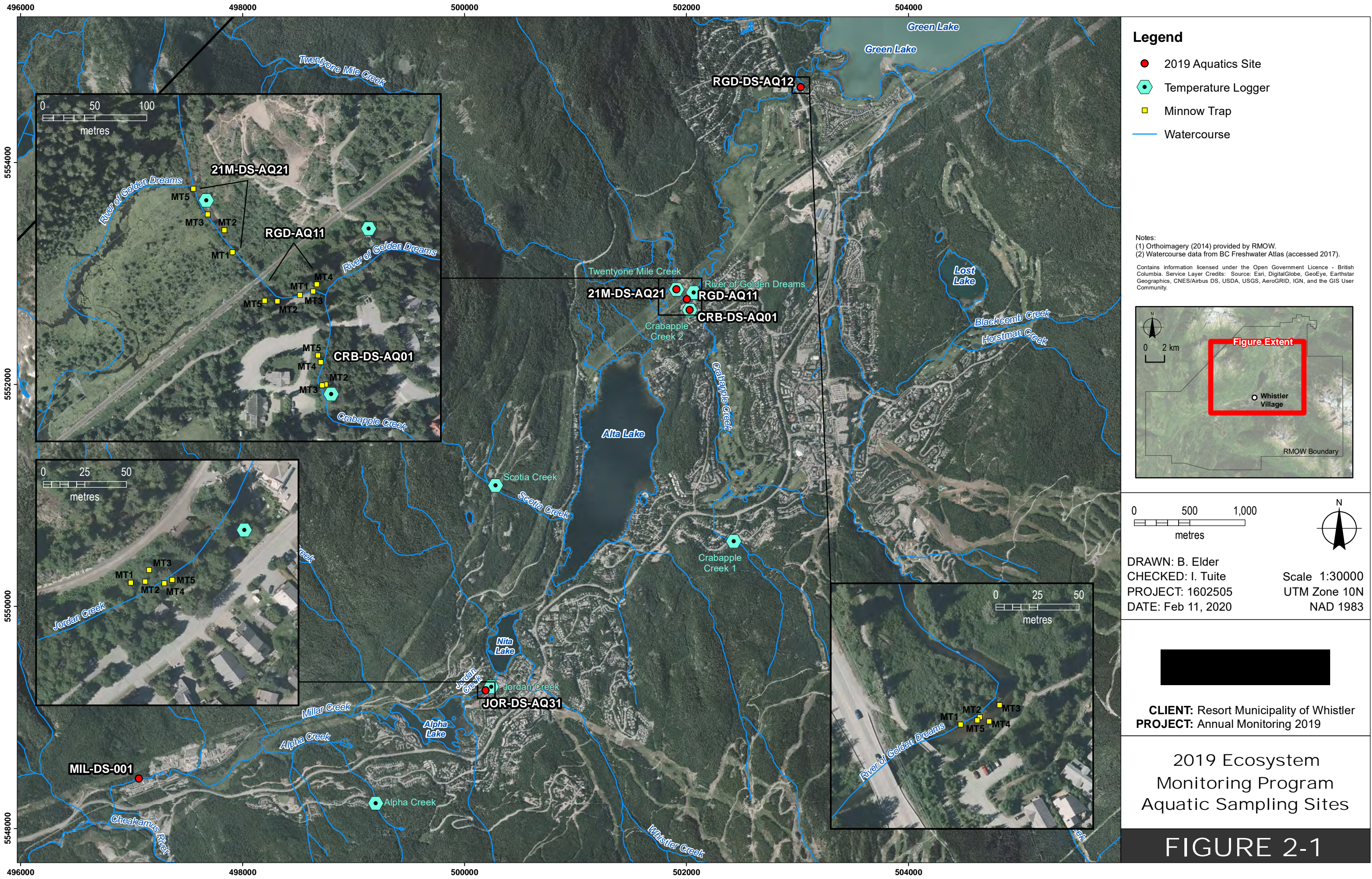
2.2 Methods

2.2.1 In Situ Water Quality

As in previous years, *in situ* water quality parameters and stream temperature were measured in 2019 using a hand-held YSI Pro plus meter, at each of the six established stream sites (Figure 2-1). Measurements were taken concurrently with benthic invertebrate sampling (Table 2-1).

Table 2-1. 2019 Ecosystem Monitoring Program sampling locations and 2019 sampling dates for *in situ* water quality (concurrent with benthic invertebrate sampling).

| Site | UTM Location (Zone 10) | | Aquatic Site ID | Access (Bridge Crossing) | Data Sampled | |
|--------------------------------|------------------------|----------|-----------------|--------------------------------|---|-----------|
| | Easting | Northing | | | In Situ Water Quality & Benthic Invertebrates | Fish |
| Jordan Creek | 500242 | 5549278 | JOR-DS-AQ31 | Lake Placid Road | 31-Jul-19 | 31-Jul-19 |
| Crabapple Creek (2) | 502030 | 5552670 | CRB-DS-AQ01 | Lorimer Road | 30-Jul-19 | 31-Jul-19 |
| River of Golden Dreams (Upper) | 502066 | 5552829 | RGD-US-AQ11 | Lorimer Road | 30-Jul-19 | 1-Aug-19 |
| River of Golden Dreams (Lower) | 503035 | 5554687 | RGD-DS-AQ12 | Off Nicklaus North Golf Course | 31-Jul-19 | 31-Jul-19 |
| 21 Mile Creek | 501910 | 5552856 | 21M-DS-AQ21 | Lorimer Road | 30-Jul-19 | 31-Jul-19 |



2.2.2 Stream Temperature

Five HOBO Water Temperature Pro v2 Data Logger (Model # U22-001) temperature loggers currently record hourly temperatures of stream systems within the RMOW study area (Table 2-2). Five loggers were deployed in December 2015. An additional two loggers were deployed for hourly temperature recordings in August of 2017. The most recently installed (2018) logger locations include 21 Mile Creek and downstream on Crabapple Creek. Table 2-2 lists the location of each temperature logger and the date the loggers were deployed and downloaded. Two temperature logging stations (Crabapple Creek (1) and 21 Mile Creek) were damaged during the spring freshet and not downloaded in 2019.

Daily and monthly summary statistics (means, maxima, and minima) were calculated during the open water period for each creek between August 2018 and July 2019. The temperature time series were examined to identify periods where data were suspect (e.g. elevated readings, when a logger may have been dry), and any suspect data were excluded from the calculations. Mean, minimum and maximum daily stream temperature data from August 2018 to July 2019 can be found in Appendix A.

Table 2-2. Location of Temperature Loggers installed for the Ecosystem Monitoring Program

| Site | UTM Location (Zone 10) | | Location Description | Aquatic Site ID | Access (Bridge Crossing) | Installation Date | Most Recent Download Date |
|---------------------------|---------------------------|----------|---------------------------|--------------------|--------------------------------|----------------------|---------------------------------|
| | Easting | Northing | | | | | |
| Alpha Creek | 499199 | 5548227 | At Tailed Frog Site #1 | - | Spring Creek Drive | 15-Dec-15 | 01-Aug-19 |
| Jordan Creek | 500242 | 5549278 | Near Aquatics Site | JOR-DS- AQ31 | Lake Placid Road | 15-Dec-15 | 31-Jul-19 |
| Scotia Creek | 500280 | 5551092 | At Tailed Frog Site #2 | - | Stone Bridge Drive | 15-Dec-15 | 01-Aug-19 |
| Crabapple Creek (1) | 502426 | 5550589 | At Tailed Frog Site #2 | - | Sunridge Drive | 15-Dec-15 | 03-Aug-18* |
| Crabapple Creek (2) | 502030 | 5552670 | At Aquatics Site | CRB-DS- AQ01 | Lorimer Road | 02-Aug-17 | 31-Jul-19 |
| River of Golden Dreams | 502066 | 5552829 | Near Aquatics Site | RGD-US- AQ11 | Lorimer Road | 15-Dec-15 | 01-Aug-19 |
| 21 Mile Creek | 501910 | 5552856 | At Aquatics Site | 21M-DS- AQ21 | Lorimer Road | 02-Aug-17 | 31-Jul-18** |

* Crabapple Creek (2) temperature logger was not downloaded in 2019

** 21 Mile Creek temperature logger was not downloaded in 2019

2.3 Results and Discussion

2.3.1 Water Quality

2.3.1.1 In Situ Parameters

In situ readings for DO, pH, Specific Conductance and water temperature, (recorded during benthic invertebrate sample collection), are presented in Table 2-3. Results from all years are shown for comparison.

Table 2-3. In situ water quality results, 2016-2019.

| Creek | Site ID | Date | Dissolved oxygen (mg/L) | pH | Specific Conductance (µS/cm) | Water Temperature (°C) |
|--------------------------------|-------------|----------------------|-------------------------|------------|------------------------------|------------------------|
| Jordan Creek | JOR-DS-AQ31 | 03-Aug-2016 | 9.3 | 7.1 | 64 | 15.8 |
| | | 26-Jul-2017 | 8.9* | 7.1 | 105 | 14.9 |
| | | 01-Aug-2018 | 7.7* | 7.1 | 65 | 18.8 |
| | | July 30, 2019 | 9.4 | 7.7 | 78 | 17.4 |
| Crabapple Creek | CRB-DS-AQ01 | 02-Aug-2016 | 9.4 | 7.6 | 218 | 12.7 |
| | | 25-Jul-2017 | 11.6 | 7.4 | 336 | 12.0 |
| | | 01-Aug-2018 | 7.5* | 7.5 | 194 | 16.0 |
| | | July 30, 2019 | 10.0 | 7.6 | 235 | 13.9 |
| 21 Mile Creek | 21M-DS-AQ21 | 03-Aug-2016 | 9.4 | 6.3* | 40 | 12.0 |
| | | 25-Jul-2017 | 11.3 | 7.1 | 40 | 11.6 |
| | | 31-Jul-2018 | 14.6 | 6.2* | 38 | 19.9 |
| | | July 30, 2019 | 9.8 | 7.0 | 52 | 13.3 |
| River of Golden Dreams (Upper) | RGD-US-AQ11 | 03-Aug-2016 | 8.3* | 7.3 | 64 | 11.7 |
| | | 25-Jul-2017 | 11.0 | 7.1 | 50 | 10.5 |
| | | 31-Jul-2018 | 7.5* | 7.2 | 36 | 15.5 |
| | | July 30, 2019 | 9.8 | 6.8 | 33 | 12.8 |
| River of Golden Dreams (Lower) | RGD-DS-AQ12 | 05-Aug-2016 | 9.9 | 7.8 | 69 | 15.2 |
| | | 25-Jul-2017 | 9.8 | 7.0 | 73 | 13.0 |
| | | 01-Aug-2018 | 8.2* | 6.7 | 48 | 17.8 |
| | | July 31, 2019 | 9.9 | 7.6 | 61 | 13.1 |

Notes: 2019 results are **bolded**; values below guideline are identified with an asterisk (*) next to the value.

Dissolved oxygen ranged from 7.5 mg/L to 14.6 mg/L across all sites and years. The BC WQG for dissolved oxygen is an instantaneous minimum of 5 mg/L for all fish life stages other than buried embryo/alevin and 9 mg/L for buried embryo/alevin life stages (BC MOE, 1997). This guideline for dissolved oxygen is not specific to benthic invertebrates, however; low dissolved oxygen can result in reduced benthic invertebrate community diversity. *In situ* dissolved oxygen was above 5 mg/L at all sites in all years. Recorded dissolved oxygen levels in 2019 ranged between 9.4 mg/L and 10.0 mg/L and were thus all above the upper guideline. PH ranged from 6.2 to 7.8 across all sites and years. The BC water quality guideline for pH is 6.5 to 9.0. Readings below 6.5 were recorded at 21 Mile Creek in 2016 and 2018. pH values recorded in 2019 ranged between 6.8 and 7.7 and were thus above the lower guideline.

Specific conductance ranged from 33.3 $\mu\text{S}/\text{cm}$ to 336.3 $\mu\text{S}/\text{cm}$ across all sites and years. The lowest value of 33.3 $\mu\text{S}/\text{cm}$ was recorded at the upper River of Golden Dreams site in 2019. There is no BC guideline for specific conductance. High specific conductance is associated with high dissolved ions. Crabapple Creek had notably higher specific conductance (194.4 to 336.3) than all other sites. Conductance values recorded in 2019 ranged from 33.3 $\mu\text{S}/\text{cm}$ to 234.9 $\mu\text{S}/\text{cm}$. Further monitoring will be required to determine if there are trends present in these parameters.

2.3.1.2 Laboratory Parameters

Laboratory water chemistry data were provided by the RMOW and were reviewed alongside the CABIN sampling results (Table 2-4). The RMOW's water and sediment quality monitoring program rotates streams on an annual basis to achieve a larger coverage of streams in the area.

Table 2-4. Summary of overlap of benthic and water/sediment sampling, 2016-2019.

| Stream | Sample Year | RMOW Water/Sediment Sampling | | | Palmer Benthic invertebrate sampling | | |
|--------------------------------------|-------------|------------------------------|-------|----------|--------------------------------------|--|-------------|
| | | Sample Date | Water | Sediment | Benthic site | Location relative to water/sediment site | Sample Date |
| Jordan Creek (upstream) | 2016 | Oct 12 | x | | JOR-DS-AQ31 | 100 m DS | Aug 03 |
| Jordan Creek (downstream) | 2016 | Oct 12 | x | | JOR-DS-AQ31 | 250 m US | Aug 03 |
| River of Golden Dreams (downstream)* | 2016 | Oct 12 | x | | RGD-US-AQ11, RGD-DS-AQ12 | 100 m US from AQ11, 3.5 km DS from AQ12 | Aug 05 |
| 21 Mile Creek | 2017 | Sept 21 | x | x | 21M-DS-AQ21 | 4 km DS | Jul 25 |
| Crabapple Creek | 2018 | Sept 11 | x | x | CRB-DS-AQ01 | Co-located | Aug 01 |
| Millar Creek | 2018 | Sept 10 | x | x | MIL-DS-001 | Co-located | Aug 01 |
| Crabapple Creek | 2019 | Sept 17 | x | x | CRB-DS-AQ01 | Co-located | Jul 30 |
| Jordan Creek (upstream) | 2019 | Sept 17 | x | x | JOR-DS-AQ31 | 100 m DS | Jul 30 |
| Jordan Creek (downstream) | 2019 | Sept 17 | x | x | JOR-DS-AQ31 | 250 m US | Jul 30 |

Notes: DS = downstream, US = upstream; * indicates that there is another water quality sampling site on the River of Golden Dreams not listed in this table as it is upstream of the confluence with 21 Mile Creek.

Water quality and sediment sampling was conducted by RMOW at the following sites in 2019: Jordan Creek (two locations) and Crabapple Creek. Results were screened against guidelines:

- BC Approved Water Quality Guidelines: Aquatic Life, Wildlife & Agriculture (ENV 2018).

- The Canadian Council of Ministers of Environment (CCME) Canadian Water Quality Guidelines (WQG) for the Protection of Aquatic Life

The CCME WQG, and the BC Approved WQG cover protection of freshwater aquatic life by providing scientifically derived benchmarks for evaluating the potential for observing adverse biological effects in aquatic systems. The BC guidelines are used where BC and CCME guidelines differ, as the BC guidelines are intended to represent more closely the conditions in BC waters, while the CCME (federal) guidelines are more general. In BC, the definition of water quality include the sediments, therefore WQGs include sediment quality values.

Jordan Creek was sampled for both water and sediment quality in 2019. A water quality exceedance for a single metal (calcium) was detected at both upstream and downstream sampling sites as per the BC WQG (BC MOE 1997). A sediment exceedance for a single metal (copper) was detected at the upstream sample site. Sediment copper concentration was 35.9 mg/kg (dry weight) and the BC Interim Sediment Quality Guideline (ISQG) is 35.7 mg/kg.

Crabapple Creek was sampled for both water and sediment quality in 2019. Water quality exceedances in aluminum were detected as per the BC WQG (BC MOE 1997). Concentrations of 0.174 and 0.207 mg/kg were detected, compared to the maximum guideline of 0.1 mg/kg. Sediment exceedances in the following metals were detected as per CCME 2014: arsenic and copper. Sediment arsenic concentration was 6.45 mg/kg (dry weight) and the ISQG is 5.9 mg/kg. Sediment copper concentration was 42.9 mg/kg (dry weight) and the ISQG is 35.7 mg/kg.

The exceedances noted above represent single sampling events and further monitoring is required to assess if there are trends that point to deterioration in sediment or water quality.

2.3.2 Stream Temperature

Stream temperatures were downloaded from five sites within the RMOW in 2019. Crabapple Creek (at Sunridge Drive) and 21 Mile Creek (21M-DS-AQ21) loggers were not downloaded. At 21M-DS-AQ21, water levels prevented the crew from determining if the temperature logger was stuck in the standpipe or lost. Mean monthly stream temperatures in the study streams ranged from -3.86 °C in February (Scotia Creek) to 18.61 °C (Jordan Creek) in August (Figure 2-2). Consistent with previous years, highest temperatures were observed during July and August in all five creeks. Jordan Creek was the warmest creek throughout the year, in keeping with previous years.

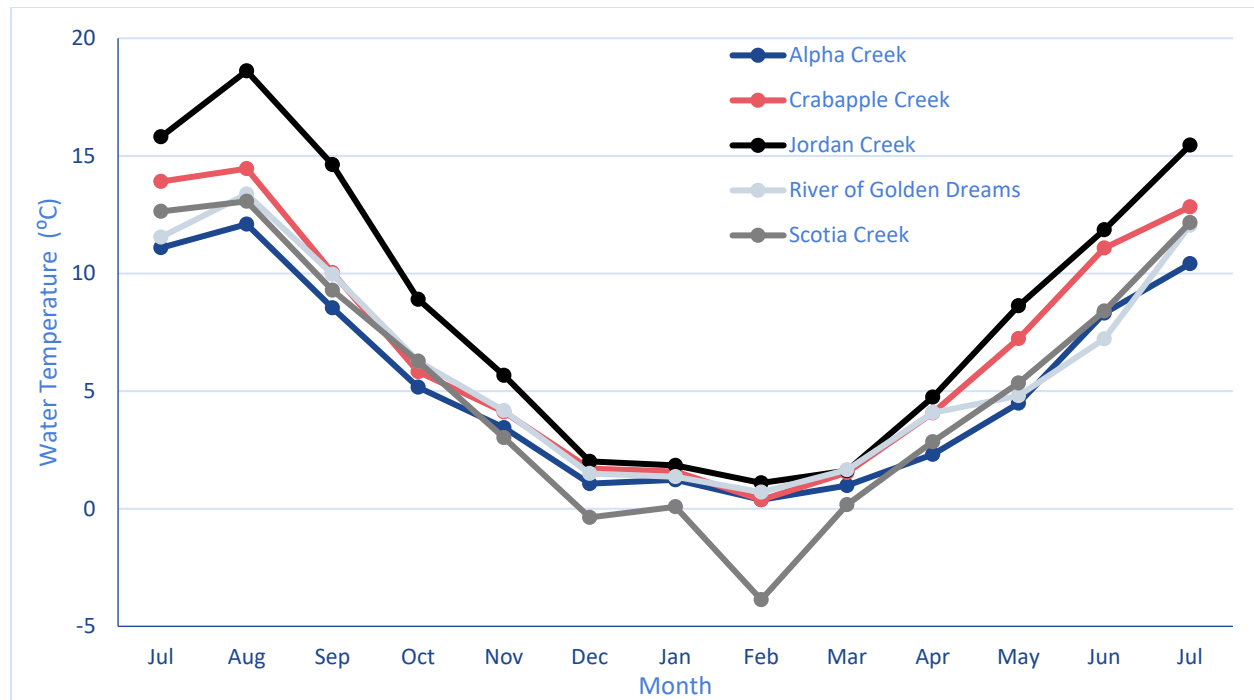


Figure 2-2. Mean monthly stream temperatures, July 2018 – July 2019.

3. Benthic Invertebrates

3.1 Introduction

Benthic invertebrates have been of central importance in biomonitoring studies for many years (Barbour et al. 1999). Advantages of using benthic invertebrates include the following;

- Many have limited migration patterns or a sessile life cycle during their aquatic phase, which means they provide a solid integrated understanding of localized, site-specific conditions.
- Many have a terrestrial winged phase to their life cycle, which means that every year to every few years an entire watershed is recolonized.
- Benthic invertebrates have a complex life-cycle that lasts approximately one year or longer, which means the community integrates the effects of transient, short-term, and seasonal variations.
- Most are relatively easy to identify to family and many taxa can be readily identified to genus or even species, which provides an in depth understanding of community structure.
- Benthic invertebrate communities are diverse and are composed of species that included a range of trophic levels, feeding strategies and pollution tolerances, which provides a comprehensive basis for interpreting community status and environmental effects.
- Sampling protocols are well established and field tested and have minimal detrimental effects on stream communities.
- Benthic invertebrates are abundant in most streams, which means that adequate numbers of organisms for a robust analysis can be easily collected.

Due to their sedentary nature, relatively long lifecycles, abundance, and high community diversity, benthic invertebrate communities provide insight into the long-term health of aquatic ecosystems within a small spatial area (i.e. site).

Benthic invertebrates have been monitored annually in the RMOW study area since 2016 (PALMER and Snowline 2017, 2018, 2019) in four streams: Jordan Creek (JOR-DS-AQ31), Crabapple Creek (CRB-DS-AQ01), River of Golden Dreams (2 sites: RGD-AQ11 and RGD-DS-AQ12), and 21 Mile Creek (21M-DS-AQ21).

3.2 Methods

3.2.1 Benthic Invertebrate Sample Collection

Rapid Bioassessment Protocols (RBP) for streams and wadeable rivers were developed decades ago in response to a need for rapid, cost-effective survey techniques that were nevertheless scientifically valid, easily translatable, and environmentally benign (Barbour et al. 1999). Integral components of an RBP include large composited samples, coordinated habitat characterization, and either multimetric analyses with performance-based evaluation or development of regional reference conditions for benthic invertebrate communities using multivariate ordination (Barbour et al. 1999).

In Canada, Environment and Climate Change Canada has developed a national RBP called the Canadian Aquatic Biomonitoring Network (CABIN) that provides a standardized sampling protocol and a multivariate Reference Condition Approach (RCA) for assessment of benthic invertebrate communities (Barbour et al.

1999, ECCC 2011). As with other RBPs, CABIN includes collection of a composited sample of benthic invertebrates, coordinated habitat characterization, and assessment of the benthic invertebrate community using the RCA.

The Canadian Aquatic Biomonitoring Network (CABIN, ECCC 2011) protocol was performed at five test sites in 2019 (Table 2-1) to collect habitat information and benthic invertebrate samples. The five sites were the same as those used in 2016, 2017 and 2018. Benthic invertebrate sampling was completed prior to fish sampling, to avoid disturbance of the substrate. At each site, a CABIN field sheet was completed, and a benthic invertebrate sample was collected using a kick-net. The CABIN method entails kick-net sampling for benthic invertebrates in the erosional zone (riffle, straight run, or rapid) of a representative watercourse reach. A triangular kick-net sampler with 400-micron mesh and detachable collection cup was employed for each kick-net sample. To collect a sample, one collector walked backward in the upstream direction, tracing a zig-zag pattern, and dragging the net along the bottom. The collector kicked the substrate in front of the net while moving upstream for three minutes. All invertebrates were removed from the net, placed in a clean 500 mL sampling jar, preserved using 85% ethanol and submitted to Cordillera Consulting (Summerland, BC) for taxonomic analysis. In the laboratory, benthic invertebrates were identified to the lowest possible taxonomic group by Cordillera.

Samples from site RGD-US-AQ11 were sieved using the bucket swirling method to remove excess debris from the samples (ECCC 2011). A QA/QC sample was collected from the remaining debris at this location. The sample of excess debris was processed in the laboratory to ensure that the method was effective in removing benthic invertebrates.

Once the kick-sample was collected, habitat characteristics were recorded at each site including canopy coverage, macrophyte coverage, riparian vegetation, periphyton coverage, substrate composition (pebble count) and slope (Appendix C). Average and maximum velocity were determined by measuring velocity at 6 points along a transect of the stream using the Velocity Head Rod technique, according to CABIN protocol (ECCC 2011).

3.2.2 Laboratory Analysis

Cordillera Consulting identified and enumerated organisms to the genus-species level, where possible. Enumeration was undertaken using a Marchant box: cells were extracted and enumerated in the order indicated by a random number table. Sorting and counting continued until the 300th organism was identified. If the 300th organism was found part way into sorting a cell, then the balance of the cell was sorted. If the organism count had not reached 300 by the 50th cell then the entire sample was sorted (Appendix B). Organisms were identified to the lowest practical level using Standard Taxonomic Effort lists compiled by the CABIN manual (McDermott et al. 2014, SAFIT 2015 and PNAMP 2015).

The 2019 benthic invertebrate taxonomic richness was reported as number of families, the standard protocol for CABIN reports that accounts for potential misidentification of invertebrates at lower taxonomic levels (e.g. genus or species level). Organisms were grouped as follows: Ephemeroptera (mayflies), Plecoptera (stoneflies), Trichoptera (caddisflies), Diptera (true flies) +non-insects, and other. The grouping of Diptera+non-insects includes true flies, bivalves, molluscs, mites and worms.

3.2.2.1 Quality Assurance/Quality Control

Cordillera Consulting has over ten years' experience in taxonomic analysis of benthic invertebrates from streams, rivers and lakes of western Canada. The following QA/QC procedures were followed by Cordillera Consulting:

- Complete, blind re-identification and re-enumeration was completed in-house by a second taxonomist (*i.e.* not the taxonomist who originally processed the samples)
- Samples for taxonomic quality control were randomly selected and quality control procedures were conducted as the samples progressed through the laboratories.
- The second taxonomist calculated and recorded four types of errors:
 - Misidentification error;
 - Enumeration error;
 - Questionable taxonomic resolution error; and
 - Insufficient taxonomic resolution error.

The percent total identification error rate was calculated as:

$$(\text{Sum of incorrect identifications} \div \text{total organisms counted in audit}) \times 100$$

The average identification error rate of audited samples did not exceed 5%. All samples that exceeded a 5% error rate were re-evaluated to determine whether repeated errors or patterns in error contributed (Appendix B).

3.2.3 Data Analysis

3.2.3.1 CABIN Multivariate Reference Condition Approach and Assessment

The 2019 benthic invertebrate sampling results (habitat and taxonomy data) were entered into the online CABIN database. Data from 2016 to 2019 sampling are stored in the database for ease of access, data security and to allow CABIN analyses to be performed. The benthic invertebrate data were analysed using the Reference Condition Approach (RCA) adopted from Environment Canada's CABIN protocols (ECCC 2011; Palmer and Snowline 2017, 2018). In 2016 and 2017, the benthic invertebrate data were compared with the Fraser River-Georgia Basin Reference Model (2005) to make this assessment. However, in 2018 the Fraser River-Georgia Basin Reference Model error rates of correctly assigning a site to the appropriate reference group were noted to be unacceptably high. Environment Canada recommended conducting the RCA analysis (for all years and sites) using the updated Fraser River Reference Model (2014) to provide more reliable results. Accordingly, for this report, the data (all years) were compared with the Fraser Basin Reference Model, a long-standing reference model first developed in 1999 (Rosenberg et al. 1999), updated in 2005 (Sylvestre et al. 2005), and updated again in 2014 (Strachan et al. 2014).

The model assigned each site to a reference group based on habitat variables as well as the type and proportion of taxa present (Sylvestre *et al.* 2005). The samples from the five sites and four years were assigned to one of a total of four reference groups (Table 3-1).

The multivariate ordination (Appendix B) used in the RCA was developed using Bray-Curtis Index (BCI) data calculated for the RCA as a complete data matrix. For the test sites, the Bray-Curtis Index (BCI) was

then calculated based on the expected relative abundance of the taxa present for that reference group; these BCI data were then used to locate each site on the ordination.

For the BCI, a value of 0 indicates that a site is identical in community structure to the reference condition and a value of 1 indicates a site is entirely different from the reference condition with no species in common. Within that range, between site variability is considered low if BCI values are less than 0.40 moderate if BCI values are between 0.40 and 0.80, and high if BCI values are greater than 0.80. The latter category is also problematic because the correlation between BCI values and ecological 'distance' becomes sharply non-linear above approximately 0.80. (Beals 1984). Site comparisons with BCI values greater than 0.80 should therefore be interpreted with caution. For the reference sites, the mean BCI values ranged from 0.41 to 0.55 and were therefore considered moderately variable on average (Table 3-1).

The CABIN analysis provided an assessment of whether test sites were in reference condition, mildly divergent from reference condition, or divergent from reference condition. The assessment was further developed through comparison of test sites with reference sites using the River Invertebrate Prediction and Classification System (RIVPACS). The RIVPACS compares the observed taxon richness with the taxon richness predicted from the reference model, reported as an Observed:Expected (O:E) ratio. A ratio less than one indicates fewer taxa than expected and a ratio greater than one indicates more taxa than expected. From an assessment perspective, it is considered that impairment would result in a loss of taxa richness and therefore O:E ratios less than one. For the CABIN assessment, however, divergence would result from either a high or low O:E ratio.

3.2.3.2 *Benthic Invertebrate Community Characterization*

To further characterize the benthic invertebrate community, the following metrics were calculated:

- Abundance, calculated as the total number of individuals per kick-net per site;
- EPT relative abundance, calculated by dividing the abundance of mayfly (Ephemeroptera), stonefly (Plecoptera) and caddisfly (Trichoptera) by the total abundance. These three orders of aquatic insects are typically most sensitive to habitat disturbance;
- Taxa richness, calculated as the total number of families present at each site;
- EPT taxa richness, defined as the total number of mayfly (Ephemeroptera), stonefly (Plecoptera) and caddisfly (Trichoptera) families per site;
- Percentage composition, calculated by dividing the abundance of the five most dominant groups by the total abundance; and
- Shannon-Wiener diversity index, defined as: $H' = \sum_{i=1}^R p_i (\ln p_i)$
- Where R is taxa richness and p_i is the total number of individuals in the i^{th} species divided by the total number of organisms in the sample. The Shannon-Wiener diversity index characterizes taxa diversity in a community and accounts for taxa richness as well as the proportion of each taxa (evenness).

3.2.3.3 *Multimetric Performance-Based Assessment*

There are two approaches to development of Rapid Bioassessment Protocols; the multivariate approach used in Canada, and the multimetric approach used in the United States (Barbour 1999). The advantage of the multivariate approach is that it uses all the data to provide a comprehensive assessment. The

disadvantage is that the multivariate approach assesses whether a test site is divergent from reference, but not how it is divergent. In contrast, the multimetric performance-based approach uses only a fraction of the available data, but if the metrics are chosen with care, provides a solid understanding of how sites are divergent. For the test sites at Whistler, use of the multimetric approach in addition to the multivariate approach was considered useful in providing a weight-of-evidence to test site classification. There are five steps to the multimetric approach (Barbour 1999);

1. Use biological information and habitat data to group reference sites into homogenous classes. This step is identical for both the multivariate and multimetric approaches.
2. Identify candidate attributes of the benthic invertebrate community that are ecologically relevant.
3. Select core metrics that are sensitive to watershed stressors and are informative of the relationship of the benthic community to specific stressors or cumulative impacts. The six core metrics selected for this report were taxon richness, EPT richness, EPT relative abundance, total abundance, dominant taxa (%), and the Shannon-Wiener Diversity Index (Table 3-1).
4. Transform the core metrics to dimensionless numbers for aggregation. For this study, transformation was undertaken by dividing the test site metric value by the mean of the reference site metric. This resulted in each metric score being expressed as a fraction.
5. Establish thresholds of impairment. For the purposes of this assessment, it was considered that the final average assessment values were within reference condition (i.e. unimpaired) if the calculated metric score was greater than 0.75. This was based on the median coefficient of variation (standard deviation/mean) of the reference metrics, which was calculated as 25% (Appendix B). It is noted, however, that the coefficient of variation was variable amongst the six metrics, with Shannon-Wiener Diversity having the lowest variability and total abundance having the highest variability (Table 3-1). Consequently, it was considered that the individual metrics were considered within reference if the metric score was within one standard deviation of the mean and mildly divergent if outside of one standard deviation. Similarly, individual metrics were considered divergent from reference if the metric score was greater than two standard deviations from the reference mean. This latter threshold is consistent with the Critical Effect Size as defined in the Environmental Effects Monitoring program within the Metal and Diamond Mining Effluent Regulations under the *Fisheries Act*, and is a de facto standard for benthic invertebrate analyses within Canada.

Table 3-1. Characteristics of the Groups within the Fraser River Basin Model (Strachan et al. 2014)

| Parameter | Group | | | |
|----------------------------|--------------------|-------------------|--------------------|--------------------|
| | 1 | 3* | 4 | 5* |
| Number of Sites | 64 | 19 | 103 | 13 |
| Bray-Curtis Index | 0.48 ± 0.15 (31%) | 0.41 ± 0.17 (42%) | 0.53 ± 0.14 (26%) | 0.55 ± 0.22 (40%) |
| Total Abundance | 5011 ± 6542 (131%) | 3776 ± 2948 (78%) | 2647 ± 2773 (105%) | 13707 ± 8626 (63%) |
| EPT Relative Abundance (%) | 79.0 ± 14.8 (19%) | 78.2 ± 17.8 (23%) | 66.1 ± 26.2 (40%) | 49.6 ± 26.3 (53%) |
| Taxon Richness | 16.8 ± 4.7 (28%) | 14.8 ± 4.3 (29%) | 18.0 ± 4.5 (25%) | 16.0 ± 4.0 (25%) |
| EPT Richness | 11.0 ± 2.8 (25%) | 9.8 ± 2.6 (27%) | 10.8 ± 3.5 (32%) | 9.3 ± 3.6 (39%) |
| Five Dominant (%) | 83.1 ± 9.3 (11%) | 86.1 ± 8.2 (10%) | 82.2 ± 8.7 (11%) | 86.1 ± 8.4 (10%) |
| Shannon-Wiener Diversity | 1.9 ± 0.4 (21%) | 1.8 ± 0.4 (22%) | 1.9 ± 0.4 (21%) | 1.7 ± 0.4 (24%) |

(Coefficient of Variation)

**The minimum recommended number of sites (20) for the RCA is not satisfied for Group 3 or Group 5 (Bowman and Somers 2005).*

3.2.3.4 Hilsenhoff Index of Biotic Integrity

As a further test of the benthic invertebrate community, the Hilsenhoff Index of Biotic Integrity (HIBI) was calculated and assessed for the 2018 and 2019 benthic invertebrate data.

The HIBI is calculated using tolerance scores and relative abundance data for the benthic invertebrate community. The tolerance scores have been developed over time by experts and relate to the response of benthic invertebrates to organic pollution (Mandeville 2002). The HIBI scores range from 0 to 10, with a score of 0 indicating that a site is dominated by pollution-sensitive organisms and that there is no apparent organic pollution at the site (Table 3-2).

The HIBI is of interest because of the potential for organic pollution in an urban setting, including from stormwater runoff, septic tank leakage, industrial activity, and/or wildlife waste.

Table 3-2. Hilsenhoff Index of Biotic Integrity (Mandaville 2002)

| Assessment | Extent of Organic Pollution | Low HIBI Score | High HIBI Score |
|-------------|-----------------------------|----------------|-----------------|
| Excellent | None apparent | 0 | 3.5 |
| Very Good | Possible | 3.51 | 4.5 |
| Good | Some | 4.51 | 5.5 |
| Fair | Fairly significant | 5.51 | 6.5 |
| Fairly Poor | Significant | 6.51 | 7.5 |
| Poor | Very significant | 7.51 | 8.5 |
| Very Poor | Severe | 8.51 | 10 |

3.2.3.5 Temporal Trends Analysis

There are currently four years of benthic invertebrate data, which provides a sample size of only six temporal comparisons. These are too few for a robust analysis and so trends analysis was not undertaken this year. However, five years of data will provide a sample size of ten temporal comparisons. Temporal trends analysis using the non-parametric Mann-Kendall temporal trends test will therefore be undertaken starting after year 5 of the monitoring program.

3.3 Results and Discussion

3.3.1 Group Assignment

Based on the habitat and the type and proportion of taxa present at each site there were three sites/years assigned to Group 1, six to Group 3, five to Group 4, and six to Group 5 (Table 3-3). Confidence in the Group assignments, however, was low for the following reasons;

- Only site JOR-DS-AQ31 (Jordan Creek) was consistently classified through all four years of monitoring with the other three sites assigned to two different Groups each (Table 3-3).
- The probability of group membership was less than 50% for all but six sites/years (Table 3-3), with the lowest probability estimated at only 26%.
- Group 3 and Group 5, which were assigned to 12 of the 20 year/site combinations, are currently defined by less than 20 reference sites (Table 3-1). This is less than recommended for development of a robust understanding of reference condition (Bowman and Somers 2005).

Table 3-3. Probabilities of sorting into each reference model group (based on habitat), for aquatic sampling sites, Whistler, 2016 - 2019

| Site | Year | Group | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 |
|---|------|-------|------------|---------|------------|------------|------------|---------|
| Twenty-one Mile Creek (21M-DS-AQ21) | 2016 | 3 | 6% | 4% | <u>29%</u> | 29% | 21% | 12% |
| | 2017 | 3 | 10% | 5% | <u>33%</u> | 24% | 17% | 10% |
| | 2018 | 5 | 10% | 5% | 22% | 17% | <u>39%</u> | 6% |
| | 2019 | 3 | 10% | 5% | <u>33%</u> | 24% | 17% | 10% |
| Crabapple Creek (CRB-DS-AQ01) | 2016 | 1 | <u>44%</u> | 27% | 0% | 19% | 9% | 2% |
| | 2017 | 1 | <u>44%</u> | 27% | 0% | 19% | 8% | 2% |
| | 2018 | 1 | <u>44%</u> | 27% | 0% | 19% | 8% | 2% |
| | 2019 | 5 | 0% | 0% | 0% | 16% | <u>82%</u> | 1% |
| Jordan Creek (JOR-DS-AQ31) | 2016 | 4 | 14% | 8% | 0% | <u>55%</u> | 2% | 20% |
| | 2017 | 4 | 18% | 10% | 0% | <u>51%</u> | 2% | 18% |
| | 2018 | 4 | 10% | 7% | 0% | <u>57%</u> | 7% | 18% |
| | 2019 | 4 | 8% | 6% | 0% | <u>62%</u> | 3% | 21% |
| River of Golden Dreams (RGD-US-AQ11) | 2016 | 3 | 9% | 5% | <u>38%</u> | 22% | 17% | 10% |
| | 2017 | 3 | 8% | 4% | <u>41%</u> | 21% | 16% | 10% |
| | 2018 | 5 | 9% | 4% | 27% | 16% | <u>38%</u> | 7% |
| | 2019 | 3 | 9% | 5% | <u>39%</u> | 22% | 17% | 10% |
| River of Golden Dreams (RGD-DS-AQ12) | 2016 | 4 | 17% | 8% | 16% | <u>27%</u> | 23% | 9% |
| | 2017 | 5 | 16% | 7% | 10% | 17% | <u>46%</u> | 5% |
| | 2018 | 5 | 12% | 4% | 5% | 8% | <u>68%</u> | 2% |
| | 2019 | 5 | 18% | 7% | 10% | 16% | <u>44%</u> | 5% |

3.3.2 Multivariate Site Assessment

The test site BCI values ranged from 0.37 to 0.94 with a first quartile of 0.57 and a median value of 0.72 (Table 3-4). The BCI values for most of the test sites were therefore greater than the highest value of 0.55 for the reference sites (Table 3-1). These data indicate that the benthic invertebrate communities at test sites were moderately dissimilar to reference sites.

Based on the reference and test site BCI values, Twenty-one Mile Creek (21M-DS-AQ21) was assessed as being mildly divergent to divergent with the difference between reference and test site BCI of between 0.32 and 0.37, Crabapple Creek (CRB-DS-AQ01) was assessed as being in reference condition to mildly divergent with the difference between reference and test site BCI of between 0.05 and 0.23, Jordan Creek (JOR-DS-AQ31) was assessed as being in reference condition to divergent with the difference between reference and test site BCI of between 0.04 and 0.25, River of Golden Dreams upstream site (RGD-US-AQ11) was assessed as being mildly divergent to divergent with the difference between reference and test site BCI of between 0.29 and 0.39, and River of Golden Dreams downstream site (RGD-DS-AQ12) was assessed as being in reference condition with the difference between reference and test site BCI of between 0.04 and 0.17 (Table 3-4).

Correlation of the difference in BCI between test site and reference site and the subsequent assessment indicated that the difference between the reference and test BCI values was <0.18 for test sites in reference condition, between 0.17 and 0.34 for sites considered mildly divergent, and >0.25 for test sites considered divergent from reference condition (Table 3-4, Appendix B).

Table 3-4. Bray - Curtis Index for aquatic sampling sites, Whistler, 2016 -2019

| Site | Year | Test Site Bray-Curtis Index | Reference Bray-Curtis Index (Mean \pm SD) | RIVPACS O:E (p>0.7) | Group (Probability) | CABIN Classification |
|-------------|------|-----------------------------|---|---------------------|---------------------|----------------------|
| 21M-DS-AQ21 | 2016 | 0.74 | 0.41 \pm 0.17 | 1.17 | 3 (29%) | Mildly Divergent |
| | 2017 | 0.78 | 0.41 \pm 0.17 | 0.93 | 3 (33%) | Divergent |
| | 2018 | 0.87 | 0.55 \pm 0.22 | 1.20 | 5 (39%) | Mildly Divergent |
| | 2019 | 0.75 | 0.41 \pm 0.17 | 1.16 | 3 (33%) | Mildly Divergent |
| CRB-DS-AQ01 | 2016 | 0.71 | 0.48 \pm 0.15 | 0.96 | 1 (44%) | Mildly Divergent |
| | 2017 | 0.37 | 0.48 \pm 0.15 | 0.96 | 1 (44%) | Reference |
| | 2018 | 0.43 | 0.48 \pm 0.15 | 1.15 | 1 (44%) | Reference |
| | 2019 | 0.72 | 0.55 \pm 0.22 | 0.56 | 5 (82%) | Mildly Divergent |
| JOR-DS-AQ31 | 2016 | 0.78 | 0.53 \pm 0.14 | 0.82 | 4 (55%) | Divergent |
| | 2017 | 0.76 | 0.53 \pm 0.14 | 0.82 | 4 (52%) | Mildly Divergent |
| | 2018 | 0.73 | 0.53 \pm 0.14 | 0.95 | 4 (57%) | Mildly Divergent |
| | 2019 | 0.57 | 0.53 \pm 0.14 | 0.82 | 4 (62%) | Reference |
| RGD-AQ11 | 2016 | 0.70 | 0.41 \pm 0.17 | 1.16 | 3 (38%) | Mildly Divergent |
| | 2017 | 0.70 | 0.41 \pm 0.17 | 1.16 | 3 (41%) | Mildly Divergent |
| | 2018 | 0.94 | 0.55 \pm 0.22 | 1.20 | 5 (38%) | Divergent |
| | 2019 | 0.71 | 0.41 \pm 0.17 | 1.16 | 3 (39%) | Mildly Divergent |
| RGD-DS-AQ12 | 2016 | 0.57 | 0.53 \pm 0.14 | 1.18 | 4 (26%) | Reference |
| | 2017 | 0.72 | 0.55 \pm 0.22 | 1.22 | 5 (46%) | Reference |
| | 2018 | 0.59 | 0.55 \pm 0.22 | 1.17 | 5 (68%) | Reference |
| | 2019 | 0.39 | 0.55 \pm 0.22 | 1.21 | 5 (44%) | Reference |

The RIVPACS O:E ratios ranged from 0.56 for Crabapple Creek (CRB-DS-AQ01) to 1.22 for River of Golden Dreams downstream site (RGD- DS-AQ12) (Table 3-4). The only site with an O:E ratio consistently less than 1.0 was Jordan Creek (JOR-DS-AQ31), where the ratio ranged from 0.82 to 0.95 (Table 3-4). For the other sites, the O:E ratio was greater than 1.0 for at least one year, and was greater than 1.0 for all four years at both sites on the River of Golden Dreams. The O:E data were also variable in relation to the multivariate assessment, with some sites assessed as within reference with a relatively low O:E ratio and other sites assessed as divergent with a relatively high O:E ratio (Table 3-4)

These data indicate that taxon richness at the test sites was on average comparable to what was expected based on the richness of reference sites. The divergence from reference condition identified through the CABIN multivariate assessment therefore was not primarily based on taxon richness.

3.3.3 Multimetric Site Characterization and Assessment

3.3.3.1 Total Abundance

Total abundance is a highly variable metric; for reference sites the CV ranged from 63% to 131% (Table 3-1). What this means is that abundance is an insensitive metric of only limited value in a site assessment. However, it is commonly reported and so is included within the analysis below.

As with the reference sites, abundance was also highly variable for the test sites, both among years and among sites. The highest abundance was recorded at Crabapple Creek (CRB-DS-AQ01), which ranged from approximately 2500 organisms to 3500 organisms per sample (Figure 3-1). The lowest abundance was recorded in the River of Golden Dreams (RGD-US-AQ11), which ranged from approximately 850 organisms to almost 1200 organisms per sample (Figure 3-1). The highest temporal variability was recorded at Jordan Creek (JOR-DS-AQ31), which ranged from just over 900 organisms to just over 2500 organisms per sample (Figure 3-1).

Despite the high spatial and temporal variability, abundance values for all sites in all years were considerably less than in the associated reference site: average site metric scores ranged from 0.21 to 0.73 among the five sites (Table 3-5). However, because of the high variability of the abundance metric, all sites were within the average metric threshold value of 0.06 (Table 3-5). What this means is that even though the abundance metrics were low in comparison to the reference average, they were still within one standard deviation of the measured abundances recorded for the reference sites. Benthic invertebrate abundance at all sites in all years was therefore in reference condition and considered unimpaired.

Table 3-5. Multimetric Assessment Scores

| Parameter | AQ21 | AQ01 | AQ31 | AQ12 | AQ11 | Threshold |
|------------------------|------|------|-------------|------|------|-------------|
| Abundance | 0.36 | 0.51 | 0.73 | 0.21 | 0.24 | 0.06 |
| EPT Relative Abundance | 1.06 | 1.16 | 0.53 | 0.95 | 1.05 | 0.66 |
| Total Taxon Richness | 1.19 | 0.89 | 0.81 | 1.08 | 1.13 | 0.73 |
| EPT Taxon Richness | 1.01 | 0.73 | 0.81 | 0.90 | 1.01 | 0.69 |
| % Dominant | 1.03 | 0.92 | 0.87 | 0.98 | 0.99 | 0.89 |

| Parameter | AQ21 | AQ01 | AQ31 | AQ12 | AQ11 | Threshold |
|-----------------------------|-------------|-------------|-------------|-------------|-------------|-------------|
| Shannon-Wiener Diversity | 1.07 | 0.80 | 0.81 | 1.08 | 1.03 | 0.78 |
| Average Metric Score | 0.95 | 0.84 | 0.76 | 0.87 | 0.91 | 0.75 |

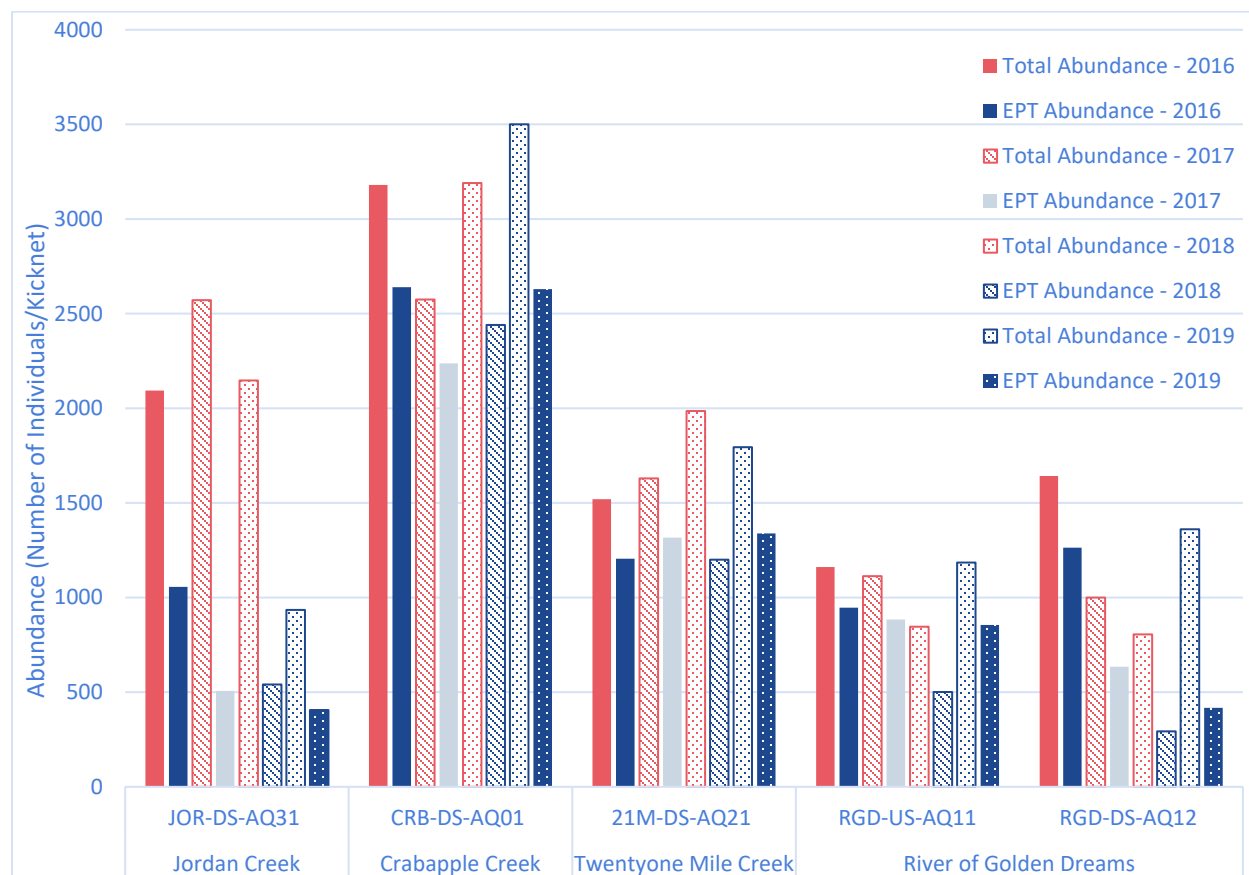


Figure 3-1. Benthic invertebrate total and Ephemeroptera, Plecoptera and Trichoptera (EPT) abundance by site and year, 2016-2019.

3.3.3.2 EPT Relative Abundance

Ephemeroptera, Plecoptera and Trichoptera (EPT) relative abundance ranged from 20% to 87% among all sites and years (Figure 3-5). The relative abundance of EPT taxa was comparable among sites in Crabapple Creek, Twentyone Mile Creek and the River of Golden Dreams, and within these sites was relatively stable varying only from 60% to 87% (Figure 3-2). In Jordan Creek EPT abundance was relatively low and variable, ranging from 20% to 50% through the four years of sampling (Figure 3-2). Of interest is the continuing decline in EPT relative abundance at the downstream site on River of Golden Dreams (RGD-DS-AQ12): a trends test in 2021 will indicate whether the decline is significant.

Average metric scores ranged from 0.53 at the Jordan Creek site (JOR-DS-AQ31) to 1.16 at the Crabapple Creek site (CRB-DS-AQ01), with three of the sites greater than 1.0 (Table 3-5). Comparison with the reference threshold of 0.66 indicates that Jordan Creek (JOR-DS-AQ31) was below the threshold, but that the other four sites were above the threshold. Benthic invertebrate EPT relative abundance at sites CRB-DS-AQ01, 21M-DS-AQ21, RGD-US-AQ11 and RGD-DS-AQ12 was therefore in reference condition and

considered unimpaired, while EPT relative abundance at site JOR-DS-AQ31 was considered mildly divergent.

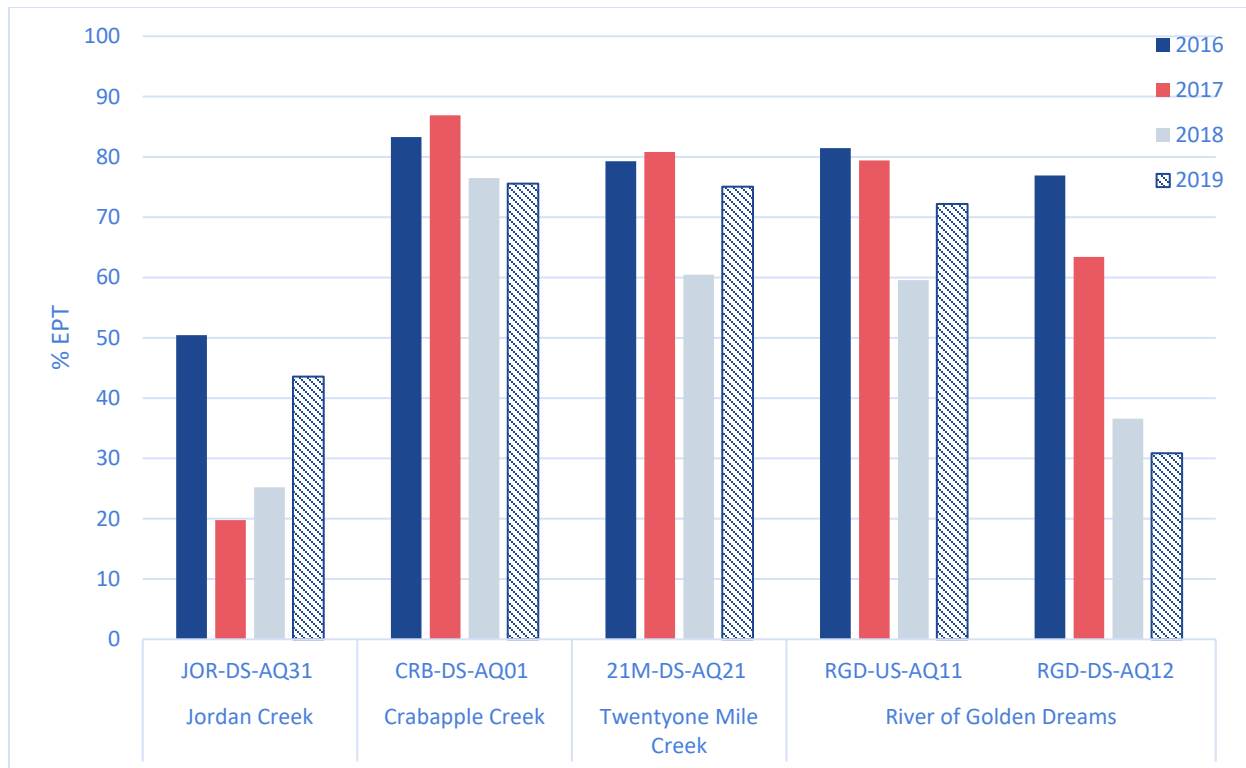


Figure 3-2. Benthic invertebrate community % EPT, 2016-2019.

3.3.3.3 Taxonomic Richness

Taxonomic richness varied from a low of 10 at site CRB-DS-AQ01 in 2017, to a high of 25 at site RGD-US-AQ11 in 2018 (Figure 3-3), but there were no consistent differences among sites. For all sites, taxonomic richness was higher in 2018 and 2019 than recorded in 2016 and 2017, ranging from 15 to 25 taxa in 2018 and 2019 and 10 to 16 taxa in 2016 and 2017 (Figure 3-3). Of interest is the continuing increase in taxa richness at site JOR-DS-AQ31 in Jordan Creek: a trends test in 2021 will indicate whether the increase is significant.

Average metric scores ranged from 0.81 at JOR-DS-AQ31 to 1.19 at 21M-DS-AQ21, with three of the sites greater than 1.0 (Table 3-4). Comparison with the reference threshold of 0.73 indicates that all sites were above the threshold. Benthic invertebrate taxon richness at all sites in all years was therefore in reference condition and considered unimpaired.

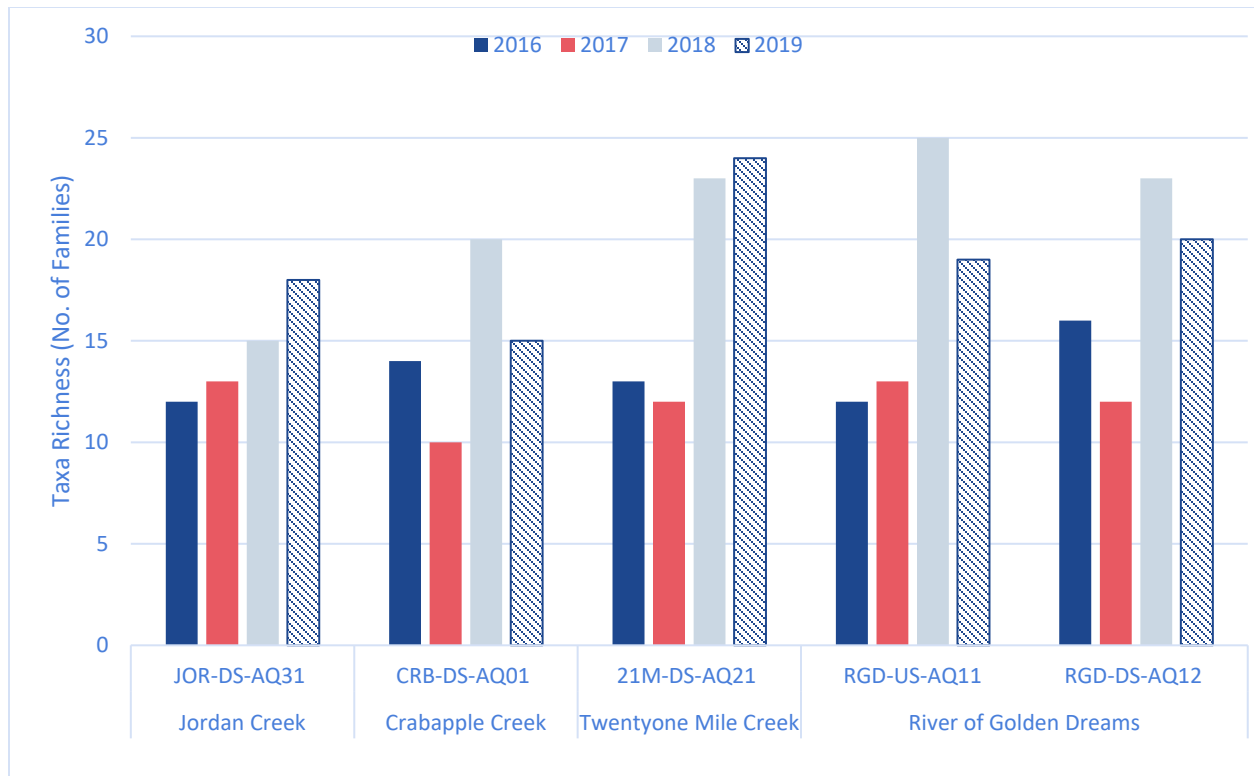


Figure 3-3. Benthic invertebrate community taxa richness, 2016-2019.

3.3.3.4 EPT Taxon Richness

The EPT richness ranged from 6 to 13 among the sites and years, but as with total richness, there were no consistent differences among sites (Figure 3-4). In 2019, richness of EPT taxa ranged from seven families at Crabapple Creek to 13 families at the 21 Mile Creek site (Figure 3-4). The 2019 EPT taxa richness was consistent with 2018 for the following sites: Jordan Creek, 21 Mile Creek, and the downstream River of Golden Dreams site (RGD-DS-AQ12). Overall, an increase in EPT taxa richness was recorded in 2018 and 2019, compared with 2016 and 2017.

Average metric scores ranged from 0.73 at CRB-DS-AQ01 to 1.01 at 21M-DS-AQ21 and RGD-US-AQ11, with two of the sites greater than 1.0 (Table 3-5). Comparison with the reference threshold of 0.69 indicates that all sites were above the threshold. Benthic invertebrate EPT taxon richness at all sites in all years was therefore in reference condition and considered unimpaired, which means that EPT taxa are present and persisting at all five of the study sites.

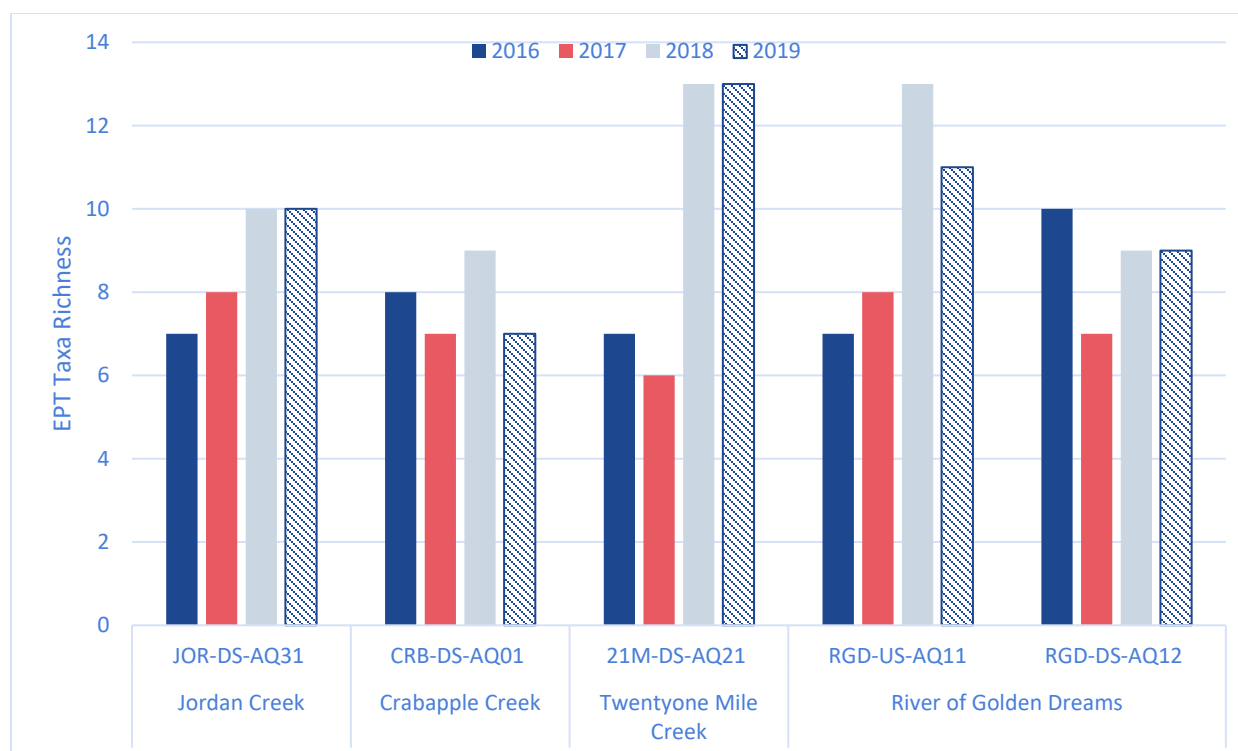


Figure 3-4. Benthic invertebrate community EPT tax richness, 2016-2019.

3.3.3.5 Community Composition

Sites in Crabapple Creek, 21 Mile Creek and the upper River of Golden Dreams site AQ11, which are all clustered in the upper River of Golden Dreams system, had similar community structure, with 27% or less of Diptera+non-insects and greater than 50% of EPT taxa (Figure 3-5).

The downstream site on the River of Golden Dreams (RGD-DS-AQ12) had a higher proportion of Diptera+non-insects (57%) relative to the upstream sites (Crabapple Creek, 21 Mile Creek and River of Golden Dreams upstream site) and a lower proportion of Ephemeroptera (28%). The site also had the highest proportion of invertebrates in the 'other' category (12%).

Diptera+non-insects comprised just over half (56%) of the benthic community at the Jordan Creek site (Figure 3-5). This proportion was similar to that recorded in 2016 (50%). Notably, in 2017 and 2018 a shift to higher proportions of Diptera+non-insects were recorded and suggested a decline in overall community health potentially caused by organic pollution.

Compilation of the five dominant species for each site resulted in average metric scores that ranged from 0.87 at JOR-DS-AQ31 in Jordan Creek to 1.03 at 21M-DS-AQ21 in Twentyone Mile Creek (Table 3-5). Comparison with the reference threshold of 0.89 indicates that site JOR-DS-AQ31 in Jordan Creek was below the threshold, but that the other four sites were above the threshold. Benthic invertebrate dominance at sites CRB-DS-AQ01, 21M-DS-AQ21, RGD-US-AQ11 and RGD-DS-AQ12 was therefore in reference condition and considered unimpaired, while dominance at site AQ31 was considered mildly divergent.

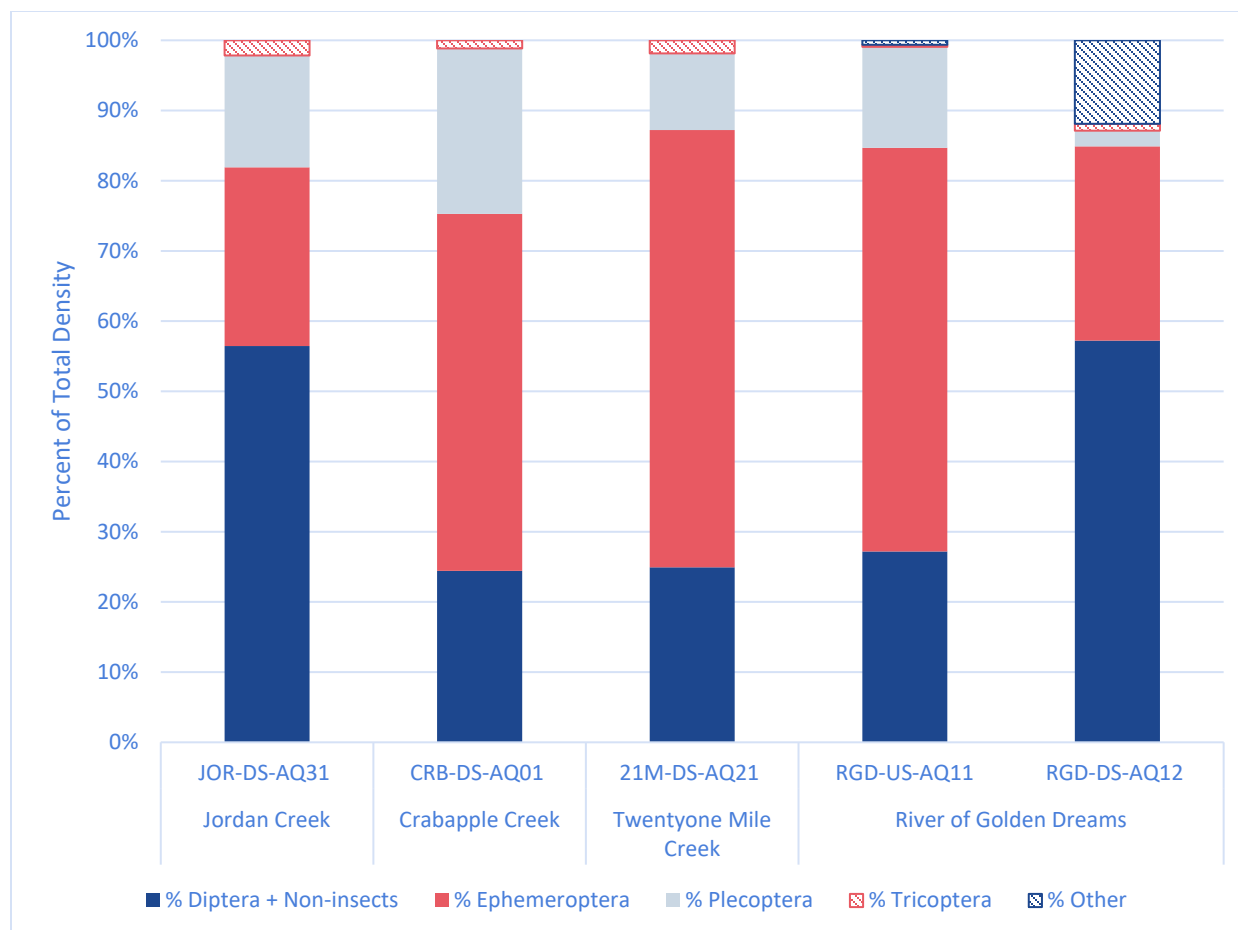


Figure 3-5. Relative densities of benthic invertebrate communities by site, 2019.

3.3.3.6 Shannon-Wiener Diversity

Diversity ranged from 1.07 at CRB-DS-AQ01 in 2017 to 2.32 at RGD-DS-AQ12 in 2018 (Figure 3-6). On average, the lowest diversity was recorded at site JOR-DS-AQ31 in Jordan Creek and the highest diversity was recorded at Site RGD-DS-AQ12 in the River of Golden Dreams. Diversity was relatively stable through the period of record, but it did appear that diversity was slightly higher in 2018 and 2019 than in 2016 and 2017 (Figure 3-6).

Average metric scores for diversity ranged from 0.80 at CRB-DS-AQ01 to 1.08 at RGD-DS-AQ12, with three of the sites greater than 1.0 (Table 3-5). Comparison with the reference threshold of 0.78 indicates that all sites were above the threshold. Benthic invertebrate diversity at all sites in all years was therefore in reference condition and considered unimpaired.

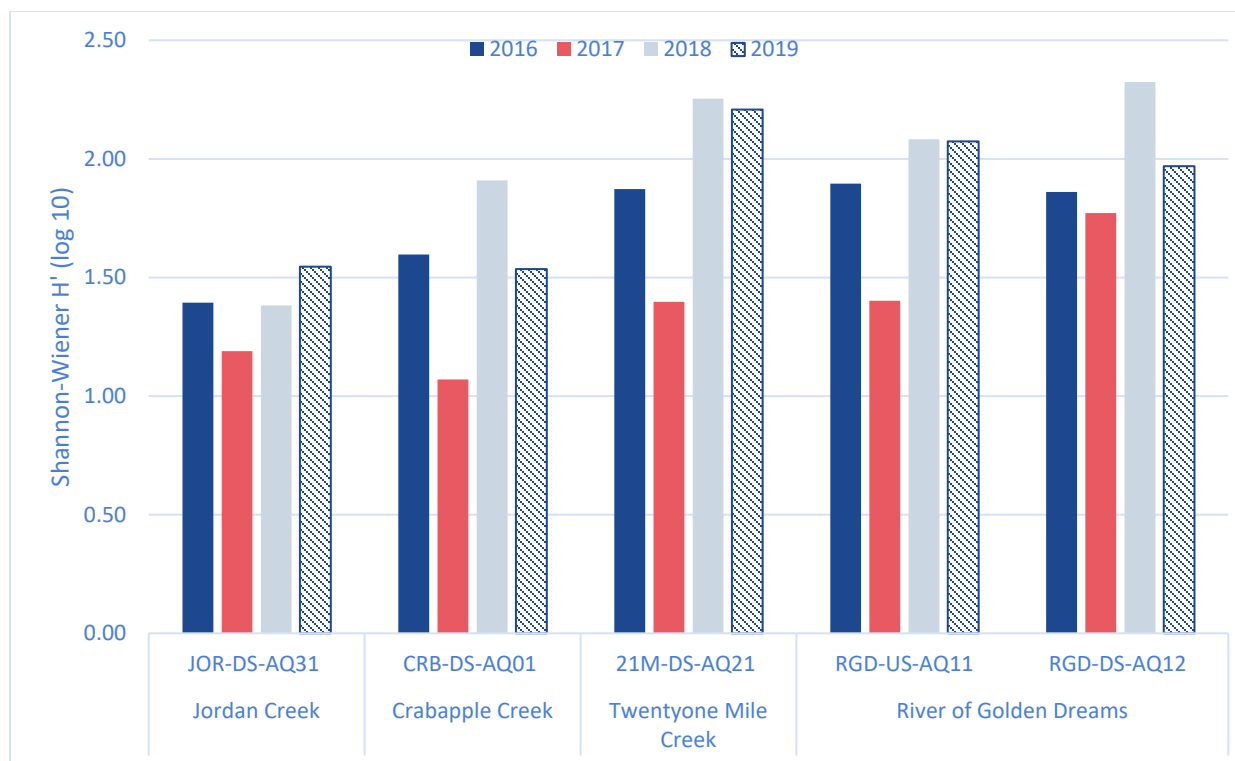


Figure 3-6. Shannon-Weiner indices, 2016-2019.

3.3.4 Hilsenhoff Index of Biotic Integrity

Compilation of the HIBI scores was undertaken and indicated that the scores varied from 3.58 at site 21M-DS-AQ21 to 5.43 at site RGD-DS-AQ12 (Table 3-6). Based on these scores, classification ranged from good to very good, indicating that the potential for some organic pollution existed at site RGD-DS-AQ12 and site JOR-DS-AQ31, but that there was only the possibility of slight organic pollution at the rest of the sites (Table 3-6). These results indicate that significant input of organic pollutants has not occurred to date within the test streams.

Table 3-6. Hilsenhoff Index of Biotic Integrity Assessment Scores

| Site | Stream Order | 2018 | 2019 | Classification* |
|--|--------------|------|------|-----------------|
| AQ01 – Crabapple Creek | 2 | 3.93 | 4.25 | Very Good |
| AQ11 – River of Golden Dreams Upstream | 3 | 3.93 | 3.98 | Very Good |
| AQ12 – River of Golden Dreams Downstream | 3 | 5.14 | 5.43 | Good |
| AQ21 – Twentyone Mile Creek | 3 | 3.75 | 3.58 | Very Good |
| AQ31 – Jordan Creek | 2 | 5.21 | 4.66 | Good |

*See Table 3-2

3.4 Assessment Conclusions

Conclusions regarding the condition of the four creeks are provided below for each site.

3.4.1 Crabapple Creek Site CRB-DS-AQ01

Crabapple Creek was assessed as in reference condition to mildly divergent using the CABIN assessment and was assessed within reference condition for all six core metrics with an average metric score of 0.84. The site was also assessed as in very good condition using the HIBI. These results suggest that Crabapple Creek through the period of record was generally in reference condition and unimpaired through the period of record.

3.4.2 River of Golden Dreams Upstream Site RGD-US-AQ11

The upstream site in the River of Golden Dreams was assessed as mildly divergent to divergent using the CABIN assessment. In contrast, the site was assessed as within reference condition for all six core metrics with an average metric score of 0.91 and in very good condition using the HIBI. The site also had a greater number of taxa than expected. The weight-of-evidence therefore suggests that the upstream reach of the River of Golden Dreams was in reference condition and unimpaired through the period of record.

3.4.3 River of Golden Dreams Downstream Site RGD-DS-AQ12

The down stream site in the River of Golden Dreams was assessed as in reference condition for all four years using the CABIN assessment. The site was also assessed as within reference condition for all six core metrics with an average metric score of 0.87 and in good condition using the HIBI. The site also had a greater number of taxa than expected. The weight-of-evidence therefore suggests that the downstream reach of the River of Golden Dreams was in reference condition and unimpaired through the period of record.

3.4.4 Twentyone Mile Creek Downstream Site 21M-DS-AQ21

The downstream site in Twentyone Mile Creek was assessed as mildly divergent to divergent using the CABIN assessment. In contrast, the site was assessed as within reference condition for all six core metrics with an average metric score of 0.95 and in very good condition using the HIBI. The site also had a greater number of taxa than expected. The weight-of-evidence therefore suggests that the downstream reach of Twentyone Mile Creek was in reference condition and unimpaired through the period of record.

3.4.5 Jordan Creek Downstream Site JOR-DS-AQ31

The downstream site in Jordan Creek was assessed as ranging from reference to divergent using the CABIN assessment. The site was also assessed as within reference condition for only four of the six core metrics with an average metric score of 0.76, which was the lowest of all the sites. The site also consistently had a fewer number of taxa than expected. The site was, however, assessed as in good condition using the HIBI. The weight-of-evidence therefore suggests that the downstream reach of Jordan Creek is likely mildly divergent from reference condition and may be slightly impaired at present.

4. Fish Community

4.1 Introduction

The objective of the aquatic species monitoring program was to assess relative aquatic health of local watercourses using important indicator species such as Kokanee Salmon (*Oncorhynchus nerka*), Bull Trout (*Salvelinus confluentus*), Rainbow Trout (*O. mykiss*) and Coastal Cutthroat Trout (*O. clarkii clarkii*). The 2019 fisheries program remained consistent with previous years. In 2018 minnow trapping was added as a sampling method to augment data collected via electrofishing and facilitate fish sampling in the River of Golden Dreams.

Kokanee Salmon are present in the study streams, with known spawning areas in the River of Golden Dreams. Bull Trout, as well as Cutthroat Trout, are native to the Whistler area, but observations of these species are rare. Both species are blue-listed, indicating that they are considered vulnerable in BC. The Lower Mainland populations of Coastal Cutthroat Trout are in serious decline (BC MoFLNRO 2017a). Within the Whistler area, Cutthroat Trout are believed to have hybridized with Rainbow Trout. Populations of Bull Trout are also in decline in BC and throughout the global range of this species (BC MoFLNRO 2017b). Bull Trout are very similar in shape and coloration to Dolly Varden (*Salvelinus malma*) and genetic analysis is required to definitively differentiate individuals of these species. Rainbow Trout are ubiquitous in the study streams and were stocked in Rainbow Lake (the headwater lake of 21 Mile Creek) in the late 1970s or early 1980s (Eric Crowe, pers. comm). Sculpin (*Cottus sp.*) and Threespine Stickleback (*Gasterosteus aculeatus*) are also common.

4.2 Methods

Streams were sampled for fish between July 31st and August 2nd, 2019. Table 4-1 provides a complete list of 2018 fish sampling sites.

The fish community within RMOW streams were sampled in 2019 under the Scientific Fish Collection Permit SU19-524197 issued by the BC Ministry of Forests Lands and Natural Resource Operations (BC MoFLNRO). The fish community was sampled using a combination of backpack electrofishing and minnow traps. Electrofishing was not completed in The River of Golden Dreams due to human safety concerns; only minnow traps were used at these locations. Electrofishing was completed at stream sites by a two-person crew using a Smith-Root LR-24 backpack electrofisher following methods outlined in Johnston et al. (2007). Only one electrofishing pass was made at each site; no stop nets were used. Electrofisher voltage, duty cycle and frequency settings were adjusted based on site conditions in order to maximize efficiency and minimize the risk of injury to fish. Electrofisher settings are summarized in Table 4-1. The electrofishing effort was recorded for each site.

Minnow traps were set at each stream site after electrofishing was completed as well as at the River of Golden Dreams where no electrofishing could be conducted. Traps consisted of two cylinders made of 6.35 mm galvanized steel wire mesh with a conical entrance, measuring 42 cm long and 23 cm in diameter. The cylinders were clipped together, baited with cat food and set overnight. Table 4-1 provides a summary of total fishing effort for gear used at each stream site.

All fish captured were identified to species, enumerated and measured for length (to the nearest 1 mm) and wet weight (to the nearest 0.1 g using a Scout Pro 400 g scale). Fork length was measured for salmonid fish species and total length was measured for other species. Any lesions, parasites, or other anomalies on fish were recorded before the fish were released live back at the site of capture.

Table 4-1. Fish sampling methods and effort at stream sites in the RMOW areas, 2019

| Creek | Site ID | Gear Type | Date Sampled/Set | Minnow Trapping | | | Electrofishing | | | |
|--------------------------------|-------------|-----------|------------------|-----------------|-----------------|--------------------|----------------|----------------|----------------|--------------------|
| | | | | Date Retrieved | Number of Traps | Total Effort (hrs) | Voltage (V) | Frequency (Hz) | Duty Cycle (%) | Total Effort (sec) |
| Jordan Creek | JOR-DS-AQ31 | EF | 01-Aug-19 | - | - | - | 220 | 30 | 15 | 402 |
| | | MT | 31-Jul-19 | 01-Aug-19 | 5 | 122.5 | - | - | - | - |
| Crabapple Creek | CRB-DS-AQ01 | EF | 01-Aug-19 | - | - | - | 225 | 30 | 12 | 1558 |
| | | MT | 01-Aug-19 | 02-Aug-19 | 5 | 130 | - | - | - | - |
| River of Golden Dreams (Upper) | RGD-US-AQ11 | MT | 01-Aug-19 | 02-Aug-19 | 5 | 119.01 | - | - | - | - |
| River of Golden Dreams (Lower) | RGD-DS-AQ12 | MT | 31-Jul-19 | 01-Aug-19 | 5 | 127.5 | - | - | - | - |
| 21 Mile Creek | 21M-DS-AQ21 | EF | 31-Jul-19 | - | - | - | 450 | 35 | 20 | 844 |
| | | MT | 31-Jul-19 | 02-Aug-19 | 5 | 222.5 | - | - | - | - |

4.2.1 Data Analysis

Field identification of juvenile trout can be confounded where Rainbow Trout occur in the same geographic area and frequently encounter on another (sympatry) with Coastal Cutthroat Trout, in part because hybridization commonly occurs between the two species and because hybrids themselves are difficult to differentiate (Baumsteiger 2005). Visual identification error rates for juvenile trout (sympatric Coastal Cutthroat and Rainbow Trout populations) can be quite high without genetic analyses to corroborate genotypes. Similar to 2018 (Palmer and Snowline 2019), 2019 field crews did not identify any suspected hybrid offspring of Coastal Cutthroat and Rainbow Trout (Photo 4-1). In the absence of genetic analyses to provide accurate identification of individual fish and the fact that a suspected hybrid was identified in 2016 within the Ecosystem Monitoring Program study area (Photo 4-2; Palmer and Snowline 2017), results are discussed in terms of ‘unknown’ trout within this report.



Photo 4-1. Rainbow Trout (fork length 136 mm) captured in Crabapple Creek (CRB-DS-AQ01) during 2018 electrofishing efforts. Date: August 1, 2019.



Photo 4-2. Suspected hybrid trout (fork length 84 mm) captured in 21 Mile Creek in 2016 (21M-DS-AQ21). Date: August 6, 2016.

Fish Abundance

Fish community data was summarized by calculating catch-per-unit-effort (CPUE) for each individual fishing effort, gear type and fish species captured. CPUE is an index of relative abundance that can be used to compare fish populations among different areas with the assumption that catch is proportional to the amount of effort for each gear-type used. CPUE is defined as the number of fish captured per sampling device per unit time. CPUE is summarized for each gear type and by species.

Electrofishing:

$$CPUE = \text{number of fish caught} * [100 / (\text{electrofishing effort, hr})]$$

Minnow Traps:

$$CPUE = \text{number of fish caught per trap} * [24 \text{ hr} / (\text{set time, hr})]$$

Length, Weight and Condition

Mean length and weight were calculated for each fish species; further analyses were only completed on trout, as this species was proposed as an indicator species in the past and the focus of analysis in previous Whistler Ecosystem Monitoring reports (Palmer and Snowline 2017-2019).

Site-specific length-age regressions for trout were calculated as:

$$\log_{10}(W) = a + b \times \log_{10}(L) \quad (1)$$

where W = weight (g), L = length (mm), a = the intercept of the regression and b = the slope of the regression.

One sample t-tests were performed on estimated weight-length slope coefficients to determine if slopes significantly differed from the isometric growth value of three. Isometric fish growth occurs when length and weight increase at the same rate as the fish grows, whereas allometric growth occurs when length and weight increase at different rates during fish growth. Isometric and allometric growth are used to understand length-weight relationships in organisms. Slope coefficients of the estimated weight-length slope used in t-tests were estimated using species-specific linear regressions. Isometric growth is a requirement for calculating fish condition using the Fulton condition factor (K), as it assumes that fish shape does not change with increasing length. Trout condition could not be assessed using the Fulton condition factor, due to allometric growth. Instead, the relative condition factor (K_n) was used to characterize fish condition:

$$K_n = \frac{W}{W'} \quad (2)$$

where W = fish actual weight (g) and W' = predicted length-specific weight using the length-weight regression outlined in Equation 1.

4.2.2 Quality Assurance/Quality Control

Field equipment was calibrated prior to the start of the field season, properly maintained and kept clean and free of excess water. The YSI meter was re-calibrated multiple times while in the field. All scales were regularly tared to maintain accuracy while in use. Care was taken to clean equipment between samples to prevent cross contamination.

All data was recorded on waterproof paper and examined for completeness and accuracy. All captured fish were identified to the lowest possible taxonomic level and a subset were photographed for verification of species identification.

All fisheries field data were transferred to electronic spreadsheets in the office. The spreadsheets were compared with the field notes to identify and correct transcription errors. A variety of other measures were taken to further ensure the validity of the data. For example, fish weights were plotted against fish lengths for each species separately to identify outliers that may have been due to errors in recording or transcription. Outliers were excluded from the analyzed dataset.

4.3 Results and Discussion

In 2019, fish community assessments were completed at five stream sites within the RMOW study area (Figure 2-1). Fish catch-per-unit-effort (CPUE) by species and sampling gear are presented in Table 4-2 and Table 4-3. Biological data for fish sampled in the RMOW study area are presented in Appendix D.

A total of 179 fish were captured during 2019 electrofishing and minnow trap efforts. Similar to 2018 only three fish species were identified in streams sampled in 2019, including unidentified trout, Sculpin and Threespine Stickleback. No Bull Trout or Dolly Varden were observed. Sculpin represented the overall dominant fish species captured during 2019 electrofishing efforts, comprising over 67% of the capture in Jordan Creek, Crabapple Creek and 21 Mile Creek (Figure 4-1). Threespine Stickleback composed 50% or more of the catch using minnow trapping at all five sampling locations (Figure 4-2). Overall, trout have shown a decline in capture since 2016. Similar to 2018 no trout were captured at the upstream site of the River of Golden Dreams where only minnow traps were deployed (Palmer and Snowline 2019).

Table 4-2. Electrofishing effort and fish caught in surveys conducted in the RMOW study area, 2019.

| Site | Creek | Date | Effort (seconds) | Catch (number of individuals) | | | | CPUE (#/100s) | | | |
|-------------|-----------------|-----------|------------------|-------------------------------|----|-----|-------|---------------|------|------|-------|
| | | | | CC | TR | TSB | Total | CC | TR | TSB | Total |
| CRB-DS-AQ01 | Crabapple Creek | 01-Aug-19 | 1558 | 4 | 1 | 1 | 6 | 0.26 | 0.06 | 0.06 | 0.39 |
| JOR-DS-AQ31 | Jordan Creek | 01-Aug-19 | 402 | 26 | 4 | 1 | 31 | 6.47 | 1.00 | 0.25 | 7.71 |
| 21M-DS-AQ21 | 21-Mile Creek | 31-Jul-19 | 844 | 30 | 11 | 0 | 41 | 3.55 | 1.30 | 0.00 | 4.86 |

Notes: CPUE = catch-per-unit-effort, CC = Sculpin (General), TR = trout, TSB = Threespine Stickleback

Table 4-3. Minnow trap effort and fish caught in surveys conducted in the RMOW study area, 2019.

| Site | Creek | Date Traps Set | Date Retrieved | Number of Traps | Effort (hrs) | CC | TR | TSB | Total | CC | TR | TSB | Total |
|-------------|------------------------|----------------|----------------|-----------------|--------------|----|----|-----|-------|------|------|------|-------|
| CRB-DS-AQ01 | Crabapple Creek | 01-Aug-19 | 02-Aug-19 | 5 | 130 | 0 | 4 | 4 | 8 | 0.00 | 0.15 | 0.15 | 0.30 |
| JOR-DS-AQ31 | Jordan Creek | 31-Jul-19 | 01-Aug-19 | 5 | 122.5 | 1 | 5 | 9 | 15 | 0.04 | 0.20 | 0.35 | 0.59 |
| RGD-DS-AQ12 | River of Golden Dreams | 31-Jul-19 | 01-Aug-19 | 5 | 136.65 | 6 | 2 | 55 | 63 | 0.23 | 0.08 | 2.07 | 2.37 |
| 21M-DS-AQ21 | 21-Mile Creek | 31-Jul-19 | 02-Aug-19 | 5 | 222.5 | 0 | 4 | 6 | 10 | 0.00 | 0.09 | 0.13 | 0.22 |
| RGD-AQ11 | River of Golden Dreams | 01-Aug-19 | 02-Aug-19 | 5 | 127.5 | 0 | 0 | 6 | 6 | 0.00 | 0.00 | 0.23 | 0.23 |

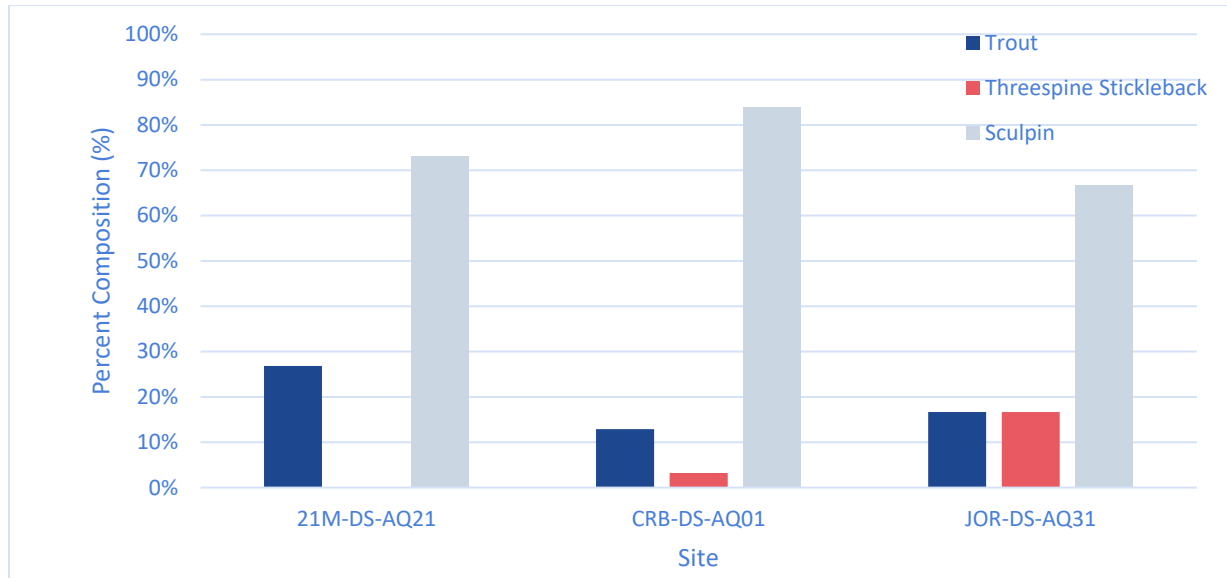


Figure 4-1. *Percent composition of fish species captured electrofishing streams in RMOW study area, 2019.*



Figure 4-2. *Percent composition of fish species captured in streams using minnow traps, 2019.*

4.3.1.1 *Lengths, Weights and Condition*

The mean length and weight of each fish sampled in 2019 is presented in Table 4-4. A length-frequency analysis for trout sampled in 2016 – 2019 is presented in Figure 4-3. Note that the larger numbers of fish captured in 2016 (n=102) relative to 2017 (n=9), 2018 (n=20) and 2019 (n=31) is most likely due to the multiple pass depletion electrofishing method used in that year (Palmer and Snowline 2017).

In 2019, trout were consistently the largest fish species captured in the study area. Amongst all sites, trout ranged in length from 27 mm to 136 mm and in weight from less than 0.1 g to 48.0 g (Table 4-4). The largest trout were captured in Crabapple Creek and Jordan Creek.

Table 4-4. Length and weights of fish captured in the RMOW study area, 2018.

| Creek | Site ID | Species | Number | Length (mm) | | | | Weight (g) | | | |
|--------------------------------|-------------|---------|--------|-------------|------|-------|------|------------|------|------|------|
| | | | | Min | Mean | Max | SD | Min | Mean | Max | SD |
| Jordan Creek | JOR-DS-AQ03 | CC | 5 | 58.0 | 74.2 | 94.0 | 14.8 | 2.2 | 14.2 | 48.0 | 19.3 |
| | | TR | 6 | 46.0 | 81.7 | 128.0 | 32.1 | 1.3 | 8.7 | 21.6 | 7.7 |
| | | TSB | 10 | 47.0 | 51.9 | 59.0 | 4.2 | 1.1 | 1.6 | 2.0 | 0.3 |
| Crabapple Creek | CRB-DS-AQ01 | CC | 8 | 28.0 | 86.1 | 136.0 | 36.8 | 0.6 | 2.4 | 8.3 | 1.7 |
| | | TR | 8 | 28.0 | 86.1 | 136.0 | 36.8 | 0.1 | 9.1 | 23.3 | 8.0 |
| | | TSB | 5 | 52.0 | 59.0 | 70.0 | 8.0 | 1.8 | 2.4 | 3.3 | 0.7 |
| River of Golden Dreams (Upper) | RGD-AQ11 | TSB | 6 | 48.0 | 58.0 | 80.0 | 12.0 | 1.0 | 2.4 | 6.2 | 2.0 |
| River of Golden Dreams (Lower) | RGD-DS-AQ12 | CC | 6 | 55.0 | 61.3 | 68.0 | 4.9 | 0.3 | 0.8 | 1.7 | 0.3 |
| | | TR | 2 | 73.0 | 78.5 | 84.0 | 7.8 | 3.6 | 4.7 | 5.8 | 1.6 |
| | | TSB | 55 | 33.0 | 42.9 | 55.0 | 4.2 | 1.7 | 2.5 | 3.3 | 0.6 |
| 21 Mile Creek | 21M-DS-AQ21 | CC | 30 | 39.0 | 50.6 | 76.0 | 9.1 | 0.4 | 2.4 | 13.0 | 2.8 |
| | | TR | 15 | 27.0 | 50.9 | 87.0 | 23.3 | 0.1 | 2.4 | 7.0 | 2.8 |
| | | TSB | 5 | 46.0 | 49.4 | 52.0 | 3.1 | 0.6 | 1.3 | 1.8 | 0.6 |
| All Sites | | CC | 67 | 39.0 | 56.4 | 94.0 | 12.0 | 0.4 | 3.3 | 48.0 | 6.1 |
| | | TR | 31 | 27.0 | 67.7 | 136.0 | 32.0 | 0.1 | 5.5 | 23.3 | 6.2 |
| | | TSB | 81 | 33.0 | 46.6 | 80.0 | 7.7 | 0.3 | 1.1 | 6.2 | 0.8 |

Notes: CC = Sculpin (General), TR = trout, TSB = Threespine Stickleback

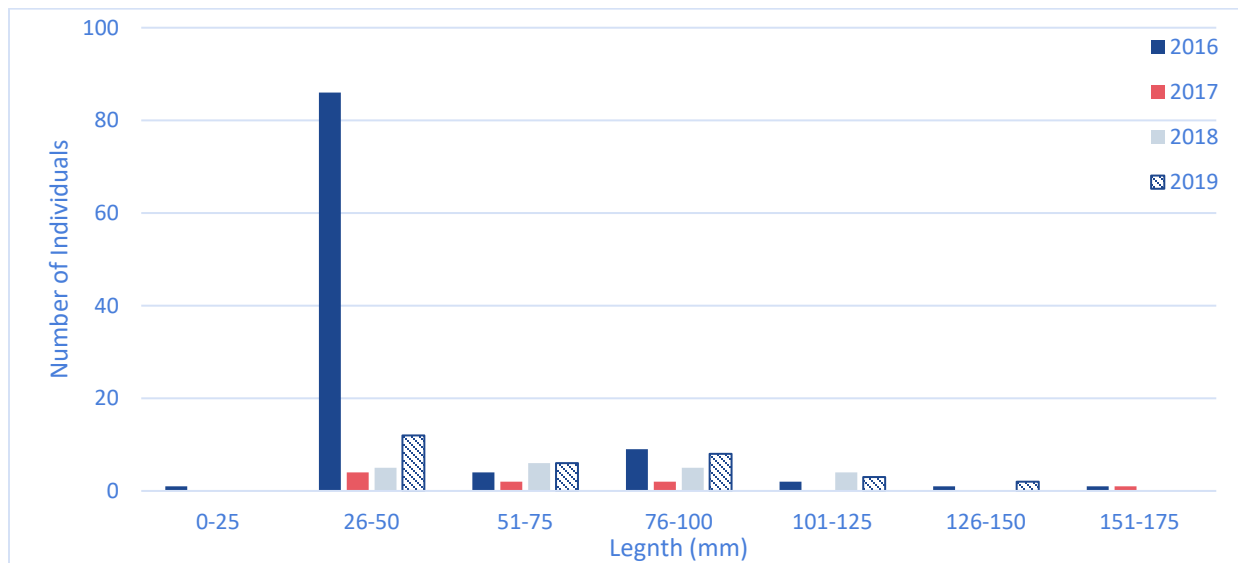


Figure 4-3. Length-frequency analysis for sampled trout (electrofishing and minnow trap collection methods) in study streams, August 2016, 2017, 2018 and 2019.

Condition

The length to weight relationship of all the trout sampled in 2016 - 2019 are presented in Figure 4-4. The length-weight linear regression for juvenile trout collected in 2019 was significant (Linear regression, slope = 3.54, $R^2 = 0.97$, $df = 29$, $p < 0.05$). Due to a slope value greater than 3.0 (3.54), trout growth was shown to be positively allometric (t-test, $t = 2.74$, $df = 29$, $p < 0.05$), that is, fish length increased more quickly relative to weight. This positive allometric growth was also in juvenile trout captured in 2018 (Palmer and Snowline, 2019). Conversely, in 2017 trout growth was shown to be isometric (t-test, $t = 0.76$, $df = 7$, $P = 0.47$) with fish having relatively similar ratios between growth in length and weight (Palmer and Snowline 2018). In 2016, trout growth was also allometric but in this year was negatively allometric (slope value less than 3.0), thus showing weight increasing quickly relative to length (Palmer and Snowline, 2017).

Due to the low sample size of trout in 2017 and 2018, statistics derived from this data have limited power and therefore results should be interpreted with caution. For example, even though the length-weight relationships of 2016 - 2019 trout appear similar (Figure 4-4), statistical analysis showed significant differences in trout growth relationships (isometric vs. allometric). As trout captured in 2018 and 2019 show allometric growth, relative condition was used to assess fish condition. Refer to Section 4.2.2 for detailed analyses.

Mean relative condition (K_n) for trout captured in RMOW from 2016 -2019 is presented in Figure 4-5. In salmonids, a condition of 1 is considered normal for a healthy fish. Therefore, all trout sampled within the RMOW in 2019 were considered healthy based on the relative condition index. Within the four sites where trout were captured, all showed similar condition values for 2019.

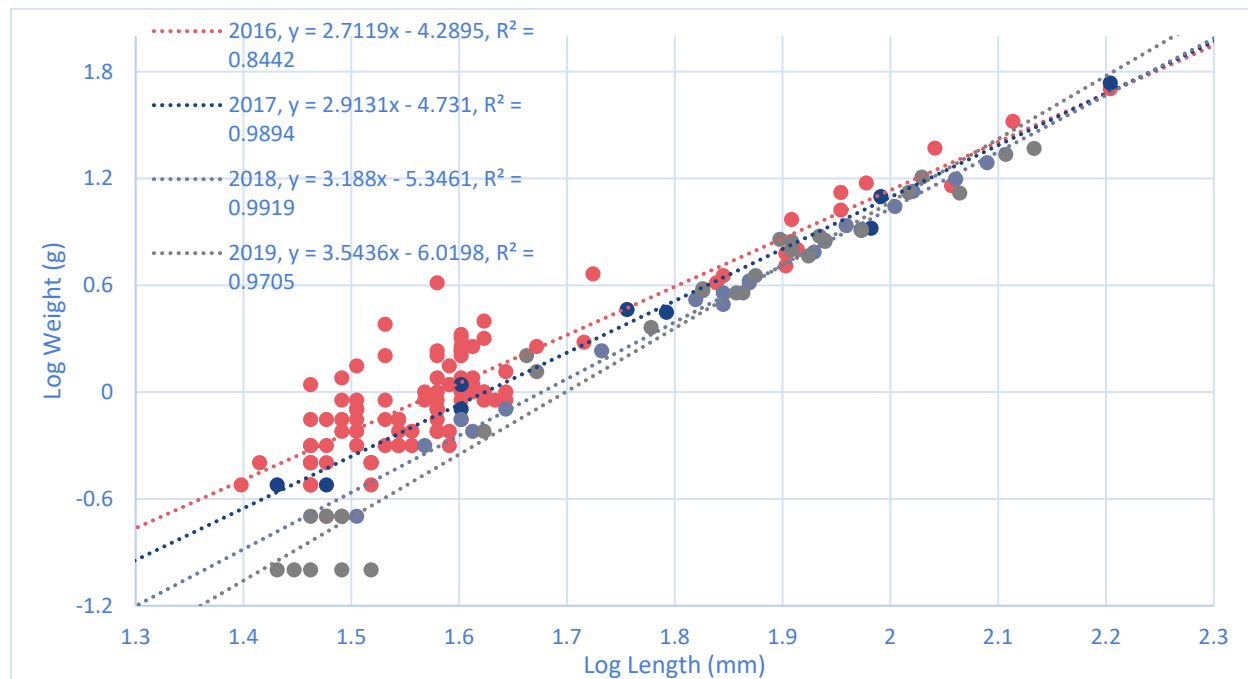


Figure 4-4. Weight-length relationship for juvenile trout captured in the RMOW study streams, 2016 - 2019.

The purpose of the fish sampling program was to develop a greater understanding of the fish communities in the streams within the study area and to help identify any potential impacts to these sites. The 2019 sampling program built upon the work completed in 2016 to 2018. Fish community health is a product of the environment in which they live in. As fish occupy a higher trophic level and are longer-lived compared to other aquatic organisms (e.g. benthic invertebrates), fish community data can provide information on the long-term health of a system.

The fish communities within the Ecosystems Monitoring Program study area are inhabited by 0+ year fry and juvenile trout, demonstrating the importance of the study reaches as rearing and feeding habitat. At present, trout collected could represent either cutthroat or rainbow trout. As recommended in previous years, the collection of aging structures and the collection of fin clips to conduct genetic analysis from trout captured within the study area would contribute to a greater understanding of the trout community inhabiting these systems within the RMOW.

Changes in length weight relationship between years is not a currently a concern. The specific species of trout captured remains unknown and mixed catch can affect the length/weight relationship between years and sites. Trout have been captured from several distinct streams and habitats; the proportion of trout captured from each stream changes from year to year. This proportional difference in sampled fish can effect the length weight relationship.

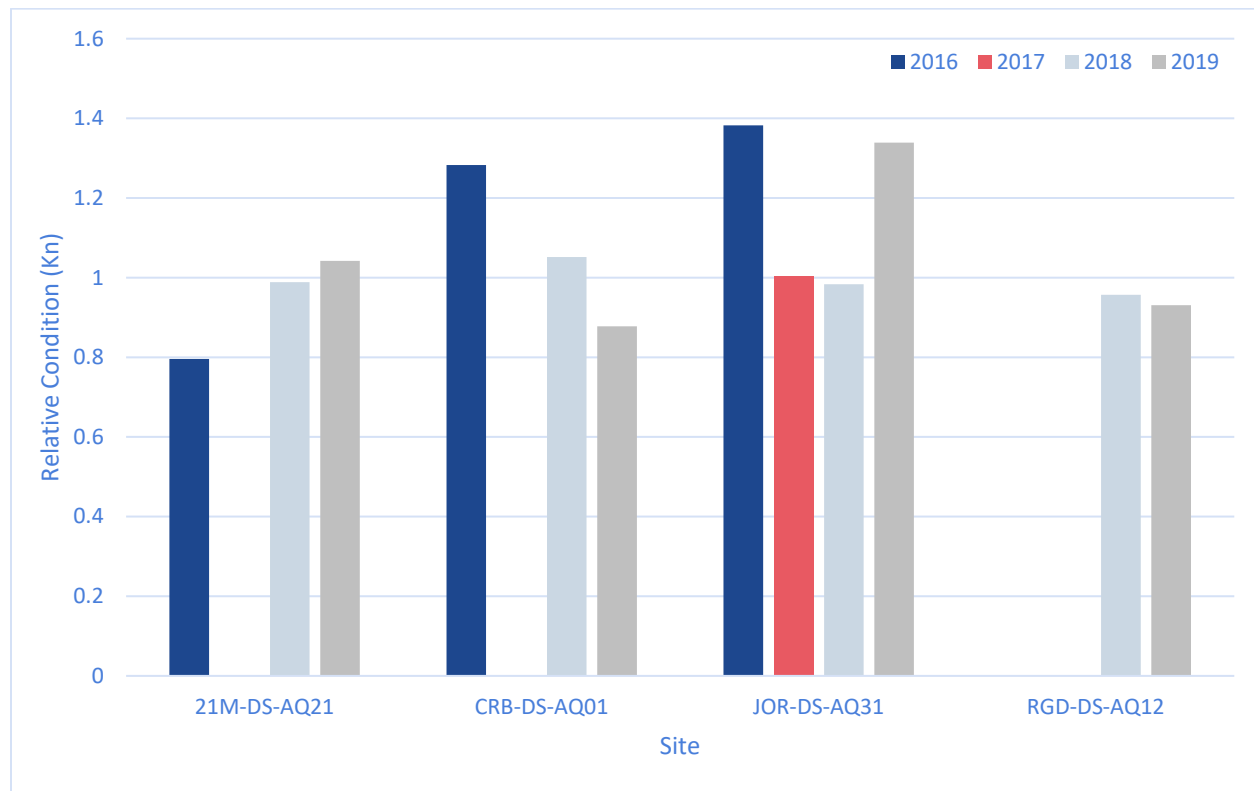


Figure 4-5. *Relative condition of trout captured in the RMOW study area, 2016 - 2019.*

5. Coastal Tailed Frogs

Lead Biologist and Author: Bob Brett

Additional Surveyors: Jagoda Kozikowska and Hillary Williamson

5.1 Introduction

Amphibians have long been used as indicators of ecosystem health. They have physiological constraints and sensitivities due to subcutaneous respiration, specialized adaptations and microhabitat requirements, as well as a dual life cycle that utilizes aquatic and terrestrial habitats. These characteristics make them susceptible to perturbations in both habitat types and suitable indicator species of ecosystem health.

Stream-dwelling amphibians such as the Coastal Tailed Frog (*Ascaphus truei*) serve a vital role as indicators of stream health as they require flowing, clear, cold water throughout their lifecycle (Matsuda et al. 2006) and are vulnerable to habitat alteration and degradation such as siltation and algal growth. They are also highly philopatric,³ long-lived and maintain relatively stable populations. These attributes make them more trackable and reliable as indicators of potential biotic diversity in stream ecosystems than anadromous fish and their relative abundance can be a useful indicator of stream condition (Welsh and Ollivier 1998).

Ideal habitats for tailed frogs are smaller, fast-flowing (gradients usually >10%) mountainside streams that are cool (typically 10 to 15°C in late summer, but at least 5° C for egg development), have a cobble-boulder substrate with rounded to subangular-shaped rocks, and a cascade or step pool morphology (Matsuda et al. 2006; BC MOE 2015). These characteristics describe many of the streams that drain into the Whistler Valley. Tadpoles have been detected in most Whistler streams surveyed to date (Wind 2005-2009; Cascade 2014-2016; Palmer and Snowline 2017-2019).

As of 2004, the only public documentation of Coastal Tailed Frogs near the RMOW was in Brandywine Creek (Leigh-Spencer 2004), presumably from surveys before the construction of the Independent Power Project (IPP) built on that creek. In late 2004, the Whistler Biodiversity Project began the first valley-wide survey of breeding populations (tadpoles) in 16 creeks in the area (Wind 2005-2009; Brett 2007). Surveys conducted since then, as part of the Environmental Monitoring Program (Cascade 2013-2015; Palmer and Snowline 2017-2019), have continued to expand our understanding of the distribution and abundance of this species. In 2017, Coastal Tailed Frogs were meanwhile down-listed in BC from Blue (Special Concern) to Yellow (Secure; CDC 2020). It remains a species of Special Concern under the Species at Risk Act (Government of Canada 2020).

³ Adults typically breed in the stream in which they hatched.

5.2 Methods

5.2.1 Site Selection

Most creeks surveyed since 2013 have been surveyed in at least two successive years; others have been surveyed three or more times. Sites since 2016 have been selected to: (a) allow results to be compared from year to year; (b) have the greatest geographic coverage that includes both east-side and west-side creeks; and/or (c) because adjacent developments could potentially affect them. It is important to include creeks on both sides of Whistler Valley since creeks on the east side of the valley are more likely to be glacier fed than those on the west side. Creeks with a glacial source typically have higher and more sustained flows. They are also more sensitive to climate change since glacier melt has and will reduce the volume and timing of water flows. Sites were again selected in 2019 to achieve the best balance between these three objectives.

Since it began in 2013, the Ecosystems Monitoring Program has surveyed for tailed frogs in 11 creeks, five of which are on the east side of Whistler Valley and six of which are on the west side (Table 5-1). More creeks on the east side of the valley pass through the ski area and suburban developments which make them easier to access and more susceptible to potential disturbance. For these reasons, more sites have been surveyed since 2013 on the east (47) than west (36) side of the valley. Surveys in 2018 and 2019 have begun to address this imbalance by sampling more creeks on the west side.

Table 5-1. Coastal Tailed Frog sampling sites, 2013 to 2015 (Cascade) and 2016 to 2018 (Palmer and Snowline)

| Creek | Valley Side | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | Total |
|---------------------|-------------|------|------|------|------|------|------|------|-------|
| Agnew Creek | West | | | | | 3 | 3 | | 6 |
| Alpha Creek | East | 3 | 3 | 3 | 3 | | | | 12 |
| Archibald Creek | East | | 3 | 3 | 3 | 3 | 3 | 3 | 18 |
| Blackcomb Creek | East | | | | | | | 1 | 1 |
| FJ West Creek | West | | | | | | 2 | 3 | 5 |
| Horstman Creek | East | | | | | 3 | | | 3 |
| Nineteen Mile Creek | West | | 2 | 2 | | | | | 4 |
| Scotia Creek | West | 3 | 3 | 3 | 3 | | 1 | | 13 |
| Sproatt Creek | West | | | | | | 1 | 3 | 4 |
| Van West Creek | West | | | | | | 2 | 2 | 4 |
| Whistler Creek | East | | | | 4 | 3 | 3 | 3 | 13 |
| | Total | 6 | 11 | 11 | 13 | 12 | 15 | 15 | 83 |
| | East | 3 | 6 | 6 | 10 | 9 | 6 | 7 | 47 |
| | West | 3 | 5 | 5 | 3 | 3 | 9 | 8 | 36 |

The 2016 program expanded and standardized the elevational range of sites. The program continued the previous approach of surveying three reaches on each creek but changed some sampling sites to achieve a standardized range in which one site was near valley bottom, one at approx. 800 m and one at approx. 1000 m. This elevational range was meant to include one site within the development footprint, a second

at the upper end of it and a third above the development footprint (as a control site). Due to access and/or topography of the area, it was not feasible to establish equivalent elevations on some creeks (Figure 5-1).

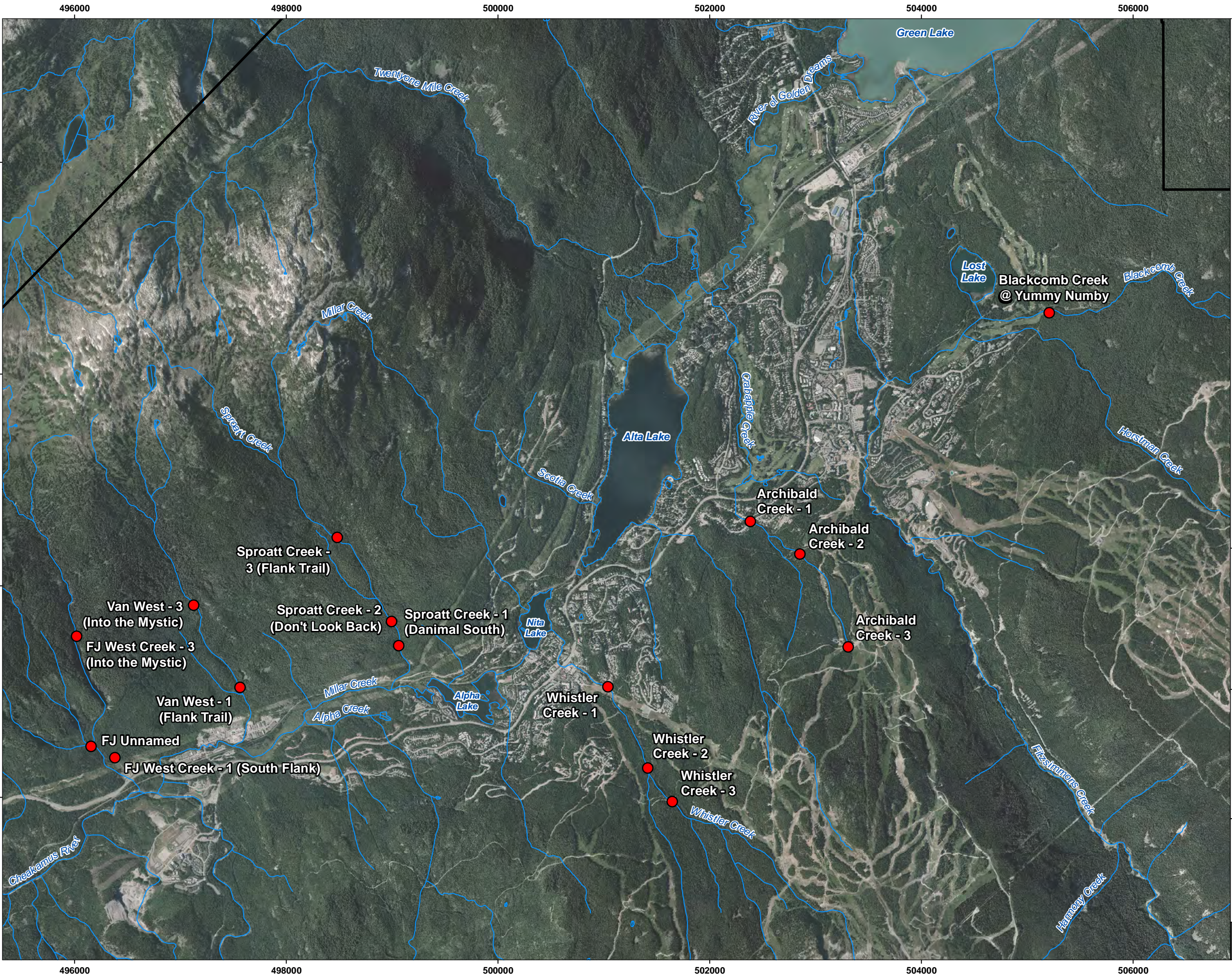
The 2017 program retained Archibald Creek and Whistler Creek to continue multi-year comparisons in these heavily used areas on Whistler Mountain. Two new creeks, Horstman Creek and Agnew Creek were added as replacements for Alpha and Scotia creeks). Alpha and Scotia creeks had been surveyed extensively in the past by the WBP and Cascade (Wind 2005-2009; Cascade 2014-2016) and detections of tailed frogs remained relatively similar in 2016. Horstman Creek was added in 2017 as it detected many tailed frogs in surveys conducted by the WBP in 2006 yet had not been surveyed since. This site was added to increase the spatial distribution of creeks northward, as well as to add another monitoring year to a creek within the ski area footprint. Agnew Creek was also added to the 2017 program to increase the representation of creeks on the west side of Whistler Valley. This area has relatively few creeks that are easily accessible and/or suitable for standard sampling methods. Prior to 2017, Agnew Creek had not previously been surveyed.

In 2018, 15 sites were surveyed on seven creeks; more than any year to date in the Ecosystem Monitoring Program. Whistler and Archibald creeks were again retained in the survey to allow multi-year comparisons, especially as the Whistler Bike Park continued to expand in those areas. Agnew Creek was retained for a second year to detect the presence of tailed frogs, while Horstman Creek was rotated out of the program to allow new creeks to be surveyed. Most notably, the 2018 survey included the sampling of three new creek sites on the west side of the valley: FJ West Creek (two sites), Sproatt Creek (one site), and Van West Creek (two sites).

As in 2018, the 2019 surveys included 15 sites on seven creeks. Archibald Creek has now been surveyed each year since 2013 to monitor impacts from the Whistler Bike Park and other potential disturbances. Whistler Creek was surveyed for the fourth straight year for the same reason. Agnew Creek was rotated out of the surveying program after two years of non-detections.

The main emphasis for the 2019 program was to establish three sites on each of the three west-side creeks newly surveyed in 2018. That goal was achieved on Sproatt Creek (two new sites) but not on the other two creeks. Chasms in the middle section of Van West Creek prevented establishing a third site, at least in 2019. Even more difficult terrain above the valleybottom break in slope on FJ West Creek likely precludes surveys in that area ever. Even below the break in slope, the streambeds for FJ West Creek and the adjacent creek to the west (FJ Unnamed) were significantly changed by a blowout during a fall 2017 storm (discussed in Section 5.3.3.3). The exact location of the lower streambeds for these two creeks no longer matches available mapping and prevented establishing a third site on FJ West Creek. Instead, a site was surveyed on FJ Unnamed Creek. These additions brought the number of west-side creeks surveyed in 2019 to eight.

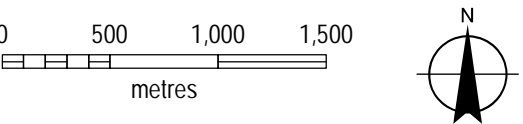
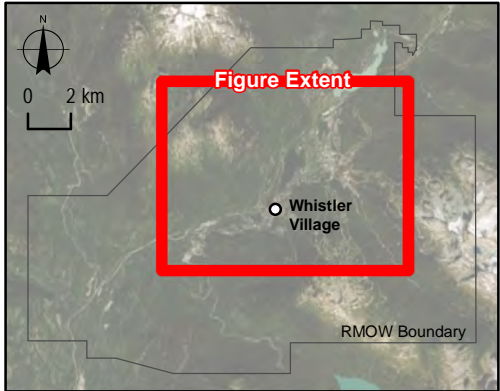
A seventh east-side site, Blackcomb Creek was also surveyed in 2019. Blackcomb Creek was last surveyed in 2006 at four sites from valleybottom to 1377 m (Wind 2006) and no tadpoles were detected. This is the coldest creek surveyed to date: 4.0°C at 1377 m and only 6.3°C at 859 m (at the RMOW water intake). The goal of including a site on this creek was to test if current temperature regime could support a tailed frog population. Even if not, future climate change could melt the Blackcomb Glacier enough that the creek becomes warm enough to support tailed frogs which adds to the value of monitoring it.



Legend

- 2019 Tailed Frog Site
- Watercourse
- RMOV Boundary

Notes:
(1) Orthoimagery (2014) provided by RMOV.
(2) Watercourse data from BC Freshwater Atlas (accessed 2017).
Contains information licensed under the Open Government Licence - British Columbia. Service Layer Credits: Source: Esri, DigitalGlobe, GeoEye, Earthstar Geographics, CNES/Airbus DS, USDA, USGS, AeroGRID, IGN, and the GIS User Community.



DRAWN: B. Elder
CHECKED: B. Brett
PROJECT: 1602505
DATE: Jan 13, 2020
Scale 1:35000
UTM Zone 10N
NAD 1983

Palmer™

CLIENT: Resort Municipality of Whistler
PROJECT: Annual Monitoring 2019

Coastal Tailed Frog
Sampling Sites, 2019

FIGURE 5-1

5.2.2 Sampling Design

All previous surveys for tailed frog tadpoles in the RMOW study area by the WBP (Wind 2005-2009) and the RMOW's Environmental Monitoring Program (Cascade 2014-2016; Palmer and Snowline 2017-2019) have followed similar methodologies in conducting Coastal Tailed Frog surveys. The only variation in in sampling design since 2005 occurred in 2013 to 2015 surveys when Cascade used an area-constrained approach in contrast to the time-constrained approach used by the WBP (2005 to 2009) and surveys since 2016.

The area-constrained approach can be used to measure relative abundance as recommended by the BC Resource Inventory Committee (BC MELP 2000). Surveys from 2013 to 2015 therefore surveyed at fixed 5 m stream lengths for a total of 30 minutes (Cascade 2014-2016). Whistler-area streams, however typically have low tadpole densities and the area-constrained method had low detections result in a low number of detections,⁴ and definitely lower than previous WBP results from time-constrained surveys (Wind 2005-2009). Surveys since 2016 have therefore returned to the time-constrained approach of 30 minutes total sampling time, regardless of area, which greatly increased detections and therefore statistical power (Malt et al. 2014a, 2014b) of the study.⁵

Both approaches measure total sampling area at each site and survey for the same amount of time (30 minutes per site) which allows direct comparisons between years, regardless of method (within some statistical limitations). In addition, the total area surveyed at each site since 2016 with the time-constrained approach remained remarkably similar to that surveyed using the area-constrained approach. This similarity should also increase the reliability of comparisons between the two approaches.

Data collection methods were otherwise the same for all tailed frog surveys since 2004 and followed recommendations of the BC Resource Inventory Committee (BC MELP 2000). The in-stream surveys consisted of overturning unembedded cover objects such as rocks with dipnets held immediately downstream to catch any dislodged animals (Photo 5-1, Photo 5-2). Rocks were also swept by hand to detect any clinging tailed frog larvae before being set back in their original positions, as were large anchored rocks and large woody debris. Data collected at each site included:

- Site characteristics such as location, weather, overhead cover and stand type;
- Stream characteristics such as morphology, substrate size and shape, slope and bankfull and wetted widths;
- Overhead canopy cover, forest type (coniferous, deciduous, or mixed) and forest successional stage;
- Water and air temperature; and
- Total survey area (measured with a cloth tape to the nearest 0.1 m).

⁴ Bruce Bury (in a 2016 email to Brent Matsuda and Bob Brett) recommends that detections should be >2 tadpoles/m² to ensure statistical power. Virtually all sites sampled to date in Whistler have revealed densities far lower.

⁵ These increases are reported in a multi-year comparison included in the results section (Section 5.3).



Photo 5-1. Hillary Williamson dipnetting for tadpoles in Whistler Creek.



Photo 5-2. Captured tadpoles are transferred to a bucket until they are measured, classified to cohort and development stage, and released upstream

Data collected for tadpole captures also followed standard methods, including a measurement of total length for tadpoles (snout to ventral length for later stages). From 2013 to 2015⁶ and again in 2016, tadpoles were classed into cohorts defined by Malt et al (2014a,b) which served as proxies for age classes (e.g., first year - T1; second year - T2, etc.) as follows:

- T0 (hatchling <15 mm);⁷
- T1 (tadpole, no visible hind legs);
- T2 (tadpole, recognizable hind legs with knees that do not extend beyond the anal fold (Photo 5-3);
- T3 (tadpole, conspicuous hind legs with knees that extend out from body (Photo 5-4); and
- Non-tadpole – metamorph (tail plus front legs), juvenile (no tail, small, no nuptial pads); and adult (larger than juvenile, male has tail and nuptial pads, females larger than males).

Doubts about this classification scheme emerged in 2016 regarding how closely these classes reflected true and consistent age classes, especially in different streams. Tests conducted prior to 2017 surveys again revealed overlaps between length and developmental stages within and between streams. These observations intensified questions about whether developmental stages were reliable proxies for the number of years since hatching, especially between streams that have different growing conditions. This doubt was later strengthened by Pierre Friele⁸ who emphasized that the link between developmental stage, length and age is even more tenuous when applied across large geographic gradients in which climate differs. As a result, surveys since 2017 measured the length of each tadpole and classified them by more detailed developmental stages as follows:

⁶ Candace Rose-Taylor, 2016 email to Bob Brett.

⁷ No hatchlings have been reported to date in Whistler surveys conducted in late August and September.

⁸ Pierre Friele email to B. Brett and follow-up phone conversation, December 2017.

Table 5-2. Tadpole Developmental Stages and Classifications

| Developmental Stage | Cohort (Malt 2014a,b) Equivalent |
|--|---|
| 0 – Hatchling <15 mm | T0 |
| 1 - No visible hind legs | T1 |
| 2 - Bulge only, hind legs not defined | T1 |
| 3 - Hind legs visible but covered | T2 |
| 4 - Hind feet protruding | T2 |
| 5 - Hind knees protruding outside body | T3 |



Photo 5-3. Tadpole cohort 2 (T2). This individual's developmental stage is transitional between developmental stages 2 and 3 (hind legs covered but just starting to be defined).



Photo 5-4. Tadpole cohort 3 (T3); and developmental stage 5 (hind knees protruding outside body).

For consistency with past reports, the classes above were grouped according to Malt et al's (2014a, 2014b) cohorts. That is, Developmental Stages 1 and 2 were grouped into Malt's T1 cohort, and Developmental Stages 3 and 4 were grouped into Malt's T2 cohort. Future analyses may be able to use these detailed classifications to calibrate a reliable relationship between age and developmental stage in Whistler-area creeks. For the purposes of this report, most of the analysis and discussion is based on Malt's cohorts.

To prevent recaptures, all tadpoles were placed in buckets and released upon completion of the site survey (BC MELP 2000). Non-tadpoles, or post metamorphosis individuals, were classed as metamorphs (non-resorbed tail), juveniles (no tail, smaller than adults, no nuptial pads on males) or adults (larger than juveniles, males have a cloacal "tail," nuptial pads and are smaller than females; Corkran and Thoms 1996; Jones et al. 2005). Early September was the targetted survey window, late enough in the season that low streamflow would increase the detectability of tadpoles.

5.2.3 Data Analysis

The total number of tadpoles per site (reach) detected in 2019 was compared to surveys since 2015 (the last year of the time-constrained approach). Results were also reported as detections per unit area (per 100 m²) to permit comparisons between the 2015 area-constrained method and the time-constrained method used for the past four surveys.

5.2.4 Quality Assurance/Quality Control

Although the ideal way to ensure consistency between sites and years would be to use the same surveyor(s), that is seldom achievable due to changes in available personnel. To maximize consistency in 2019, the two main surveyors from 2018 were again used for all but two sites at which one of the surveyors was unavailable and replaced by another who had participated in 2017. A trial survey was first used to ensure that measurements were consistent between surveyors. Special care was taken to ensure that cohort classes and developmental stages (see above) were recorded consistently. Photos of representative tadpoles in each class were used as guides to improve consistency between surveyors (e.g. Photo 5-3 and Photo 5-4).

5.3 Results and Discussion

5.3.1 Tadpole Surveys

Fifteen sites were surveyed from September 3 to 6, 2019 (Figure 5-1, Table 5-4; Appendix E). Seven sites on three creeks were on the east side of the valley and eight sites on four creeks were on the west side of the valley. As expected, Blackcomb Creek was the coldest creek surveyed. While surveys during warm, sunny weather are preferable to maximize detections, weather changes during the survey window meant that four sites were sampled in non-sunny weather. Air temperatures were nonetheless fairly warm at these sites which presumably offset any potential reduction in detections (Appendix E).

Table 5-3. Coastal Tailed Frog sampling sites, 2019.

| Site | Valley Side | Date | Easting | Northing | Elev. (m) | Weather | Water Temp. (°C) | Air Temp. (°C) |
|-------------------------------------|-------------|------------|---------|----------|-----------|----------|------------------|----------------|
| Archibald Creek - 1 | East | 2019-09-04 | 502387 | 5550606 | 695 | Sunny | 11.4 | 18 |
| Archibald Creek - 2 | East | 2019-09-04 | 502854 | 5550298 | 835 | Sunny | 11.2 | 15 |
| Archibald Creek - 3 | East | 2019-09-04 | 503310 | 5549422 | 1026 | Sunny | 9.4 | 17 |
| Blackcomb Creek @ Yummy Numby | East | 2019-09-06 | 505211 | 5552576 | 762 | Sunny | 8.0 | 11 |
| FJ Unnamed | West | 2019-09-05 | 496157 | 5548481 | 699 | Cloudy | 11.0 | 18 |
| FJ West Creek - 1 (South Flank) | West | 2019-09-05 | 496383 | 5548374 | 648 | Cloudy | 11.2 | 18 |
| FJ West Creek - 3 (Into the Mystic) | West | 2019-09-03 | 496022 | 5549522 | 1119 | Sunny | 11.3 | 14 |
| Sproatt Creek - 1 (Danimal South) | West | 2019-09-03 | 499063 | 5549434 | 692 | Lt. Rain | 12.9 | 16 |

| Site | Valley Side | Date | Easting | Northing | Elev. (m) | Weather | Water Temp. (°C) | Air Temp. (°C) |
|-------------------------------------|-------------|------------|---------|----------|-----------|----------|------------------|----------------|
| Sproatt Creek - 2 (Don't Look Back) | West | 2019-09-03 | 498996 | 5549662 | 790 | Lt. Rain | 12.3 | 17 |
| Sproatt Creek - 3 (Flank Trail) | West | 2019-09-03 | 498483 | 5550455 | 996 | Sunny | 12.0 | 15 |
| Van West - 1 (Flank Trail) | West | 2019-09-05 | 497563 | 5549038 | 706 | Sunny | 12.5 | 16 |
| Van West - 3 (Into the Mystic) | West | 2019-09-03 | 497125 | 5549816 | 1036 | Sunny | 11.7 | 14.5 |
| Whistler Creek - 1 | East | 2019-09-06 | 501041 | 5549045 | 692 | Sunny | 11 | 11 |
| Whistler Creek - 2 | East | 2019-09-05 | 501417 | 5548276 | 879 | Sunny | 10 | 11 |
| Whistler Creek - 3 | East | 2019-09-05 | 501649 | 5547961 | 972 | Sunny | 10.2 | 11 |

A total of 60 tadpoles were detected in 2019 (Table 5-4, Appendix F). This total is lower than in the past two years but still higher than in 2016 (39) and much higher than when the surveys used an area-constrained approach (only nine tadpoles were detected in the September 2015 survey). Fluctuations between yearly totals at two creeks, Archibald Creek and Whistler Creek, explain most of the difference in annual totals since 2016. These have been the two most productive creeks throughout the surveys. In contrast, only two tadpoles were detected at four sites located on FJ West Creek, FJ Unnamed Creek, and Blackcomb Creek. This is the second year of that only one tadpole was found at FJ West Creek. No hatchlings (T0 cohort = <15mm length) were detected in 2019 which is consistent with all September surveys conducted to date in the RMOW. Contrary to most years, no juveniles (metamorphs) or adults were detected at any site in 2019.

Table 5-4. Tadpole surveys by creek, 2015-2019.

| Survey Year | Valley Side | Site | No. of Sites | Total Survey Area (m ²) | Average Survey Area (m ²) | Number of Tadpoles Detected | Tadpoles /100m ² | Average Water Temp. (°C) |
|-------------|-------------|-----------------------|--------------|-------------------------------------|---------------------------------------|-----------------------------|-----------------------------|--------------------------|
| 2015 | East | Alpha Creek | 3 | 69.6 | 23.2 | 4 | 5.7 | 7.5 |
| | East | Archibald Creek | 3 | 46.9 | 15.6 | 4 | 8.5 | 8.7 |
| | West | Scotia Creek | 3 | 45.8 | 15.3 | 1 | 2.2 | 8.8 |
| | West | 19 Mile Creek | 3 | 73.6 | 24.5 | 0 | 0.0 | 7.9 |
| | | All 2015 Sites | 12 | 235.9 | 19.7 | 9 | 3.8 | 8.2 |
| 2016 | East | Alpha Creek | 3 | 72.5 | 24.2 | 9 | 12.4 | 7.0 |
| | East | Archibald Creek | 3 | 45.2 | 15.1 | 5 | 11.1 | 6.4 |
| | West | Scotia Creek | 3 | 86.7 | 28.9 | 3 | 3.5 | 10.1 |
| | East | Whistler Creek | 4 | 97.6 | 24.4 | 22 | 22.5 | 8.8 |
| | | All 2016 Sites | 13 | 302.0 | 23.2 | 39 | 12.9 | 8.1 |
| 2017 | West | Agnew Creek | 3 | 56.2 | 18.7 | 0 | 0.0 | 8.8 |
| | East | Archibald Creek | 3 | 88.2 | 29.4 | 33 | 37.4 | 12.0 |
| | East | Horstman Creek | 3 | 56.2 | 18.7 | 6 | 10.7 | 9.3 |

| Survey Year | Valley Side | Site | No. of Sites | Total Survey Area (m ²) | Average Survey Area (m ²) | Number of Tadpoles Detected | Tadpoles /100m ² | Average Water Temp. (°C) |
|-------------|-------------|-----------------------|--------------|-------------------------------------|---------------------------------------|-----------------------------|-----------------------------|--------------------------|
| | East | Whistler Creek | 3 | 36.2 | 12.1 | 48 | 132.6 | 13.0 |
| | | All 2017 Sites | 12 | 236.8 | 19.7 | 87 | 36.7 | 10.8 |
| 2018 | West | Agnew Creek | 3 | 82.3 | 18.7 | 0 | 0.0 | 8.1 |
| | East | Archibald Creek | 3 | 55.5 | 18.7 | 30 | 54.1 | 8.1 |
| | West | FJ West Creek | 2 | 18.0 | 18.7 | 1 | 5.6 | 9.0 |
| | West | Scotia Creek | 1 | 9.5 | 18.7 | 2 | 21.1 | 9.0 |
| | West | Sproatt Creek | 1 | 19.5 | 18.7 | 11 | 56.4 | 9.1 |
| | West | Van West Creek | 2 | 30.0 | 18.7 | 17 | 56.7 | 10.0 |
| | East | Whistler Creek | 3 | 89.0 | 18.7 | 21 | 23.6 | 8.1 |
| | | All 2018 Sites | 15 | 303.8 | 18.7 | 82 | 27.0 | 8.8 |
| 2019 | East | Archibald Creek | 3 | 60.5 | 20.2 | 14 | 23.1 | 10.7 |
| | East | Blackcomb Creek | 1 | 23.5 | 23.5 | 0 | 0.0 | 8.0 |
| | West | FJ Unnamed | 1 | 23.5 | 23.5 | 1 | 4.3 | 11.0 |
| | West | FJ West Creek | 2 | 32.0 | 16.0 | 1 | 3.1 | 11.3 |
| | West | Sproatt Creek | 3 | 48.0 | 16.0 | 11 | 22.9 | 12.4 |
| | West | Van West Creek | 2 | 25.0 | 12.5 | 6 | 24.0 | 12.1 |
| | East | Whistler Creek | 3 | 51.9 | 17.3 | 27 | 52.0 | 10.4 |
| | | All 2019 Sites | 15 | 264.4 | 17.6 | 60 | 22.7 | 10.8 |

Several conclusions can be drawn from a comparison of surveys conducted over the past five years (Figure 5-2):

- Survey area per site has been relatively consistent, even including the switch from area-constrained searches in 2015 to time-constrained searches since.
- Detections increased with the switch to time-constrained searches in 2016 and were even higher between 2017 and 2019.
- The 2018 report showed a possible though weak connection between water temperature and detections. Results from 2019 surveys, which had lower detections than in 2018 in spite of warmer water, weaken this hypothesis.

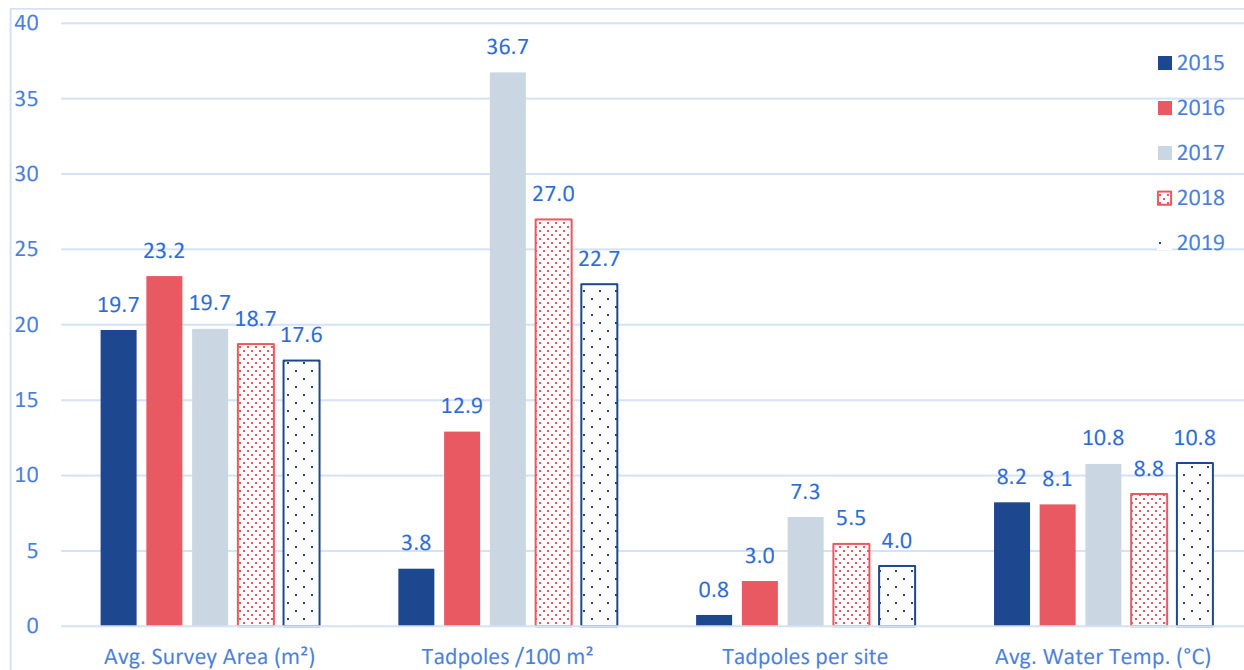


Figure 5-2. Average survey area per site, number of tadpoles per 100 m², number of tadpoles per site and average water temperature for September Coastal Tailed Frog surveys, 2015-2019.

The proportion of tadpole by cohort has fluctuated since 2016 with no clear pattern emerging (Table 5-5). Statistical analysis for this relatively small number of detections would have low power to reliably detect trends but one observation is still possible. Even without statistics it is reasonable to conclude that there is no indication of lower survivorship over those years. The proportion of older (T3) tadpoles has been relatively stable and the absolute number has actually increased each year.

Table 5-5. Tadpole detections by year, site, elevation and cohort, 2016-2019.

| Year | Cohort (No. of Tadpoles) | | | Cohort (% of Tadpoles) | | |
|-------|--------------------------|----|----|------------------------|-----|-----|
| | T1 | T2 | T3 | T1 | T2 | T3 |
| 2016 | 25 | 5 | 9 | 64% | 13% | 23% |
| 2017 | 63 | 11 | 13 | 72% | 13% | 15% |
| 2018 | 64 | 2 | 16 | 78% | 2% | 20% |
| 2019 | 26 | 14 | 20 | 43% | 23% | 33% |
| Total | 178 | 32 | 58 | 66% | 12% | 22% |

Surveys in 2016 recorded tadpoles by cohorts recommended by Malt (2014a, b), that is, T0, T1, T2, and T3. Since 2017, surveyors have recorded more detailed developmental stages that broadly coincide with Malt's cohorts but potentially increase the ability to ensure they are true cohorts (see Section 5.2.2 for more details).

A comparison of the 229 tadpoles captured since 2017 shows that while there is a relationship between length and developmental stage, there is also significant overlap between classes, whether Malt's three cohorts or the five more detailed developmental stages also recorded for this study (Table 5-6). It is clear

from this comparison that length alone is a poor proxy for developmental stage, regardless of how detailed that classification is. Another conclusion from this data is that Developmental Stages 1 and 2 are similar enough that grouping them together in Malt's cohort T1 is reasonable. Similarly, Developmental Stage 5 probably separates out strongly enough from Developmental Stage 4 to show that Malt's cohort T3 is distinct, except when the tadpoles are transitional between the two (though see below). The largest overlaps are in the middle stages: Developmental Stages 3 and 4 and their assumed equivalent, cohort T2.

Table 5-6. Length comparisons between detailed developmental stages (upper) and age classes/cohorts (lower; Malt et al 2014a, b).

| | Development Stage | | | | |
|-------------------------------|-------------------|---------------------------------------|----------------------------------|-------------------------|--------------------------|
| | 1 - No hind legs | 2 - Bulge only, hind legs not defined | 3- Hind legs visible but covered | 4- Hind feet protruding | 5 -Hind knees protruding |
| Number of Tadpoles | 63 | 90 | 15 | 33 | 28 |
| Mean Length (mm) | 31 | 35 | 39 | 45 | 50 |
| Median Length (mm) | 30 | 35 | 38 | 45 | 50 |
| Smallest (mm) | 25 | 27 | 33 | 40 | 43 |
| Largest (mm) | 43 | 45 | 50 | 54 | 60 |
| Length Range (mm) | 25 to 43 | 27 to 45 | 33 to 50 | 40 to 54 | 43 to 60 |
| Largest to Smallest | 1.7 | 1.7 | 1.5 | 1.4 | 1.4 |
| Cohort (Malt et al. 2014a, b) | | | | | |
| | T1 | T2 | | T3 | |
| Number of Tadpoles | 153 | 48 | | 28 | |
| Mean Length (mm) | 33 | 43 | | 50 | |
| Median Length (mm) | 33 | 44 | | 50 | |
| Smallest (mm) | 25 | 33 | | 43 | |
| Largest (mm) | 45 | 54 | | 60 | |
| Length Range (mm) | 25 to 45 | 33 to 54 | | 43 to 60 | |
| Largest to Smallest | 1.8 | 1.6 | | 1.4 | |

Notes: No hatchlings <15mm (T0 or development stage 1) have yet been detected in a September survey in Whistler.

The most difficult classifications have been between tadpoles demonstrating intermediate stages between cohort 1 and 2 and between cohort 2 and cohort 3 classifications. For cohort 1 and 2, many tadpoles were transitional between having an undefined "bulge" and defined legs contained within that bulge (Photo 5-4). Between cohort 2 and 3, there were some tadpoles whose rear feet but not knees were free of the skin that covered the bulge. They were transitional to cohort 3 but without the exact characteristics described by Malt *et al.* (2014a,b). These observations and the ones above suggest caution in interpreting age cohorts from length and/or development stage, especially between streams.

Almost three times as many tadpoles have been detected in east-side than west-side creeks since 2016. (Table 5-7). As discussed in Section 5.2.1, glacier-fed creeks are predominantly on the east side of Whistler Valley. Glacial run-off increases overall flow and also means that post-snowmelt fluctuations in flow are reduced compared to creeks reliant solely on rainwater. Creeks on the east side of the valley are therefore more likely to be larger and, as found in these surveys, seem overall to offer better habitat characteristics

including: more cobbles, less embeddedness, and more riffles. These are preliminary conclusions that need to be further tested, especially since the predominance of detections from two creeks (Whistler and Archibald) affects the totals so much. Although data to date showed east side creeks to be slightly colder than west-side creeks, temperature loggers (Section 2) provide more robust data for the actual magnitude of that difference.

Table 5-7. Tadpoles detected in east-side versus west-side creeks since 2016.

| Valley Side | No. Sites | Mean Elev. (m) | Mean Tadpoles /Site | Mean Water Temp. (°C) |
|-------------|-----------|----------------|---------------------|-----------------------|
| East | 32 | 850 | 6.7 | 9.3 |
| West | 23 | 809 | 2.3 | 10.1 |

5.3.2 Remarks on Notable Streams Surveyed in 2019

5.3.2.1 Archibald Creek

In 2016, significant depositions of sand and small gravel occurred in Archibald Creek below the main part of the Whistler Bike Park (Photo 5-5; Photo 5-6). The deposition was especially deep at the lowest reach, Archibald 1, located uphill of Panorama Drive in Brio. This site was downstream of the data logger that became clogged with sand and gravel in 2016. Low detections that year were attributed to two possible causes: (i) the sedimentation; and/or (ii) low water temperatures.



Photo 5-5. Sedimentation in 2016 at Archibald Creek-1 (near Panorama Drive).

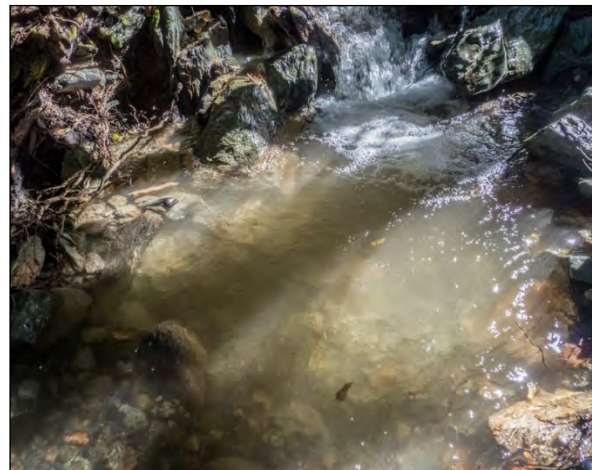


Photo 5-6. Sedimentation in 2016 at Archibald Creek-2 (near Crank It Up in the Whistler Bike Park).

No evidence of negative impacts on tailed frogs was detected in 2019, from either the 2016 event or from a rain-caused flush of sediments after a long dry spell in August 2018 (Photo 5-7 and Photo 5-8). The most tadpoles in the four years were recorded in 2017 and 2018, each with more twice as many as the 14 recorded in 2019. As in past years, pre-survey reconnaissance found variable numbers of tadpoles on bedrock in the main flow and the low number recorded during the actual survey were likely at the low end of those pre-surveys. As a result, there is no compelling evidence that there has been a significant (if any)

decline in the tailed frog population in this creek. In addition, even if there have been negative impacts on the creek from the Whistler Bike Park and/or other sources, they have (so far) not been significant enough to prevent a viable and potentially stable population of tailed frogs in the creek.



Photo 5-7. Sedimentation at Archibald Creek-1 on August 29, 2018, likely caused by runoff from a heavy rainfall that was the first significant precipitation in many weeks.



Photo 5-8. A photo near the same location as Photo 5-7 on Archibald Creek-1 one week later (September 6, 2019).

5.3.2.2 Whistler Creek

Whistler Creek was added to the monitoring program in September 2016 (when the Whistler Bike Park was being extended into this drainage) and no direct impacts on the creek were then noted. Sometime between then and 2017 surveys, however, machines imported and moved rocks at the lowest elevation site (Whistler Creek-1, located below the Snowridge pedestrian bridge). This work removed streamside vegetation and channelized the creek (Photo 5-9). The channel was also filled with large angular rocks that replaced many of the cobbles that were previously in the stream. The detections of tadpoles in the 2017 survey were nonetheless higher than those in 2016 despite the apparent degradation of habitat. By September 2018, the stream had mostly reverted to its pre-disturbed condition, presumably because the imported rocks had been washed downstream. By September 2019 no obvious evidence of that in-stream work was visible (Photo 5-10). There was also no evidence of negative impacts (such as turbidity or sedimentation) in Whistler Creek from the expanded trail network or increased traffic associated with the Whistler Bike Park..



Photo 5-9. Significant in-stream disturbance occurred at the Whistler Creek 1 site before the 2017 surveys.



Photo 5-10. The Whistler Creek 1 streambed has mostly returned to an undisturbed appearance by September 2019.

5.3.2.3 Blackcomb Creek

Blackcomb Creek, surveyed for the first time in 2006 from valleybottom to 1377 m, is the coldest creek surveyed to date. On August 25, 2006 it was only 4.0°C at 1377 m and 6.3°C at 859 m (at the RMOW water intake; Wind 2006) and no tadpoles were detected even at the two warmer sites below. Since temperatures below 5.0°C are inhospitable for egg development (Section 5.1), it is possible this creek may not support tailed frog reproduction; at least until run-off from the melting Blackcomb Glacier diminishes enough to reduce its cooling effect. This creek therefore provides a monitoring opportunity. If additional surveys continue to find no tadpoles, it is probable that the creek is too cold. If so, tailed frogs would only colonize this creek after the glacier melts to a point at which it has less influence on stream temperatures than relatively warmer sources (mainly rain, but also non-glacial tributaries). If proved true, this hypothesis would demonstrate direct effects of climate change on local habitat and species.

One test site was therefore surveyed under the bridge leading to Yummy Mummy Trail and at 762 m elevation (Photo 5-11). The glacial flour in the creek reflected its glacial source (Photo 5-12). This site was almost 100 m elevation below the water intake site surveyed in 2006 and measured 8.0°C, slightly warmer than that site in 2006. Although within a temperature range that could successfully allow egg development, no tadpoles were detected. Higher elevations on this creek are difficult to access but should ideally be added in future years' surveys.



Photo 5-11. Blackcomb Creek at the Yummy Numby bridge, looking upstream.



Photo 5-12. Blackcomb Creek at the Yummy Numby bridge, looking downstream.

5.3.3 Inconsistencies in Stream Mapping

The 2018 work plan included tailed frog sampling for the first time in small creeks accessed by the Flank Trail and new Sproatt Mountain Trails (e.g., Into the Mystic). After extensive reconnaissance, referencing of various maps, and field surveys, it became evident that the mapping of some west-side creeks was incorrect and/or that they did not have surface flow in late summer. While not originally within the scope of this monitoring project, the 2019 work plan included further investigation of these inconsistencies on three creeks: Sproatt, Van West, and FJ West.

5.3.3.1 Sproatt Creek

RMOW mapping⁹ for Sproatt Creek shows two outlets, the main channel that feeds into Miller Creek and a second that feeds southeast into Alpha Lake (Figure 5-2). Fieldwork in 2018 showed that the year-round flow at higher elevations of Sproatt Creek disappears in low flow (e.g., late summer) somewhere downstream of the break in slope and upstream of the CN Rail tracks.

Surveys in 2019 confirmed the exact spot that the last water flowed underground, approximately 110 m upstream of the CN Rail tracks (Photo 5-13; Figure 5-2). It also confirmed that the second outlet mapped as flowing southeast to Alpha Lake is not (or no longer) connected to the main stem of Sproatt Creek. A significant berm at the mapped junction with the main stem currently prevents any flow in that direction. The lack of year-round above-ground connection of Sproatt Creek with Miller Creek presumably does not negatively impact tailed frogs since: (a) an apparently healthy population inhabits the creek; and (b) valleybottom creeks like Miller Creek that support fish typically provide less hospitable habitat for tailed frogs due to predation.

⁹ <https://webmap.whistler.ca/HTML5Viewer/Index.html?viewer=ExternalGIS>



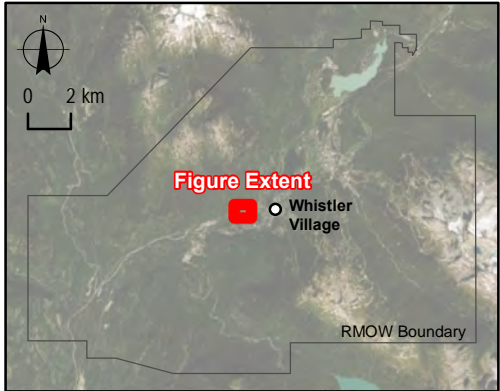
Photo 5-13. The last above-ground flow of Sproatt Creek disappears underground (as marked by arrow) ~110 m upstream of the railway tracks on September 6, 2019. See Figure 5-2 for location and context.



- Legend**
- Flowing Creek
 - Dry Creek (September, 2019)
 - Flow Direction
 - Contour (10 m)

Notes:
(1) Orthoimagery (2014) provided by RMOW.
(2) Watercourse data provided by RMOW (modified).

Service Layer Credits: Source: Esri, DigitalGlobe, GeoEye, Earthstar Geographics, CNES/Airbus DS, USDA, USGS, AeroGRID, IGN, and the GIS User Community.



0 25 50 75 100 metres

Scale 1:2500
UTM Zone 10N
NAD 1983

DRAWN: B. Elder
CHECKED: B. Brett
PROJECT: 1602505
DATE: Jan 13, 2020

Palmer™

CLIENT: Resort Municipality of Whistler
PROJECT: Annual Monitoring 2019

**Sproatt Creek
– Seasonal
Disappearance**

FIGURE 5-3

5.3.3.2 Van West Creek

Observations during the 2018 surveys also raised the possibility that at least some of the flow of Van West Creek is diverted underground below the break in slope that occurs at the uphill edge of Function Junction (specifically, just uphill of the Rona rear parking lot). On September 5, 2019, the flow at the Van West Creek-1 site at 706 m appeared to be much stronger than that observed just upstream (at 604 m) of that creek's confluence with Miller Creek in Function Junction (Photo 5-14).

An attempt to locate that possible diversion on September 26, 2019 was unsuccessful because heavy rain in the previous days and resulting high flow obscured any possible diversion. Subsequent visits confirmed that streamflow stayed strong for the remainder of fall 2019. Although not conclusive, evidence to date suggests that at least some water is diverted underground between the abrupt break in slope near 613 m and the creek ~100 downstream at 604 m (Photo 5-14). Testing of this hypothesis will require another site visit during low streamflow, ideally in early September 2020.



Photo 5-14. Van West Creek upstream of Function Junction. The top icon at 706 m shows the lowest site surveyed for tailed frogs on Van West Creek in 2018 and 2019 (Van West Creek-1 at the Lower Flank Trail). An abrupt break in slope occurs at the middle icon (~613 m). Water flow at the lower icon (at 604 m) in late summer appeared much lower than at 706 m on September 5, 2019.

5.3.3.3 FJ West Creek

Another puzzle emerged during 2018 tailed frog surveys at “FJ West Creek”¹⁰ when the streambeds on RMOw maps did not always correspond with actual flows. Similar to Sproatt Creek and, to a lesser degree Van West Creek, there is a very abrupt break in slope where the mountainside meets the valleybottom. The 2018 survey at the lower FJ West Creek site (648 m, where it meets with the South Flank Trail) saw many signs of a recent and significant flood, and the main stem of the creek appeared to have moved from its mapped location. A second site could not be established in 2018 partly due to this uncertainty.

Subsequent research showed the evidence seen in 2018 was almost certainly caused by a major rainstorm and flooding on November 23, 2017 (Photo 5-15). This hypothesis was corroborated by Google Earth images showed land scouring in and below the chasm at the break in slope on FJ West Creek sometime between July 4, 2017 and August 6, 2018¹¹ (Photo 5-16). Similar scouring seen in September 2018 at the upper Sproatt Creek survey site (996 m, at the Mid-Flank Trail) was obviously caused by a recent flood, presumably the same November 2017 event.



Photo 5-15. Flooding on November 23, 2017 washed out the railway tracks at Sproatt Creek.

Fieldwork in 2019 was able to confirm that: (a) FJ West Creek does not currently connect with Miller Creek via a southeast branch as currently mapped; (b) that FJ West Creek blew out enough during that storm to create a temporary channel to FJ Unnamed Creek; and that (c) multiple, shallow streambeds now occur in

¹⁰ No official name could be found for this creek nor for the creek adjacent to the west (“FJ Unnamed Creek”).

¹¹ No imagery between these two dates was available.

that area so the exact creek locations are difficult to discern. Further fieldwork in 2020 will help determine the location of active creekbeds and hopefully also establish a third, middle tailed frog site on FJ West Creek.

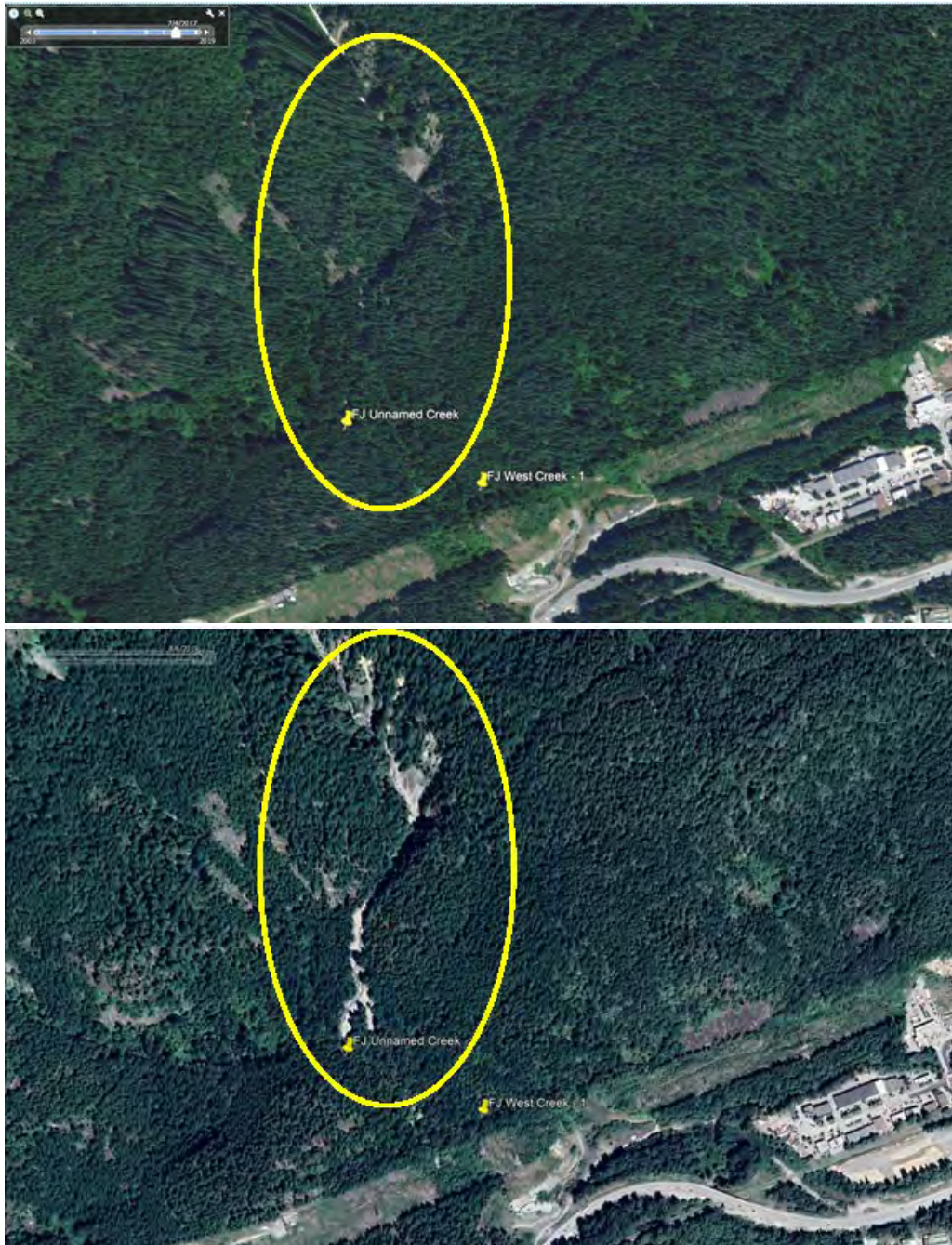


Photo 5-16. FJ West Creek chasm, uphill of the break in slope and west of Function Junction from Google Earth imagery on July 4, 2017 (top); and August 6, 2019 (bottom).



6. Beavers

Lead Biologist and Author: Bob Brett

Additional Surveyors: Kristen Jones and Jagoda Kozikowska

6.1 Introduction

Beavers are a keystone species, second only to humans in their ability to alter Whistler's landscape. The ponds and wetlands created by Whistler's beavers provide important habitat for a wide range of other species including waterfowl, amphibians, snakes, fish, mammals, aquatic plants and insects. Flooding and other damage caused by beavers can bring them into conflict with humans, which is why there is a long history of removing beavers from urban and other habitats.

Beavers are colonial animals. They maintain a family lodge which houses the adult parents and typically two years of offspring, newborns as well as yearlings (Müller-Schwarze and Sun 2003). Two-year-old beavers generally disperse to form new colonies except when quality habitat is already occupied and dispersal is sometimes delayed. A lodge can remain active indefinitely but more often it is periodically inactive or abandoned permanently (as shown by Whistler data). The dispersal of offspring, death and migration of adults indicate that the location of active lodges changes each year within the landscape (here defined as lower elevations in Whistler Valley).

Beavers provide a unique situation for field biologists because it is possible to document all colonies (overwintering lodges) in a valley the size of Whistler. This information, when combined with an estimate of number of beavers per colony, provides a population census that can be monitored without statistical analysis as required in most population surveys (statistical sampling). The human equivalent is the Canada census compared to election polling: the former includes the whole population while the latter includes a small subset and uses statistical analysis to estimate figures for the whole population.

The Whistler Biodiversity Project initiated Whistler's first beaver census in 2007 (Brett 2007; Mullen 2008). Surveys continued through 2011, the last two of which were in conjunction with RMOW staff (Mullen 2009; Pevec 2009; Tayless 2010; Tayless and Burrows 2011). The survey was reinitiated in 2013 as part of this Ecosystem Monitoring Program but focussed only on a subset of lodges (Cascade 2014-2016). The 2016 surveys (Palmer and Snowline 2017) returned to a full census approach where all possible active beaver locations within Whistler Valley were enumerated. The greater survey effort and geographic range that started in 2016 increased the number of documented colonies from nine (2015), 13 (2016), 14 (2017), and 18 (2018). The documentation of inactive lodges and other activity similarly increased. Each year, these surveys have come closer to a full census of all beaver colonies in Whistler.

The 2018 surveys increased the knowledge of active colonies on the River of Golden Dreams and other areas. It also produced the first mapping of wetlands created and/or maintained by beavers ("beaver-affected wetlands"). The main goals for 2019 surveys were to:

- Build on 2018 results to move closer to a full census of beavers in Whistler Valley. Two areas were again targeted for additional survey effort: the River of Golden Dreams and Millar Creek Wetlands;

- Obtain and tabulate historic and recent trapping records. Correlate those trapping records where possible to past and current beaver locations;
- Work with RMOW staff to convey information about beavers to avoid/mitigate conflicts, especially where the new Valley Trail is being built in Function Junction;
- Continue to communicate with local golf courses about beaver activities and possible ways to coexist better with beavers; and
- Field truth beaver-affected wetlands to reconcile the discrepancy between the RMOW wetlands layer and what was mapped for this report. The eventual goals are to be able to: (a) accurately monitor the extent of beaver-affected wetland; and ideally (b) provide a better historic baseline to which that extent can be compared.

6.2 Methods

6.2.1 Sampling Design

Sites included in the 2016 to 2019 surveys were based on the following sources: (i) locations documented in surveys dating back to 2007; (ii) incidental sightings; and (iii) anecdotal reports. Each beaver survey recorded all past and current beaver activity, e.g., freshly cut branches and trees, tracks, food caches submerged in the water, new twigs and branches on dams, new construction on lodges (fresh mud or branches; Photo 6-1), tunnels through terrestrial vegetation and exit slides from water edges (Photo 6-2).



Photo 6-1. Fresh mud is an example of recent activity and an active lodge. This lodge photographed in November 2017 was still active at Alta Vista Pond in 2019.



Photo 6-2. Other evidence of recent beaver activity: a lodge (left); tracks (middle); and a runway through adjacent vegetation (right).

In most cases, it is possible to confidently identify where a lodge, burrow, dam, or area is active based on observations that include:

- Sightings of beavers, especially if entering and exiting structures;
- New construction or repair, especially in the fall;
- Functioning and freshly-maintained dam(s)
- Fresh food caches submerged at the entrance to a lodge;
- Beaver tracks;
- Well-worn paths (tunnels and slides) through vegetation that links to the lodge's pond; and
- Evidence of extensive clippings and cuttings along those paths.

Signs of inactivity include:

- Absence of any beaver sightings in the area;
- Absence of a structurally sound lodge;
- Absence of functioning or freshly-maintained dam(s); and
- Absence of any other fresh signs (*i.e.*, that were obviously not from the survey year).

Since it is not always possible to conclude whether there has been recent activity, past reports have included a third classification (Unknown). For 2019, this uncertainty has been recognized by question marks beside a record, that is, "Active" or "Inactive." This change forced surveyors to choose which of the two classifications was most probable and was meant to allow easier interpretation of population trends.

6.2.2 Data Analysis

Three factors introduced uncertainty into the 2019 estimate of Whistler's beaver population. Firstly, and as discussed above (Section 6.2.1), it was not always possible to conclude whether a lodge was occupied.

Secondly, it is likely that not all occupied lodges were detected, though the number of undetected lodges continued to fall as the census built on past years' results. Thirdly, the population estimate relies on a multiplier of beavers per lodge that has not been verified (and would require extensive research beyond the scope of this project). It is therefore necessary to rely on data published from other areas.

The number of beavers per colony (overwintering lodge or possibly bank burrow) is based on several factors, especially habitat type and beaver density (Müller-Schwarze and Sun 2003). In 2008, data was averaged from five studies to derive an estimate of the total Whistler beaver population based on a multiplier of 5.8 beavers per lodge (Mullen 2008). This multiplier has been used each year since to derive an estimated total population. Other studies (Müller-Schwarze and Sun 2003) reported the average number of beavers per family from twelve locations that ranged from 4.1 to 8.2 and in which half were 5.1 or below and the average was 5.6 (Table 6-1). While these figures suggest the past multiplier used for Whistler studies date is reasonable, Section 6. 3 includes low, middle, and high estimates of Whistler's beaver population.

Table 6-1. Number of beavers per family in various locations (Müller-Schwarze and Sun 2003).

| Location | Avg. No. per Family | Location | Avg. No. per Family |
|--------------|---------------------|---------------|---------------------|
| Alaska | 4.1 | Alleghany | 5.4 |
| Montana | 4.1 | Ohio | 5.9 |
| Newfoundland | 4.2 | Colorado | 6.3 |
| Adirondacks | 4.3 | Isle Royale | 6.4 |
| California | 4.8 | Massachusetts | 8.1 |
| Michigan | 5.1 | Nevada | 8.2 |

6.2.3 Quality Assurance and Quality Control

Results from beaver surveys are comparable year to year, with the caveat that the survey effort and reliability has been variable to an unknown degree. It is nonetheless certain that population estimates have become more reliable since 2016 when survey efforts were expanded. Surveys each year since 2016 has built on the previous years' results and improved the accuracy of the census.

6.3 Results and Discussion

6.3.1 2019 Surveys

For the fourth consecutive year, beaver surveys detected more active colonies and came closer to a full census of beavers in Whistler. A total of 65 lodges were surveyed of which 27 were determined to have active colonies (Table 6-2). This result continues the upward trend of both number of total lodges and active colonies surveyed since surveys began in 2007.

Table 6-2. Summary table of documented lodges from 2007 through 2019 by activity status. Surveys were not conducted in 2012.

| Status | 2007 | 2008 | 2009 | 2010 | 2011 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| Lodge - Active | 9 | 27 | 16 | 16 | 17 | 10 | 10 | 7 | 13 | 13 | 16 | 27 |
| Burrow - Active | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2? | 0 |
| Lodge - Inactive | 9 | 12 | 13 | 7 | 21 | 5 | 14 | 18 | 11 | 21 | 32 | 36 |
| Summer Only | | | | | | | | | 2 | 2 | 2 | 0 |
| Unknown | 1 | 4 | 4 | 4 | 0 | 8 | 1 | 3 | 3 | 8 | 9 | NR |
| Total | 19 | 43 | 33 | 27 | 38 | 23 | 25 | 28 | 29 | 44 | 59 | 65 |

Note: NR = not recorded. Starting in 2019, surveyors classified lodges with an uncertain status as either "Active?" or "Inactive?" based on available evidence.

The number of active colonies documented in 2019 (27) is 50% more than in 2018 (18) and the highest recorded since surveys began in 2007 with one exception (Table 6-2). Mullen (2008) also recorded 27 active lodges but almost certainly over-counted. For example, she recorded 15 active lodges on the River of Golden Dreams in 2008, more than twice the number documented in any other year (Section 6.3.2).

Two areas accounted for seven out of nine active lodges added since 2018: The River of Golden Dreams (ROGD) and the Miller Creek Wetlands (Table 6-3; Photos 6-3). Since 2007 surveys began, more active colonies have been documented on the ROGD than any other area. In 2019, the importance of this habitat was again emphasized by the detection of seven active colonies which is two more than in 2018. More surprisingly, 2019 surveys revealed that the Miller Creek Wetlands provides similarly important beaver habitat. Previous surveys have been hampered by the difficult accessing this site but in 2019 exceptional effort was directed to documenting as many lodges as possible Photos 6-3. As a result, eight active lodges were detected compared to the two found in 2018. This concentration of beaver activity demonstrates the value of the Miller Creek Wetlands to those beavers and almost certainly provides a second critical source of beaver out-migration to the ROGD.

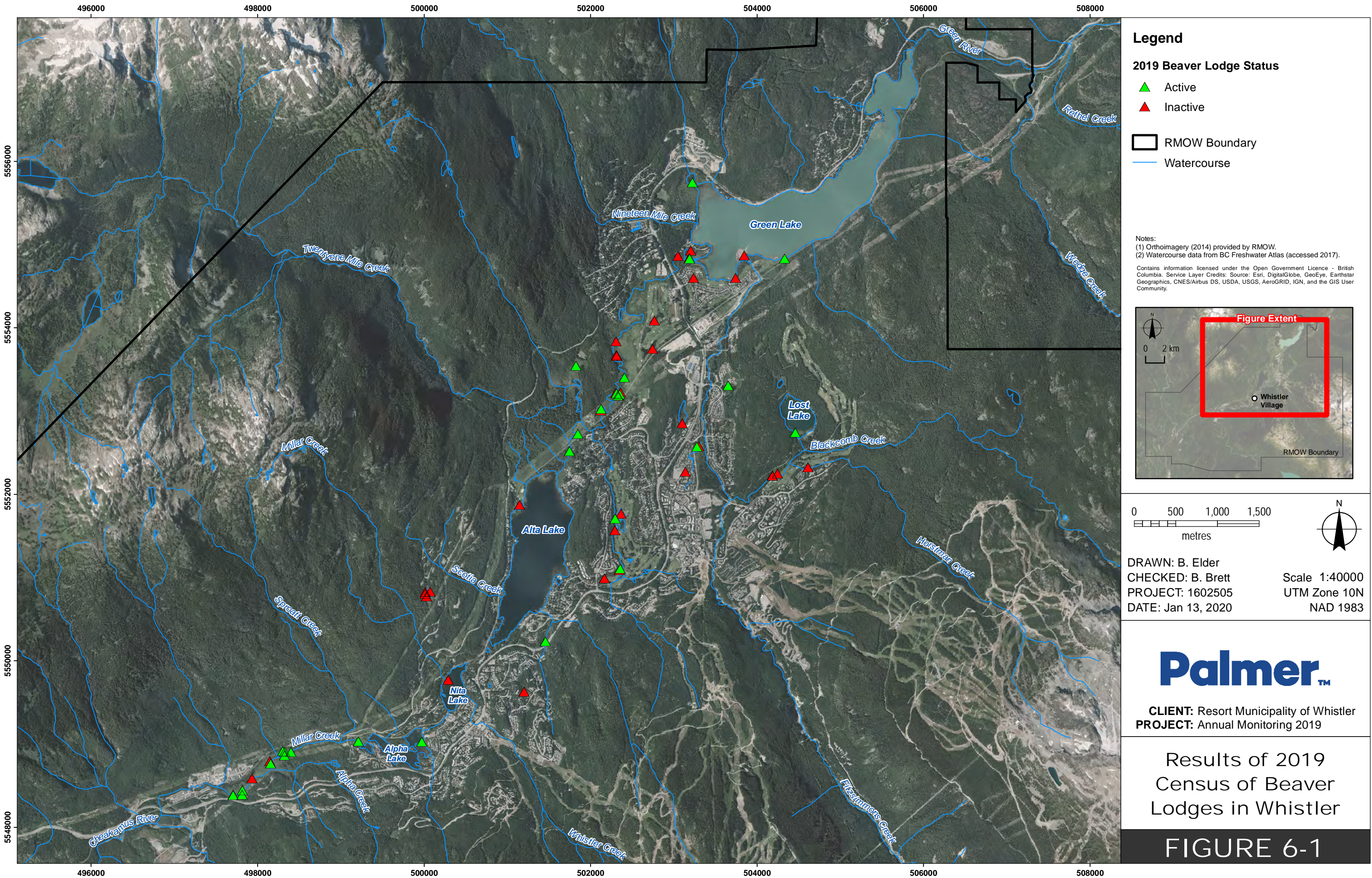
The 2018 survey recorded the first bank burrows (two) considered to house overwintering colonies. The 2019 did not find any evidence of overwintering in bank burrows. Future surveys will presumably verify if bank burrows are indeed used for overwintering in Whistler.

Table 6-3. Lodges and burrows documented in 2019.

| <u>Location</u> | <u>2017 Status</u> | <u>2018 Status</u> | <u>2019 Status</u> | <u>Eastin g</u> | <u>Northin g</u> |
|---|-------------------------------|-------------------------------|-------------------------------|----------------------------|-----------------------------|
| Alpha Lake, near dog beach | Active | Active | Active | 499970 | 5549027 |
| Alpha Lake, outlet at Millar Creek | NR | Active | Active | 499208 | 5549034 |
| Alta Vista Pond | Active | Active | Active | 501458 | 5550235 |
| Fitz Creek Pond - Blackcomb Way/Nancy Greene Dr. | NR | Active | Active | 503275 | 5552571 |
| Green Lake - ROGD, Fitz Fan, Parkhurst area] | Active? | Active? | Active? | 504330 | 5554834 |
| Lost Lake | Unknown | Active | Active | 504458 | 5552740 |
| Millar Cr. Wetlands - bet. hydro tower and Valley Tr. bench | NR | Inactive? | Active | 498301 | 5548918 |
| Millar Cr. Wetlands -FJ (Valley Trail access) | NR | NR | Active | 498321 | 5548863 |
| Millar Cr. Wetlands -FJ (Valley Trail access) | NR | NR | Active | 498324 | 5548906 |
| Millar Cr. Wetlands -FJ (Valley Trail access) | NR | NR | Active | 498398 | 5548903 |
| Millar Cr. Wetlands -FJ (water access) | NR | Active | Active | 497706 | 5548388 |
| Millar Cr. Wetlands -FJ (water access) | NR | Active | Active | 497812 | 5548393 |
| Millar Cr. Wetlands -FJ (water access) | NR | Inactive | Active | 497818 | 5548447 |
| Millar Cr. Wetlands -FJ (water access) | NR | NR | Active? | 498156 | 5548764 |
| Rainbow Wetlands, NE end near 21-Mile Creek | Active | Active | Active | 501848 | 5552727 |
| ROGD1 - Alta Lake entrance to fish weir | Active | Active | Active | 501744 | 5552517 |
| ROGD4 - RR bridge to bend nearest Valley Tr. | NR | Active | Active | 502327 | 5553188 |
| ROGD4 - RR bridge to bend nearest Valley Tr. | NR | Active? | Active | 502349 | 5553202 |
| ROGD4 - RR bridge to bend nearest Valley Tr. | NR | NR | Active | 502126 | 5553026 |
| ROGD4 - RR bridge to bend nearest Valley Tr. | NR | NR | Active | 502312 | 5553214 |
| ROGD4 - RR bridge to bend nearest Valley Tr. | NR | NR | Active | 502406 | 5553403 |
| ROGD6 - Hwy. 99 bridge to Green Lake | NR | Inactive? | Active | 503187 | 5554830 |
| Spruce Grove Park, entrance | Active | Active | Active | 503652 | 5553307 |
| Wedge Pond | Inactive | Active | Active | 503223 | 5555744 |
| Whistler GC, #15 fairway, s. of #16 outflow | Active | Active | Active? | 502356 | 5551107 |
| Whistler GC, Crabapple Cr. #10 sand trap | Active | Active | Active? | 502293 | 5551708 |
| Wildlife Refuge, middle pond | Active | Active | Active | 501825 | 5553543 |



Photos 6-3. (left) The kayak used to survey additional parts of Miller Creek Wetlands atop an active beaver dam. (right) A second active dam in the middle of the Millar Creek Wetlands.



6.3.2 Estimated Whistler Beaver Population in 2019

The 27 lodges documented in 2019 significantly raise the estimated beaver population in Whistler from past estimates, primarily due to the large concentration of beaver lodges detected for the first time in Miller Creek Wetlands. The middle estimate for the population (based on 5.8 beavers per lodge; Section 6.2.2) has risen from 104 in 2018 to 157 in 2019 (Table 6-4; Figure 6-2). Even the lowest multiplier of 4.1 beavers per lodge yields an estimated population of over 100 beavers living in Whistler in 2019.

Table 6-4. Estimated number of beavers overwintering in Whistler, 2007-2019. Surveys were not conducted in 2012.

| | 2007 | 2008 | 2009 | 2010 | 2011 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
|------------------------------|------|------|------|------|------|------|------|------|------|------|------|------|
| Active lodges/burrows | 9 | 27 | 16 | 16 | 17 | 10 | 10 | 7 | 13 | 14 | 18 | 27 |
| 4.2 beavers/site | 38 | 113 | 67 | 67 | 71 | 42 | 42 | 29 | 55 | 59 | 76 | 113 |
| 5.8 beavers/site | 52 | 157 | 93 | 93 | 99 | 58 | 58 | 41 | 75 | 81 | 104 | 157 |
| 6.4 beavers/site | 58 | 173 | 102 | 102 | 109 | 64 | 64 | 45 | 83 | 90 | 115 | 173 |

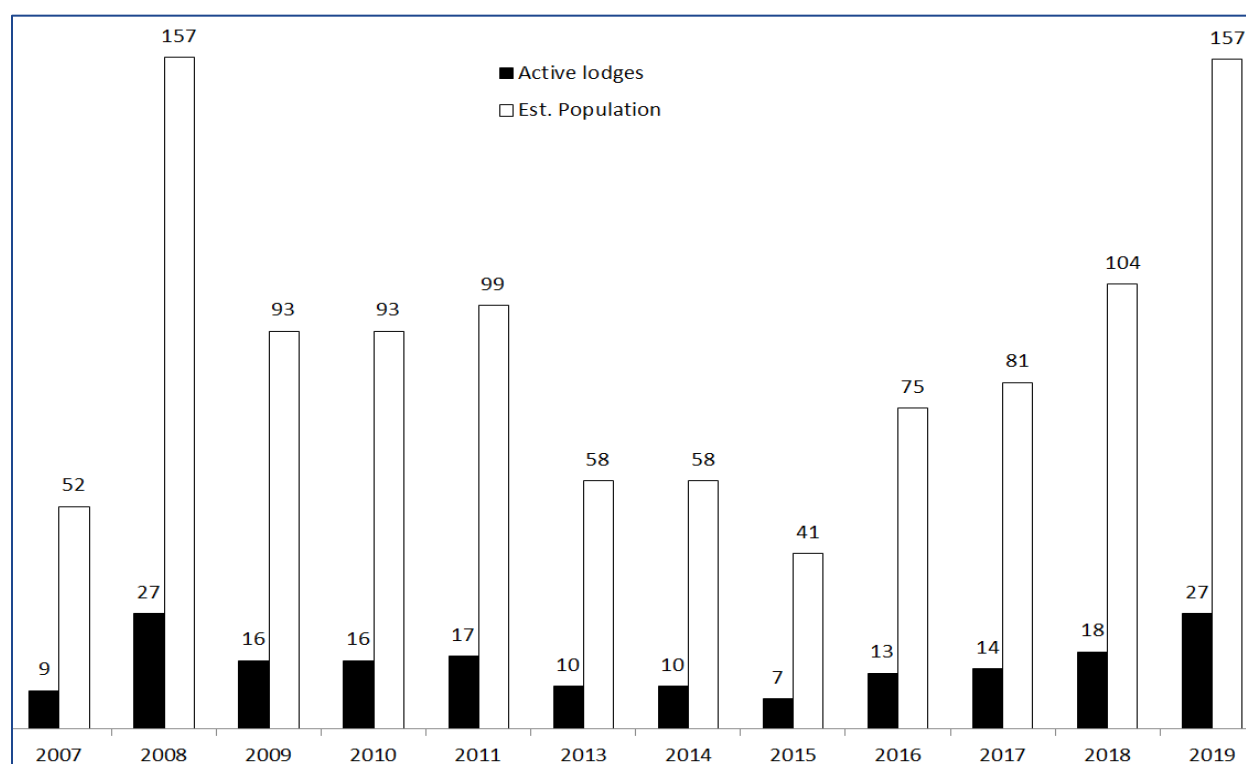


Figure 6-2. Estimated beaver population from 2007-2019 based on a multiplier of 5.8 beavers per overwintering site. Surveys were not conducted in 2012.

Results in 2019 again demonstrated the benefit of annual surveys in which past years' helped direct the next year's focus areas, especially on the River of Golden Dreams and in the Miller Creek Wetlands. Knowledge about how many beavers are in Whistler and where they are located has improved with each survey since 2016 and 2019 results again approach ever nearer to a complete beaver census. It is now unlikely that a significant beaver habitat has not been detected in the Whistler Valley.

These advances in information about beavers are already valuable in terms of knowing which areas support beavers and therefore should be considered in environmental planning. They are not, however, enough to confidently conclude any annual trends in their overall population since annual results are based on incomplete information (though least so in 2019). It is likely that monitoring the number and location of active lodges will remain more important than population estimates for environmental planning in the future since: (a) the location of lodges is by nature spatially defined; and (b) estimates of total population are not as reliable.

6.3.3 Two Major Beaver Habitats in Whistler

Previous surveys established the River of Golden Dreams (ROGD) as the single most important beaver habitat in Whistler Valley but 2019 surveys revealed the Miller Creek Wetlands have a similar importance (Table 6-2). The 2019 surveys successfully expanded the documentation of active lodges on ROGD from five to seven, a result that probably represents better documentation of existing lodges rather than an expanded population. Even more significantly, 2019 surveys in the Miller Creek Wetlands confirmed eight active lodges (versus two in 2018 and one in 2017) and therefore a similar population to that in the ROGD wetlands. Between them, these two areas account for more than half of all active lodges detected in 2019 (15 of 27).

Both of these areas are very difficult to access but for different reasons. To date, all lodges detected in the ROGD wetland complex have been at the edge of the flowing water and therefore accessible by kayak. Many of these lodges are, however, hidden even when only a few metres from the water's edge. Surveys since 2017 have therefore spent much more survey effort on land to increase detections (**Photos 6-4**).



Photos 6-4. (Left) Two surveyors search for active lodges and beaver sign on land beside the River of Golden Dreams. (Right) The lodge in the bottom left of the photo is approximately two metres from the water's edge and not visible without searches on land.

Most of the eight lodges detected in 2019 in the Miller Creek Wetlands were also beside the main stem of the creek which runs through it. Miller Creek is, however, much more difficult than the River of Golden Dreams to access by kayak because of: (a) no easily-accessible put-in locations; (b) dams and other obstructions; and (c) faster moving water. The 2019 surveys nonetheless managed to reach many previously inaccessible areas (Photos 6-3) which is the reason so many lodges were detected for the first

time. These efforts were rewarded with the discovery of six active lodges which brought the total for the area from two in 2018 to eight in 2019. Similar to new detections on the ROGD, this result almost certainly represents a more accurate tally rather than an expanded population.

6.3.4 Beaver-affected Wetlands

Beavers are well-known keystone species also referred to as “wetlands engineers” for their role in creating and maintaining wetlands (Müller-Schwarze and Sun 2003). The 2018 report described the first effort to map wetlands that have been engineered and/or directly affected by beavers within Whistler Valley (“beaver-affected wetlands”). The goal of these maps was to create a baseline calculation of beavers’ impact on wetland habitat in the form of total hectares of beaver-affected wetlands in Whistler Valley.

The 2018 report discussed the challenges in producing an accurate map of beaver-affected wetlands, especially that the exact extent of beaver impacts within these wetlands is often not obvious. The only way to truly determine that extent would be to remove the beavers and compare the wetlands before and many years after removal. With this caveat, the maps and tables presented in 2018 and updated below (Table 6-5; Table 6-6) highlight the importance of beavers within Whistler Valley.

One significant change in 2019 mapping was in the Miller Creek Wetlands where field surveys revealed almost twice as much area flooded by beaver activity than was mapped in 2018 (Table 6-5). This increase is probably not related to additional beaver activity in 2019 but rather that the mapping last year was based on aerial photos which hid flooded areas under tree cover. The current total of beaver-affected wetlands is now just over 100 ha. Alpha Lake is not included in this total even though beaver dams elevate the water level there by up to 1 m which means the lake covers approximately 7.1 ha more area than it would without damming.¹²

¹² *This estimate is based on GIS analysis performed for the 2018 report.*

Table 6-5. Location and area of beaver-affected wetlands in Whistler, 2019.

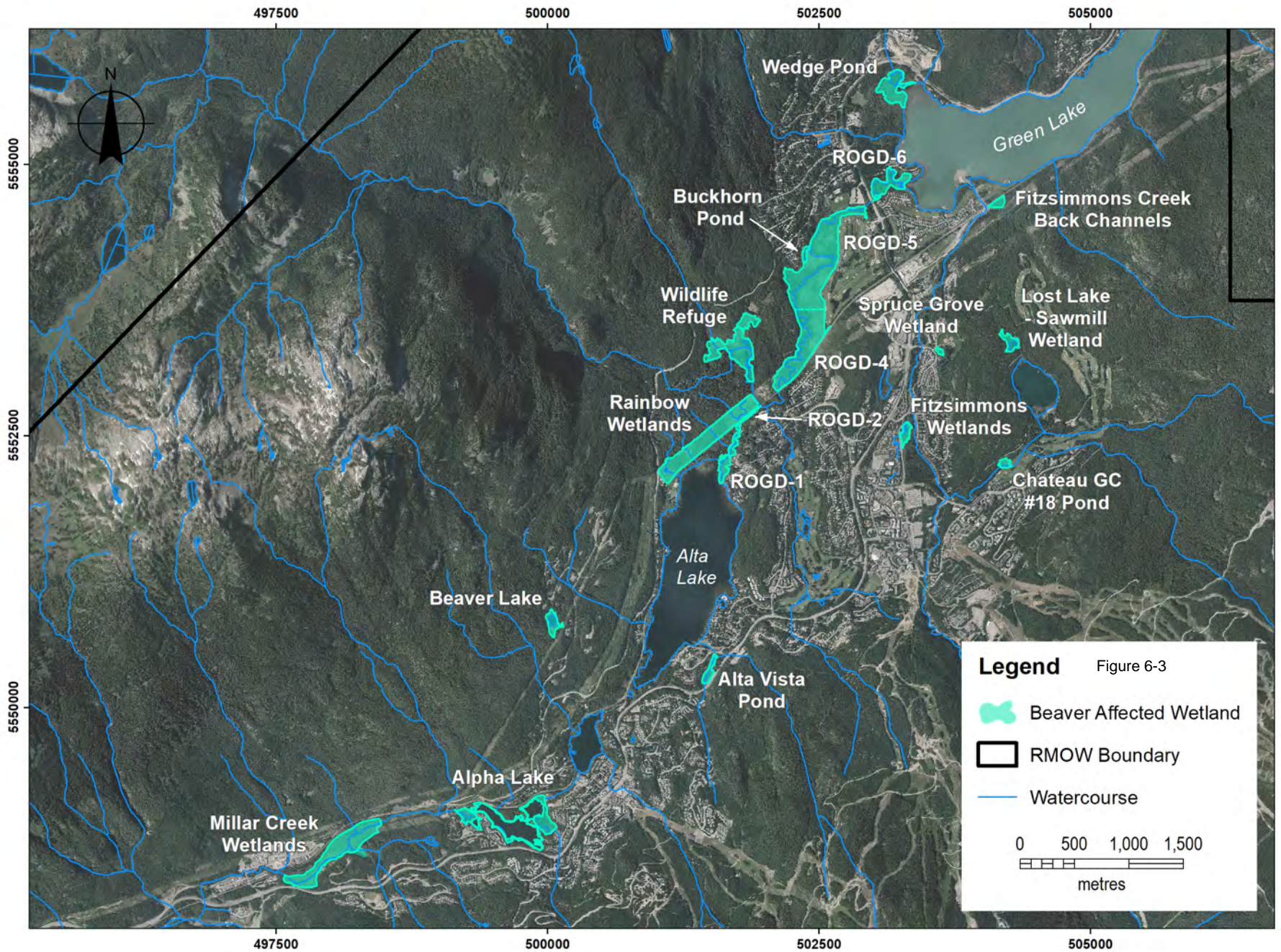
| Wetland (South to North) | 2018(ha) | 2019 (ha) | 2019 (%) |
|---------------------------------------|--------------|--------------|-------------|
| Millar Creek Wetlands | 7.6 | 13.3 | 13% |
| Beaver Lake | 1.8 | 1.8 | 2% |
| Alta Vista Pond | 1.3 | 1.3 | 1% |
| Rainbow Wetlands | 14.7 | 14.7 | 15% |
| Fitzsimmons Wetlands | 1.4 | 1.4 | 1% |
| Chateau GC #18 Pond | 0.7 | 0.7 | 1% |
| Wildlife Refuge | 10.4 | 10.4 | 10% |
| Spruce Grove Wetland | 0.3 | 0.3 | 0% |
| Lost Lake - Sawmill Wetland | 1.6 | 1.6 | 2% |
| Buckhorn Pond | 0.5 | 0.5 | 0% |
| River of Golden Dreams | 47.9 | 47.9 | 48% |
| Fitzsimmons Creek Back Channels | 0.9 | 0.9 | 1% |
| Wedge Pond | 5.5 | 5.5 | 5% |
| Total beaver-affected wetlands | 94.7 | 100.3 | 100% |
| Alpha Lake (flood effect of dam) | 7.1 | 7.1 | |
| Total beaver effect | 101.8 | 107.4 | |

The River of Golden Dreams accounted for almost half of all beaver-affected wetlands in 2019 (Table 6-5 and Table 6-6). By far, the largest part of this area is the main section between the CN Rail bridge and Highway 99 (ROGD-4 and ROGD-5; Section 6.3.4.2). Two of the next largest beaver-affected wetlands are the Rainbow Wetlands and the Wildlife Refuge (Table 6-5). Before development, these two wetlands areas would have been linked with the River of Golden Dreams in a complex spanning from Alta Lake to Green Lake on either side of the current railway line, as well as what is now the Whistler Golf Course (McBlane 2007). There would also have been much greater connectivity southward to the Miller Creek Wetlands. In total, these four wetlands account for 86% of all beaver-affected wetlands. Descriptions of all 13 wetlands as well as beaver impacts on Alpha Lake wetland are included below as Section 6.3.4.2.

Table 6-6. Areal extent of beaver-affected wetlands of different sections along the River of Golden Dreams (ROGD).

| ROGD Survey Area | Area (ha) | Area (%) |
|--------------------------------------|-------------|-------------|
| ROGD-1 (Alta Lake to fish weir) | 3.0 | 6% |
| ROGD-2 (fish weir to 21-Mile Creek) | 0.1 | 0.2% |
| ROGD-4/5 (railway bridge to Hwy. 99) | 40.4 | 84% |
| ROGD-6 (Hwy. 99 to Green Lake) | 4.4 | 9% |
| Total | 47.9 | 100% |

Note: ROGD-3 is located between the junction with 21 Mile Creek and railway bridge; this section is not included because no beaver activities have yet been detected there. See Section 6.3.4.2 for mapping of the other sections.



6.3.4.1 Historic Context

There were at least two changes that significantly impacted beavers since the railway grade was established before operations began in 1913:

1. the railbed for that railway and the ensuing increase in human presence it facilitated; and
2. the increased urban development starting in the 1960s and continuing to the present.

The railway bisected the large wetland complex mentioned above¹³ which changed the hydrology and reduced the connectivity of that area. As Whistler's population started to grow in the 1960s and 1970s, wetlands were increasingly replaced by subdivisions, golf courses and other urban developments. By 2003, at least 72% of the original area covered by wetlands was lost to development (McBlane 2007; Table 6-7; Figure 6-3).

Table 6-7. Wetland area in the RMOW by year and scope.

| Year | Wetland Scope | Area (ha) | Compared to 1946 | Source |
|------|----------------------------------|-----------|------------------|--------------------------|
| 1946 | All RMOW | 604.4 | 100% | McBlane 2007 |
| 2003 | All RMOW | 169.9 | 28% | McBlane 2007 |
| 2014 | All RMOW | 193.4 | 32% | Palmer (unpubl.) |
| 2014 | All RMOW <800 m | 169.7 | 28% | Palmer (unpubl.) |
| 2014 | <800 m, study area only | 150.7 | 25% | Palmer (unpubl.) |
| 2018 | Beaver-affected, study area only | 94.7 | 16% | Palmer and Snowline 2019 |
| 2019 | Beaver-affected, study area only | 100.3 | 17% | This report. |

Notes: The current study area is equivalent to the RMOW Development Footprint, from Function Junction to the north end of Green Lake. McBlane (2007) compared air photos taken in 1946 and 2003 within a similar but not exact scope. The 2014 data is based on the RMOW's most recent mapping of wetlands.

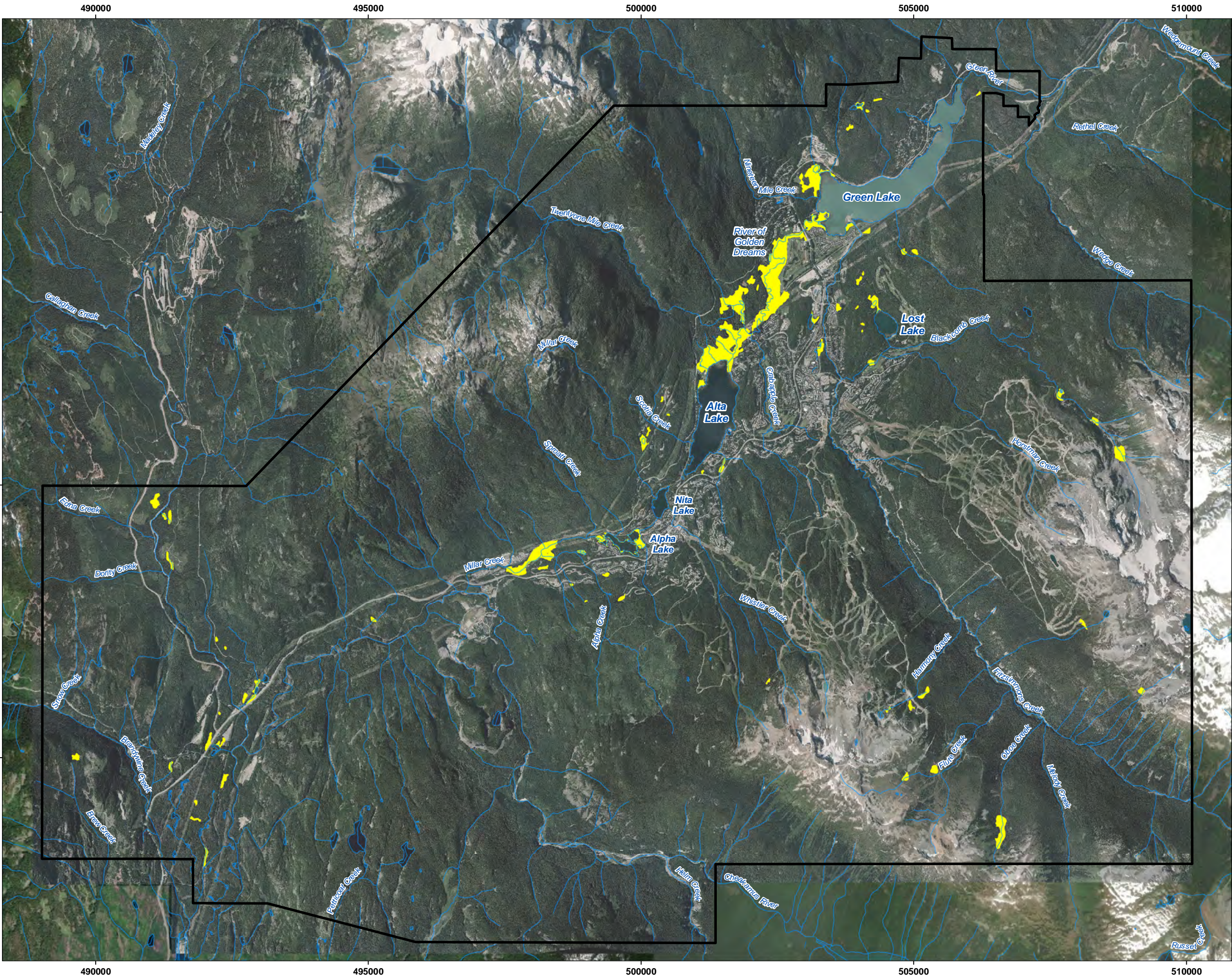
The loss of wetlands since 2003 has definitely slowed, though it is not possible with current data to provide exact figures. The RMOW's most recent mapping shows that approximately 25% of the area originally occupied by wetlands remain below 800 metres and within the Development Footprint¹⁴ remained in 2014 (Table 6-7).

A comparison of wetlands affected by beavers (Figure 6-3) and all wetlands (Figure 6-3) highlights the importance of beavers in Whistler Valley. Beavers have created or at least affected approximately two-thirds (100.3 of 150.7 ha) of all wetlands in Whistler's Development Footprint: as of 2019 (Table 6-7).¹⁵

¹³ Rainbow Wetlands, Wildlife Refuge, and River of Golden Dreams.

¹⁴ Roughly from Function Junction north to Emerald Estates and mostly below 800 metres.

¹⁵ The 2019 total is 5.6 ha larger than in 2018 due to more accurate ground-based mapping.

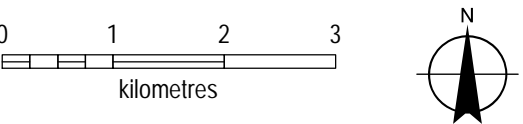


Legend

- Wetland
- RMOW Boundary
- Watercourse

Notes:
(1) Orthoimagery (2014) provided by RMOW.
(2) Wetlands displayed include features from Whistler Official Community Plan (2007/2008) and delineated "Beaver Affected" wetlands (Annual Monitoring - Palmer).
(3) Watercourse and lake data from BC Freshwater Atlas (accessed 2017).

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DRAWN: B. Elder
CHECKED: B. Brett
PROJECT: 1602505
DATE: Dec 15, 2020

Scale 1:68000
UTM Zone 10N
NAD 1983

Palmer™

CLIENT: Resort Municipality of Whistler
PROJECT: Annual Monitoring 2019

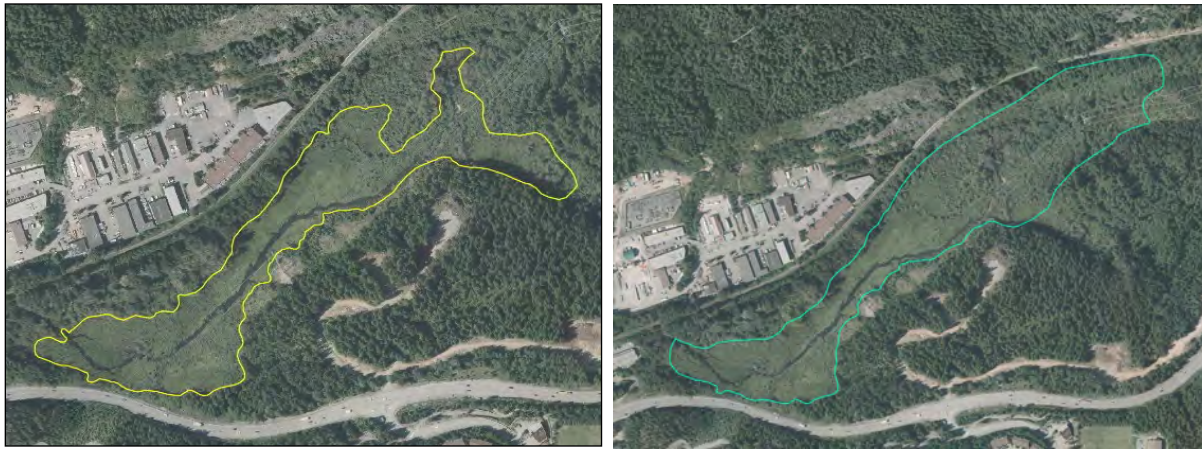
Wetlands

FIGURE 6.4

6.3.4.2 Detailed descriptions of Beaver-affected Wetlands (from South to North)

Millar Creek Wetlands

The 2019 beaver survey included two objectives for the Miller Creek Wetlands: (i) confirm the number of active lodges; and (ii) ground-truth the true extent of areas flooded by beaver activity. Both of these objectives were met. The number of active lodges detected in 2019 raised the number of confirmed lodges from two in 2018 to eight, a total that represents a beaver population as significant as that inhabiting the River of Golden Dreams (Section 6.3.3). Ground-truthing also revealed that almost twice as much area was flooded by beavers than showed in air photos: 13.3 ha versus 7.6 ha, respectively (Table 6-5; Photo 6-5).



Photos 6-5. (left) Air photo analysis in 2018 estimated that 7.6 ha were by beavers in the Miller Creek Wetlands. (right) Ground-truthing in 2019 showed 13.3 ha were actually affected by beavers. Most of this increase was due to flooded areas under tree cover that were not detectable from air photos.

Beaver Lake

In the past, Beaver Lake had four active lodges, but beaver activity has not been detected in the area since 2006. While active lodges are not present, the old lodge structures are still visible, and the related dams still impound water (Photo 6-6). There is no known impediment to recolonization by beavers in this location.



Photo 6-6. The beaver-affected wetland at Beaver Lake.**Alta Vista Pond**

The old lodge at Alta Vista Pond was recolonized by beavers in 2016 and has been active since. In 2018 and early 2019, damming of the outlet weir raised water levels enough to create one large pond (vs three different water levels behind two dams; (Photos 6-7). In September 2019, RMOW crews installed a pipe to lower the water level approximately 1.0 to 1.5 m (Section 6.3.6).

Due to the lowered water level, two previously built (and still functional) dams are visible and the pond again has three water levels: the lodge pond, a middle pond, and the lowest pond feeding into the weir. The total flooded area remained the same even with this change. (See Section 6.3.6 for a description and photos of the change in water level.)



Photos 6-7 (left) The approximate outline of the beaver-affected area of Alta Vista Pond. (right) The active lodge is shown in the left foreground of this photo from April 2019 before the pond was partially drained in September 2019 (Section 6.3.6).

Rainbow Wetlands

The Rainbow Wetlands complex is a large swath of partially inundated land with a long history of beaver activity (Photo 6-8). Until recently there was an active lodge at the west end of Rainbow Park and multiple dams upstream that impounded the water in the area nearest to the Rainbow Park lower parking lot (6-8, right, foreground). Most of the current beaver activity is now in the northern half of this area where a long-standing lodge was again confirmed active. Abundant signs suggested at least one more active lodge may not have been detected.

The RMOW's wetland layer of the Rainbow Wetlands area includes moist, forested areas especially on the upstream side of 21 Mile Creek (Figure 6-4) which means that the area of wetland calculated in this report is conservative.



Photo 6-8. (left) The approximate outline of the beaver-affected area of the Rainbow Wetlands includes the entire hydro corridor. (right) Inactive beaver dams at the southwest end of the wetlands still impound water, as seen in the foreground of this photo.

Fitzsimmons Wetland

Hillary Williamson (RMOW staff) recorded an active lodge in the Fitzsimmons Wetland in 2018. Detailed surveys in 2019 confirmed one large active lodge and one inactive lodge.

The Fitzsimmons Wetland (Photo 6-9) is the only remaining remnant of the large wetland that the Village North development replaced in the 1990s. Even when inactive (for an unknown number of years before 2017), old beaver dams maintained raised water levels and current dams have likely raised them even further. In spite of this activity, the construction of the Montebello bioswale in 2007 has apparently reduced water flow north into the Fitzsimmons Wetland and resulted in encroachment into it by cattails and other vegetation.



Photo 6-9. The Beaver-affected wetlands at Fitzsimmons Wetland: outline (left) and the lodge that has been active since at least 2018 (right).

Chateau Golf Course #18 Pond

The Chateau Golf Course #18 Pond (Photo 6-10, and Photos 6-11) is another remnant of a historically larger wetland. A very large dam (Photos 6-11) impounded water for many years (at least dating back to the first beaver surveys) and other dams have also changed water flow. The two lodges below the dam remained inactive in 2019, but recent beaver signs showed evidence of a lodge nearby (the closest detected lodges were in Lost Lake and Fitzsimmons Wetland).

The main pond drained in 2018 apparently due to lack of maintenance by beavers (Photos 6-11).¹⁶ In 2019, the pond rose partway to its height in 2017 which suggests the resumption of at least some dam maintenance by beavers. Since non-resident beavers are unlikely to maintain a dam, the 2020 survey will need to confirm whether there was an undetected lodge nearby or whether changes in water levels are unrelated to beaver activities.



Photo 6-10. Chateau Golf Course #18 Pond

¹⁶ Dan Nash, the course Superintendent, confirmed his staff did not drain the pond (pers. comm., Oct. 2018).



Photos 6-11. The Chateau Golf Course #18 pond in fall 2017 (top); fall 2018 (middle), and fall 2019 (bottom). After many years of creating this pond, the dam was not functional in late 2018. Beaver activity was again detected below the dam in 2019 (though no active lodges). This presence

apparently included enough maintenance that the dam again raised the water partway to 2017 levels.

Wildlife Refuge Wetland

The Wildlife Refuge Wetland (Photos 6-12, left) has had an active beaver population for at least 20 years¹⁷ and it is almost certain beaver activity predated the railway in 1913. Beaver dams have raised water levels and signs of beaver activities are common in the area. One long-active lodge was again confirmed active in 2019.



Photos 6-12. Beaver-affected wetlands in the Wildlife Refuge (left) and Spruce Grove Park (right).

Spruce Grove Wetland

There has been beaver activity in Spruce Grove Park for at least the last three years (Photos 6-12). The beavers have blocked the outflow weir to impound water behind it. The active lodge was located for the first time in 2018 and was very active in 2019.

Lost Lake – Sawmill Wetland

The 2019 surveys found evidence of feeding and repairs to the outflow dam in the old sawmill site north of Lost Lake (Photos 6-13, left). Since there is no evidence of new or old lodge structures in this area, the beavers associated with that activity are likely from the active lodge in Lost Lake itself.

¹⁷ A photo of an old beaver dam at the south end of the wetland dates back to 2000 (B. Brett photo -- <https://www.whistler.ca/services/environmental-stewardship/ecosystem-monitoring>).



Photos 6-13. (left) Beaver-affected wetlands at north of Lost Lake at the old sawmill site; and (right) at Buckhorn Pond.

Buckhorn Pond

Buckhorn Pond is connected to the River of Golden Dreams Wetlands and is the only large pond within the complex (Photos 6-13, right). Water levels are maintained by a dam that, although functional, has not been repaired in at least two years. A resident saw beavers eating aquatic vegetation in this pond in 2016 (Palmer and Snowline 2017) and beavers from lodges on the River of Golden Dreams likely continue to access this source of food.

River of Golden Dreams Wetlands

The River of Golden Dreams wetland complex contains almost one-half of the area of beaver-affected wetlands in Whistler (Table 6-5; Table 6-6). This is a complex system to survey and to describe which is why reports since 2016 have segmented the river into six sections:

- ROGD-1 (Alta Lake entrance to fish weir);
- ROGD-2 (fish weir to junction with 21 Mile Creek);
- ROGD-3 (21 Mile Creek to railway bridge);
- ROGD-4 (railway bridge to closest approach to Valley Trail – about midway through this section);
- ROGD-5 (closest approach to Valley Trail to Highway 99 bridge); and
- ROGD-6. (Highway 99 bridge to Green Lake).

All except for the ROGD-3 segment have been affected by beaver activities.

ROGD-1 (Alta Lake entrance to fish weir)

The first segment of the River of Golden Dreams (Photos 6-14; left) includes one large, long-standing lodge upstream of the Valley Trail bridge. While there are not many obvious alterations on land from this lodge, there is a small dam (frequently breached by boaters) that slightly raises the water level. In 2019, the lodge had active slides and well-used pathways leading to and from its entrance.

ROGD-2 (fish weir to junction with 21 Mile Creek)

The second segment of the River of Golden Dreams (Photos 6-14, right) is a narrow, constructed channel that is defined by the CN railbed adjacent to it. Beavers have long-used this area, most notably in recent years, with bank burrows that are presumably unoccupied in winter. Active lodges have been previously observed in this area (e.g., Tayless 2010), but none in recent years.



Photos 6-14. The southern most segment of the River of Golden Dreams (ROGD) wetland: (left) ROGD-1; (right) ROGD-2. ROGD-2 occupies the narrow channel that drains from the fish weir to 21-Mile Creek.

ROGD-4 and ROGD-5 (railway bridge to Highway 99 bridge)

This segment is by far the largest wetland through which the River of Golden Dreams flows (Photo 6-15). Yearly evidence of beaver activity is apparent throughout the area and includes: lodges, bank burrows, food caches, gnawed trees and branches, tracks, scent mounds, slides, tunnels through vegetation, as well as some direct sightings of beavers.

Five active lodges were detected in this part of the River of Golden Dreams in 2019, the highest total to date. All appeared to be older structures which were not previously detected due to how well they were hidden under dense vegetation (mainly hardhack) and back from the water's edge.



Photo 6-15. The largest contiguous wetland that the River of Golden Dreams passes between the railway bridge to the south and bridge over Highway 99 to the north. This area is coded as ROGD-4 (south end of the polygon) and ROGD-5 (north end of the polygon). The Wildlife Refuge Wetland is shown to the southwest (bottom left).

ROGD-6. (Highway 99 bridge to Green Lake)

Beavers are also active each year in the downstream segment of the River of Golden Dreams (Photo 6-16). Small dams are usually maintained, though they don't tend to impound much water. Although there were abundant signs of activity in the river and on the adjacent shore in 2019, only one active lodge could be detected.



Photo 6-16. The northmost section of the River of Golden Dreams wetland, ROGD-6, is located between the Highway 99 bridge and Green Lake.

Fitzsimmons Creek Back Channels

The Fitzsimmons Creek back channels (Photo 6-17) are on the uphill (northeast) side of the railway tracks, east of and adjacent to the Fitzsimmons Creek main channel at Nicklaus North Golf Course. Extensive beaver activity was first documented in this location in 2016 but no active structures were detected until surveys conducted in 2018. One active lodge and six burrows were found during surveys of the Fitzsimmons Creek back channels in 2018 and at least one colony was active. Surveys in 2019 were unable to detect any active colonies in spite of evidence such as feeding and caches. Burrow entrances found in 2018 appeared unused and the lodge had no signs of recent activity. It is unclear whether recent trail building in the area (as part of the Lost Lake Zappa Trails) had any impact (also see Section 6.3.6).



Photo 6-17. The approximate area influenced by beavers at the Fitzsimmons Creek back channels.

Wedge Pond

An active lodge at Wedge Pond (Photo 6-18) first detected in 2018 was again confirmed as active in 2019. Although active lodges have not been found in many survey years, there has always been extensive evidence of their presence in the area including numerous channels, dams, structures and other signs.

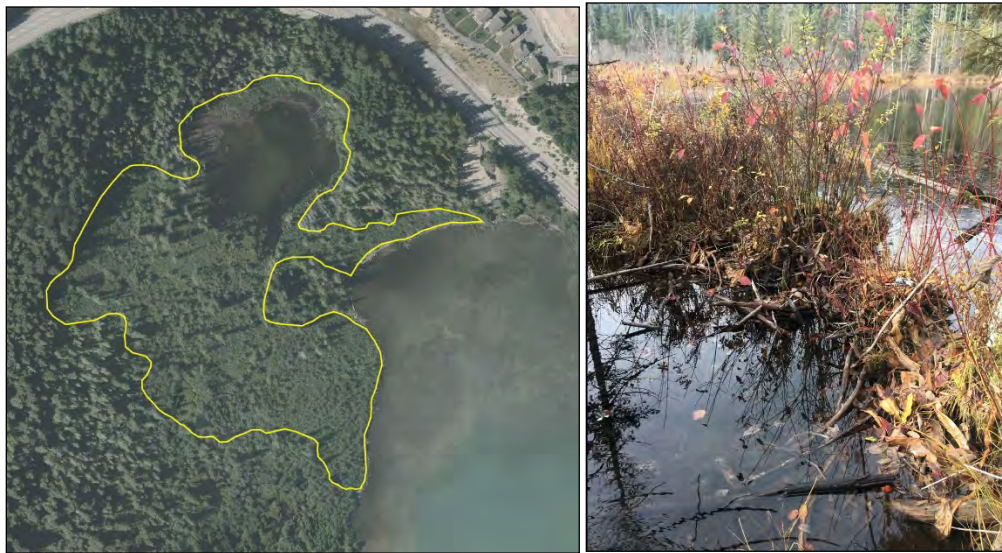


Photo 6-18. The beaver-affected wetlands at Wedge Pond.

Alpha Lake (non-wetland)

Alpha Lake is the only non-wetland area included in the discussion of beaver-affected habitats (Photo 6-19). The beaver dam at the outlet of the lake has been functional for at least 30 years (and likely far longer) and maintains water levels approximately 1 m higher than if the dam was not present. Given the large amount of area that is covered by shallow water, it is apparent that the beaver dam greatly increases the lake's surface area. At least some of these shallow areas, especially on the west and east edges, have wetland characteristics. Two active lodges were detected in 2019.



Photo 6-19. *The outlet dam created by beavers on Alpha Lake (top left). The outer yellow line shows the approximate extent of the beavers' influence (e.g., wetland vegetation). The inner line attempts to delineate how much the open lake surface would diminish without the dam. The estimated impact of beavers is the area within the two lines.*

6.3.5 Trapping Records as a Proxy for Historic Beaver Population in Whistler

A baseline population pre-development is an important component of an effective monitoring program but not always possible to determine. For beavers, it is clear they have always been an important part of Whistler's valley bottom ecosystems (e.g., Racey and McTaggart-Cowan 1935). It is almost certain beavers were much more numerous before the development of Whistler Valley based on two pieces of evidence. Firstly, the area of wetlands has decreased by approximately 75% (McBlane 2007; Section 6.3.4) which suggests the historic population of beavers was possibly four times larger than it is today. The second is a recollection of Don Maclaurin, a well-known figure who moved to Whistler in 1964 and remembered seeing six lodges on Alpha Lake at that time.¹⁸ Surveys since 2007 have meanwhile documented between one and two active lodges on Alpha Lake, a decrease consistent with the 75% loss of wetlands in Whistler as a whole.

¹⁸ Personal communication with Bob Brett, approximately 20 years ago.

One of the 2019 goals was to examine historic and recent trapping records as a proxy of the past beaver population. An extensive online search did not reveal any trapping records, even on the BC Government websites. At the time of writing, replies to emails to many government and independent biologists have not returned any leads. This effort will be pursued with the goal of including results from it with next year's report.

6.3.6 Conflict Areas in 2018

Beavers have a long history of conflict with humans, especially when urban development occurs in valley bottoms, as is the case in Whistler. Beaver conflicts are seldom made public, thus limiting the ability to report on them. All available information about 2019 conflicts is discussed below.

Millar Creek Wetlands: Line maintenance in 2018 by Fortis Gas included the new Valley Trail alignment between Alta Lake Road and Function Junction that passed through beaver-dammed parts of this wetland. Fortis BC applied for a trapping permit but was able to achieve their goals instead by breaching some of the dams.¹⁹ This situation exemplified an opportunity to design new developments to avoid beaver conflicts, especially when the developer is the RMOW. As of fall 2018, the RMOW planned to build the trail to accommodate flooding,²⁰ and no conflicts with beavers were recorded in 2019.

Alta Vista Pond: The 2018 report noted that the RMOW Roads Department was troubled by beavers blocking the outflow weir in this pond. It suggested that the RMOW investigate ways to accommodate a higher water level in the pond by lining the road subsurface and/or other measures that would protect the integrity of the road while also protecting beaver habitat.

In September 2019, the Roads Department built a gravel road to access the pond and build a more beaver-proof drain. As a result, the pond level was lowered by 1-1.5 metres (Photo 6-20). It appears other options were not considered.

¹⁹ Hillary Williamson (RMOW) email to B. Brett, September 2018.

²⁰ Heather Beresford email to B. Brett, September 2018.



Photo 6-20. Alta Vista Pond in April 2019 before draining (top) and in September 2019 after draining (bottom).

Rainbow Wetlands: CN Rail continues to breach dams each year along the section of tracks in the Rainbow Wetlands area. Dam breaches over the past four years have not prevented beavers from inhabiting the area, and the population was again robust enough in 2019 to rebuild the dams throughout the summer and into the fall.

River of Golden Dreams. The main concern for beaver conservation within this important habitat is human use of the river. On most sunny days in the summer, large numbers of people boat on the river by canoe and kayak (primarily customers of outdoor recreation companies) or by inflatable boats. Dams are routinely breached by the passage of these vessels (possibly unintentional). Without this human activity, dams would impound more water, beavers could likely remain active more hours in a day (as they would not have to avoid humans), colonies would likely be more plentiful and more area in the wetland complex would likely be inundated.

Whistler Golf Course: The golf course is built on a previous wetland which has a creek passing through the course to the west (the creek is named Archibald Creek above the course and Crabapple Creek inside the course). Since the level of the creek is not far below the level of the course, damming by beavers can flood the course and cause damage. The Whistler Golf Course has a long but unquantified history of trapping beavers,²¹ but recently has made some efforts to co-exist with them. The lodge at the #10 sand trap was trapped out and re-colonized at least once during the time beaver surveys have been conducted (2007-2018). This lodge is in a relatively benign location since it is far below the golf course and therefore some damming and beaver activity can be tolerated. The other frequent site for beavers to recolonize after being trapped out is adjacent to the #15 fairway. The elevation of the creek at that point is very close to the elevation of the golf course which means that it is less tolerable for golf course operations. Golf course staff breached dams and hired a trapper in the fall of 2018 who was unsuccessful in eliminating the beavers in the two lodges. In 2019 the #10 lodge appeared inactive but it is unknown if that was due to trapping.

Spruce Grove Park: In 2018, RMOW road crews responded to increased beaver activity blocking the outflow weir by removing material. Though not confirmed, it appears RMOW crews again removed material blocking the weir in 2019 since the water level has not increased in spite of extensive beaver activity in the pond. This site may be another opportunity for the RMOW to set a standard that allows co-existence with beavers.

Chateau Golf Course #2 and #18 Ponds: Beaver activity appears to have ceased in this area since sometime in 2017. The main dam in #18 failed in 2018 but appears to have been repaired by beavers enough to again raise water levels at least partly back to past levels. The course superintendent, Dan Nash, confirms there was no trapping again in 2019 on the course and does not have any information about the absence of active lodges.²²

Nicklaus North Golf Course: Whistler's third golf course has a long history with beavers, understandably given its location used to be a major portion of the historic Alta Lake to Green Lake wetland complex. The course is very close to the River of Golden Dreams (Photo 6-15). Beavers have used golf course ponds to forage and, less frequently, to build lodges (most recently on #10 pond, now inactive for three years). No known trapping or other control efforts took place in 2019.

²¹ Past trapping records should be maintained by the BC Government but have not yet been located (Section 6.3.5).

²² Several conversations with Bob Brett throughout summer and fall 2019.

Fitzsimmons Back Channels:

At least one beaver colony was active in this area in 2018 but no active lodges or overwintering burrows were found in 2019. It is possible a new single-track bike trail (Muffin Man; Photo 6-21) built through this area in 2018 and finished in 2019 may have negatively affected beavers.



Photo 6-21. The new Muffin Man single-track bike trail built through the Fitzsimmons Back Channels.

7. Additional Species

Lead Biologist and Author: Bob Brett

7.1 Northern Goshawks

The population of BC's Northern Goshawks (*Accipiter gentilis*) has declined precipitously in recent years, at least partly due to the loss of old forest habitat (BC MFLNRO 2018). Two subspecies occur in British Columbia. Queen Charlotte Goshawk occurs in the Whistler area (*A. gentilis laingi*; MFLNRO and Madrone 2014, 2015; CDC 2020).²³ The other subspecies, *A. gentilis atricapillus*, occurs throughout the rest of BC and other parts of North America. Both subspecies of the Northern Goshawk are listed as species at risk. The *A. laingi* subspecies is Red-listed in BC (CDC 2020) and Threatened under the Canadian Species At Risk Act (Government of Canada 2020). The other subspecies of Northern Goshawk that occurs in BC, ssp. *atricapillus*, is Blue-listed in BC but considered Not At Risk by the Canadian Government (CDC 2020; Government of Canada 2020).

Northern Goshawks were selected by the Working Group (Brett 2018) for consideration as indicators within this RMOW Ecosystems Monitoring Program (EMP). The 2018 EMP report included an initial effort towards compiling and updating current knowledge about goshawks within the RMOW to help assess how or whether they could be cost-effective indicators within the program. This section updates that information and also considers ways to include monitoring of Northern Goshawks in future years.

Available data includes 57 records of Northern Goshawk observed in Whistler since 2001 (Figure 7-1; Appendix F), including 11 new records from 2019. The BC Government has conducted or commissioned sporadic surveys for goshawks in Whistler since 2011 when they reported an active nest uphill and west of the current Whistler RV Park.²⁴ Surveys in advance of construction of an Independent Power Project (IPP) on Wedge Creek found active nests near Comfortably Numb Trail in 2014 and 2015 (MFLNRO and Madrone). Another active nest was recorded by this program in 2016 and 2017.

Goshawk surveys are time-consuming and require experienced biologists, especially when searching for new nests. While less-experienced surveyors can be used to monitor known nest sites, searching for new nests requires the surveyors to be able to distinguish calls and observations of goshawks from other raptors and also from mimics including Gray Jays (*Perisoreus canadensis*). In addition, knowledge of goshawk behaviour and the ability to detect goshawk feeding, whitewash, and other signs is important for maximizing the accuracy of the survey. These requirements mean that a comprehensive search for new nests in Whistler is outside the scope of the RMOW EMP.

An independent study in 2019 by the Association of Whistler-Area Residents for the Environment (Brett 2020) undertook the first attempt to do a comprehensive survey for goshawks in Whistler. It was able to confirm juvenile calls near the 2014 Comfortably Numb nest that showed a breeding pair was still active in 2019 (Photo 7-1). Several anecdotal reports of goshawk sightings in that area Figure 7-1; Appendix G) may

²³ See Brett (2020) for an update and discussion of the taxonomic and conservation status of Northern Goshawk.

²⁴ BC Conservation Data Centre (CDC) Species Occurrence Report Shape ID 106601. This area was recorded as Brew Creek.

have been associated with that nest. Another concentration of goshawk sightings occurred in the Kadenwood area, near enough to the 2016-17 nest to suggest a second breeding pair still resides near there. Even if there was a breeding pair there, it is unknown if it successfully bred in 2019.

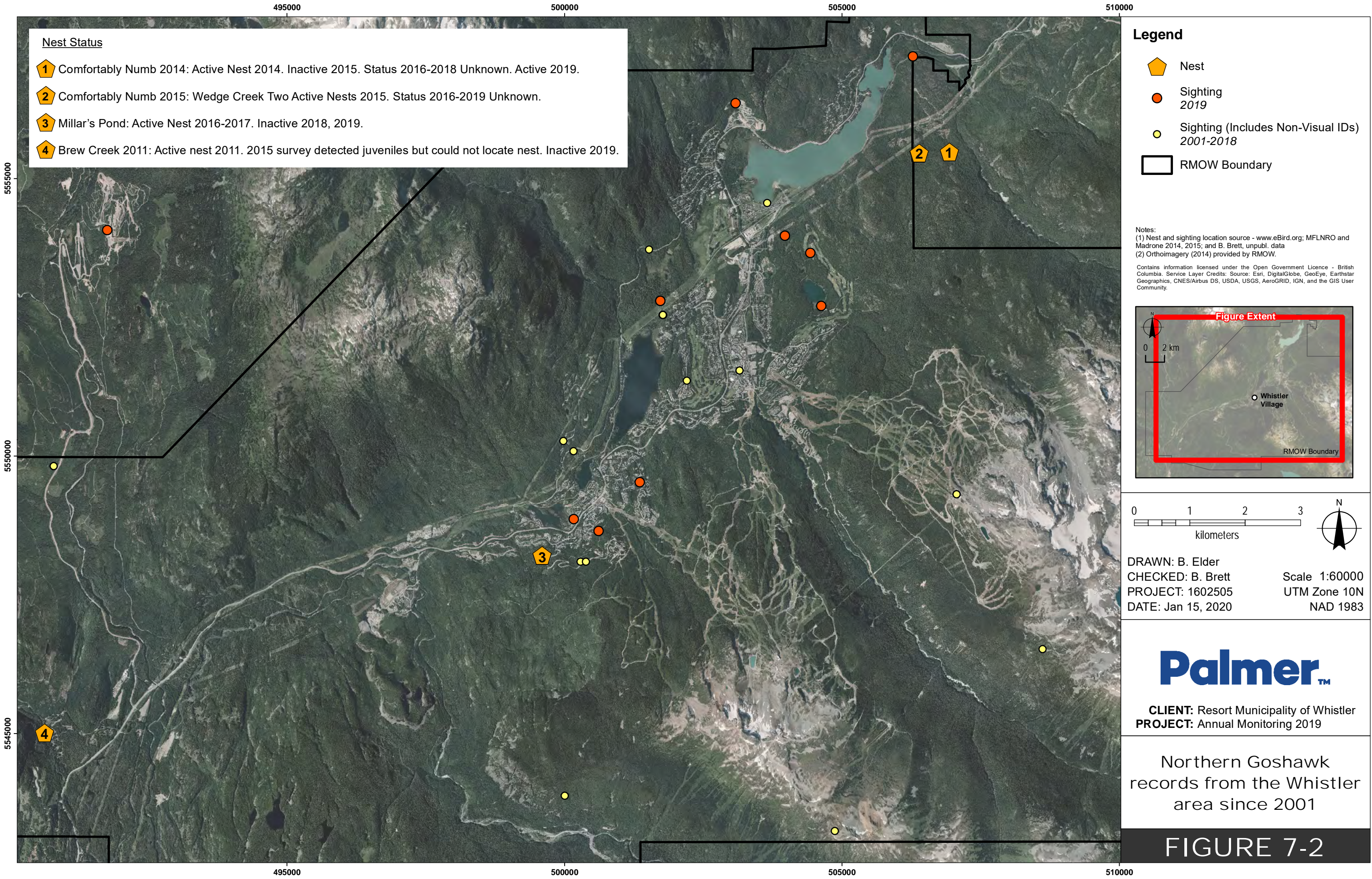


Photo 7-1. Northern Goshawk habitat near Comfortably Numb Trail, west of Wedge Creek: (left) large, wide-spaced trees, open flyways and access to the forest floor typical of excellent goshawk habitat; (right) juveniles were detected near this nest in 2019 which was last active in 2014 (Brett 2020).

The last point is important since a sustainable local population of goshawks requires breeding pairs as well as suitable habitat, for both nesting and foraging. Based on research, the ideal goshawk habitat in Whistler is in old forests that have characteristics not found in other habitats including: (a) large, tall trees with abundant potential nest platforms; (b) wide tree spacing and open flyways to facilitate hunting under the canopy; and (c) relatively unobstructed access to potential prey on the forest floor (Brett 2020; Figure 7-1). Brett (2020) mapped how logging since 2011 has reduced known and potentially available goshawk habitat near Comfortably Numb Trail and the Whistler RV Park (“Brew Creek”). It also raised questions about the future viability of the Miller’s Pond site that is currently very close to zoning for a large housing development proposed in the RMOW draft OCP.²⁵

Monitoring Northern Goshawks in Whistler remains a priority, even if the cost and effort remain a challenge. One benefit of outside research (e.g., Brett 2020 and also cooperation with BC Government biologists) is that, once located, less effort is required to monitor known nests.

²⁵ <https://www.whistler.ca/ocp>



7.2 Western Toads

The RMOW Environmental Stewardship department has monitored amphibian populations at the south end of Whistler over the past decade, especially near the Cheakamus Crossing neighbourhood. In that time, the only annual breeding site confirmed for Western Toads has been at Lost Lake. Given the number of anecdotal reports of juvenile and adult toads at the south end of Whistler, it seems reasonable to assume there would be one or more annually-used breeding sites in that part of the RMOW as well. One site was chosen by the RMOW to survey in 2018 as part of the effort to test this hypothesis: a pond on the northwest corner of the entrance from Highway 99 to the Callaghan Forest Service Road (UTM 493120E 5546435N, elevation 512m) No toads or other amphibians were detected in two shore surveys that year. The work plan for 2019 included re-surveys at that pond which occurred on May 22 and 23 (Photo 7-2) again with no detections.



Photo 7-2. The pond surveyed for Western Toads and other amphibians on May 22 and 23, 2019. This pond is on the northwest corner of the entrance from Highway 99 to the Callaghan Forest Service Road and is labelled as “1” in Figure 7-2.

There are many small wetlands near the Callaghan FSR pond, including two directly southwest and a third on the east side of the highway, south of the new Callaghan road (Figure 7-2). These ponds were assessed

on May 22, 2019 for possible inclusion in future surveys. The artificial wetlands constructed in the Brandywine snowmobile parking lot (in 2009?) were also assessed.



Figure 7-2. Wetlands surveyed May 22-23, 2019 near the Callaghan River.

Ponds 2 and 3 (Figure 7-2)

These ponds were partially filled in during the widening of Highway 99 prior to the 2010 Olympics. The road bed falls very steeply into them (Photo 7-3) and access to them would be difficult, either by kayak or by foot. It nonetheless may be worthwhile to attempt a kayak-based survey in 2020 survey.

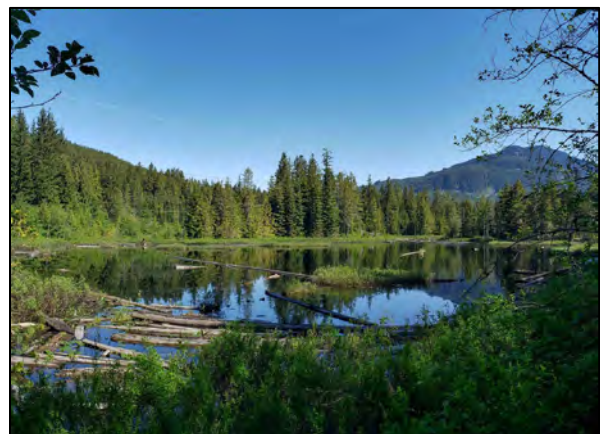


Photo 7-3. The steep road bed beside wetland #2 (Figure 7-2) complicates access for amphibian surveys.

Photo 7-4. Kayak access for amphibian surveys is relatively easy at wetland #4 (Figure 7-2).

Pond 4 (Figure 7-2)

Kayak access to this larger pond is also somewhat difficult but definitely possible (Photo 7-4). Foot access around the perimeter of the pond is not advisable due to dense vegetation and rough terrain.

Brandywine Artificial Wetlands (Figure 7-2)

The Cheakamus Community Forest created approximately four shallow depressions as replacement wetlands in approximately 2009 (Photo 7-5). All but one are now covered by an extension of the parking lot. The remainder is covered by vegetation and is similarly non-wetland.



Photo 7-5. Small depressions were created in approximately 2009 in the Brandywine snowmobile parking lot (labelled “5” in Figure 7-2). Most of that area is now covered by an extension of the parking lot (middle left). The southmost depression is now completely grown over with vegetation.

Additional surveys would be worthwhile in 2020 and beyond to determine the location of toad breeding in the southern end of the RMOW.

7.3 Western Screech-Owls

A February 2018 video of a Western Screech-Owl inside a house on Alta Lake Road was circulated to BC Government biologists, including Kym Welstead. In May 2019, she organized the installation of two Autonomous Recording Units (ARUs) in the area. No detections of owls were reported and a report will be published by the end of March.²⁶ Kym Welstead expects to conduct additional surveys in 2020 and indicated willingness for cooperation.

7.4 Black Cottonwoods

Black cottonwoods (*Populus trichocarpa*), especially when large and old, provide important habitat for a wide range of organisms in Whistler. Mapping for a conservation ranking of species and habitats (Brett 2018) showed cottonwood forests are relatively uncommon in Whistler and generally concentrated in areas associated with valley bottom wetlands and riparian areas between Alta and Green Lakes. The Working Group assembled for this report prioritized black cottonwood as a priority species which should be considered for inclusion in future years of the RMOW Ecosystems Monitoring Program. The first step towards that goal was the initial analysis of the extent and distribution of cottonwood forests presented below.

The cottonwood map layer in Brett (2018) included all ecosystem polygons that contained cottonwoods (regardless of percent cover) and showed the age of trees within (<100 years, 100-250 years and >250 years). Conclusions from that report included:

1. The largest contiguous area mapped as containing cottonwoods (between Spruce Grove and Nicklaus North) appears to have been mostly developed since that mapping. Updated mapping from air photos and/or field-truthing may be required to determine the current extent of cottonwoods in that area.
2. The main areas with old (>250-years) cottonwoods are the Edgewater forest, Rebagliati Park north on the west edge of Fitzsimmons Creek to the wetlands south of Nancy Greene Drive (Photo 7-6. Large cottonwoods near the River Runs Through It (left) and Rebagliati Park (right). The tree on the right is approximately 300 years old (Brett and Ruddy, 2019). The tree on the left has not been cored but based on the age of other trees in the area, it is likely in the range of 100 years-old. Older trees are larger, have more complex branching and provide more habitat for more organisms, especially those that benefit from tree cavities.) and in riparian areas of the Cheakamus River upstream near the park entrance.
3. Significant components of younger cottonwoods occur in the River Runs Through It area (Photo 7-6 and Photo 7-7), edges of both the River of Golden Dreams wetlands (Photo 7-8.), Millar Creek Wetlands and south of the development footprint in the riparian edges of Cheakamus River near the Sugar Cubes (across from the entrance to the Callaghan Forestry Service Road).

²⁶ Kym Welstead, January 6, 2019 email to Bob Brett.

The work plan for 2019 included the investigation of LIDAR as a way to more accurately map cottonwoods, even down to significant individual trees, but this was unsuccessful. Consultation with RMOW GIS Analyst Greg Thistle²⁷ revealed the RMOW's new LIDAR layer is not suitable for that use. As a result, field surveys combined with detailed orthophoto analysis would be required to develop better mapping of Whistler's cottonwoods. Perhaps the most promising option is the use of air photos that have been taken when cottonwoods turn yellow in fall (Photo 7-8). Unfortunately, air photos are normally taken earlier in the fall²⁸ which means another source for those photos would be needed.

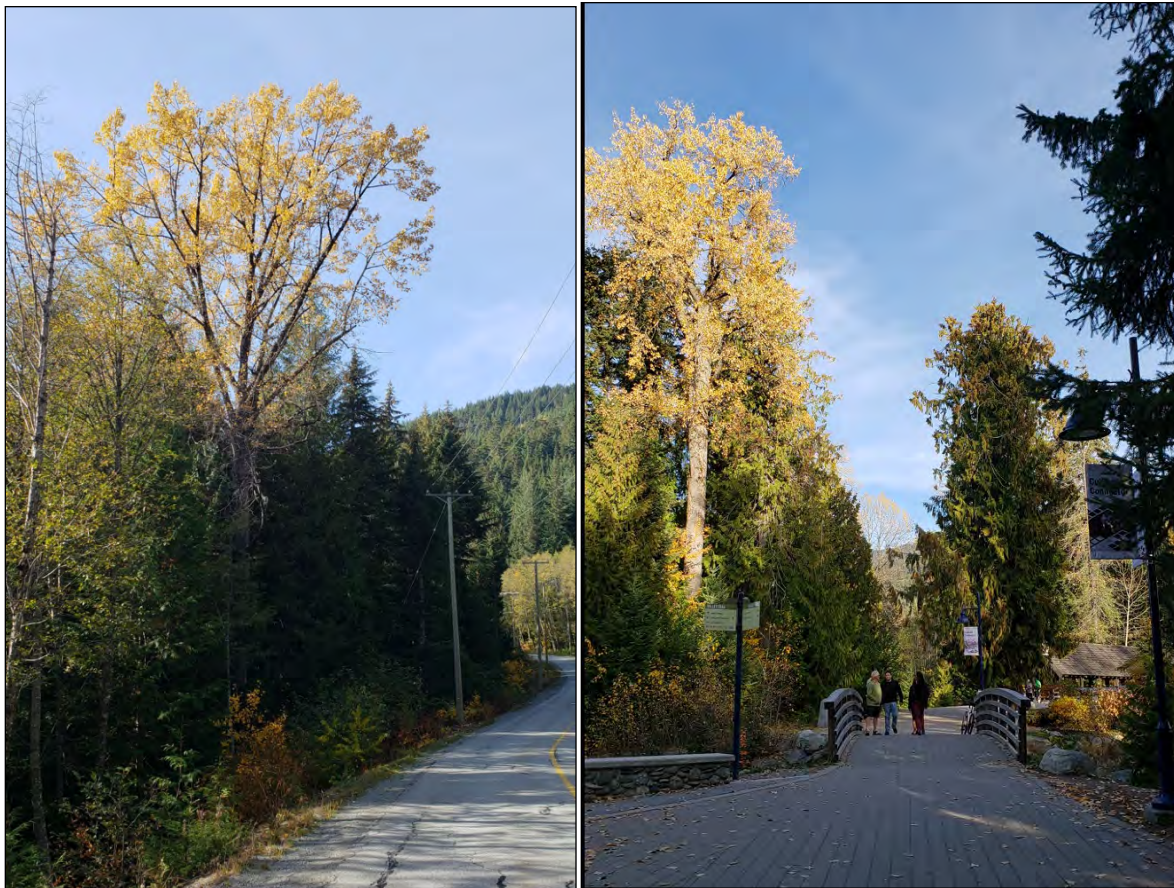


Photo 7-6. Large cottonwoods near the River Runs Through It (left) and Rebagliati Park (right). The tree on the right is approximately 300 years old (Brett and Ruddy, 2019). The tree on the left has not been cored but based on the age of other trees in the area, it is likely in the range of 100 years-old. Older trees are larger, have more complex branching and provide more habitat for more organisms, especially those that benefit from tree cavities.

²⁷ By email to Bob Brett, December 3, 2019.

²⁸ Ibid.

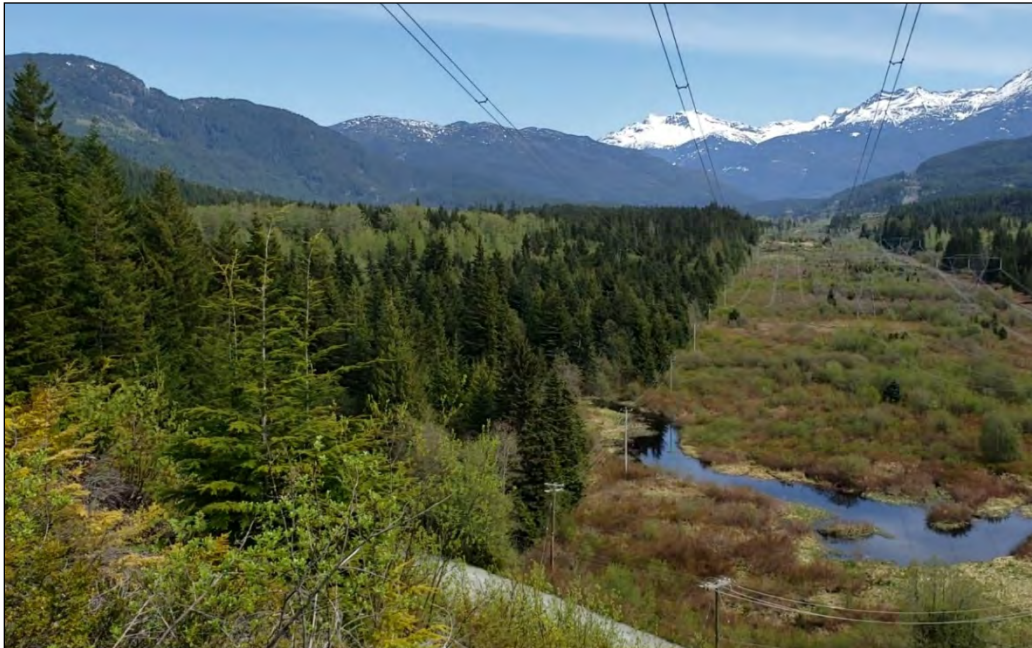


Photo 7-7. A cottonwood-dominated forest in the River Runs Through It area adjacent to 21 Mile Creek.

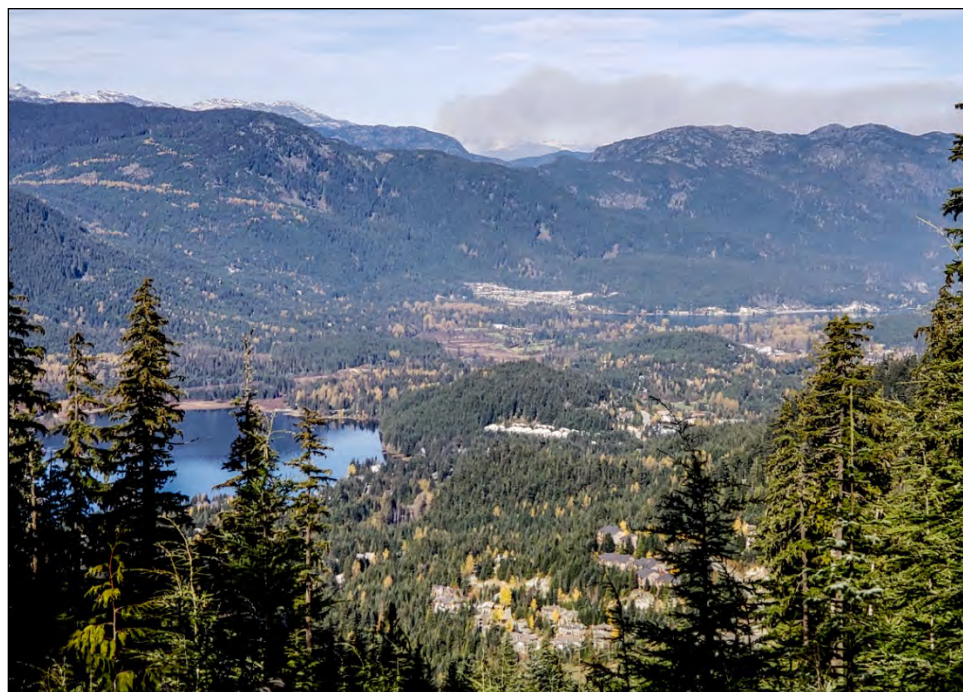


Photo 7-8. Cottonwoods are most visible after turning yellow in fall. This photo shows the view northward across Whistler Valley and highlights cottonwoods interspersed in developed areas, especially at low elevations of the wetland corridor in the mid background.

8. Climate Indicators

Lead Biologist and Author: Bob Brett

The timing and duration of ice on Alta Lake has been used as a climate indicator since the inception of the Ecosystems Monitoring Program. Cascade (2013) compiled data from two reporting periods: 1942 to 1975, and from spring 2002 to present. No data is known to have been recorded between those two periods. The current dataset is derived from the Alta Lake Ice Break Up Raffle, a fundraiser for The Point Artist-Run Centre.²⁹ The purpose of presenting and analyzing this data is to document how the timing and duration of ice on Alta Lake has changed over time to predict how it may change in the future.

Fifty years of data from 1942 through spring 2019 are presented as Appendix H. There has been a significant reduction in the duration of ice on Alta Lake between the early records and records since 2002 (Figure 8-1; Table 8-1). Nine of 10 of the winters with the longest duration of ice on Alta Lake were from the earlier dataset, and six of 10 of the winters with the shortest duration were from the current dataset (Appendix H). The median reduction in number of days that the lake was frozen between those two reporting periods was 23 days (average 27 days; Table 8-1). There has been less change in the date that Alta Lake freezes over (ice-on) than the date it thaws (ice-off). The median ice-on date is six days later in the recent dataset compared to earlier dataset: December 18th compared to December 12th. The median ice-off date for the recent dataset is April 10th compared to April 23rd for the earlier dataset, which represents a reduction of 13 days.

During the past winter, 2018-19, Alta Lake was frozen from January 1st to April 12th. While that 102 day duration was close to the average for the 2002-2019 period of records, it was still below the average of the earlier recording period. These two datasets provide convincing evidence of a changing climate that is consistent with other observations, for example, glacial recession within that same time. The conclusions would be stronger with more complete data from 1976 to 2001, and ideally data before 1942.

²⁹ Annual data has been supplied by Stephen Vogler. The 2019 date was emailed by him on January 10, 2020.

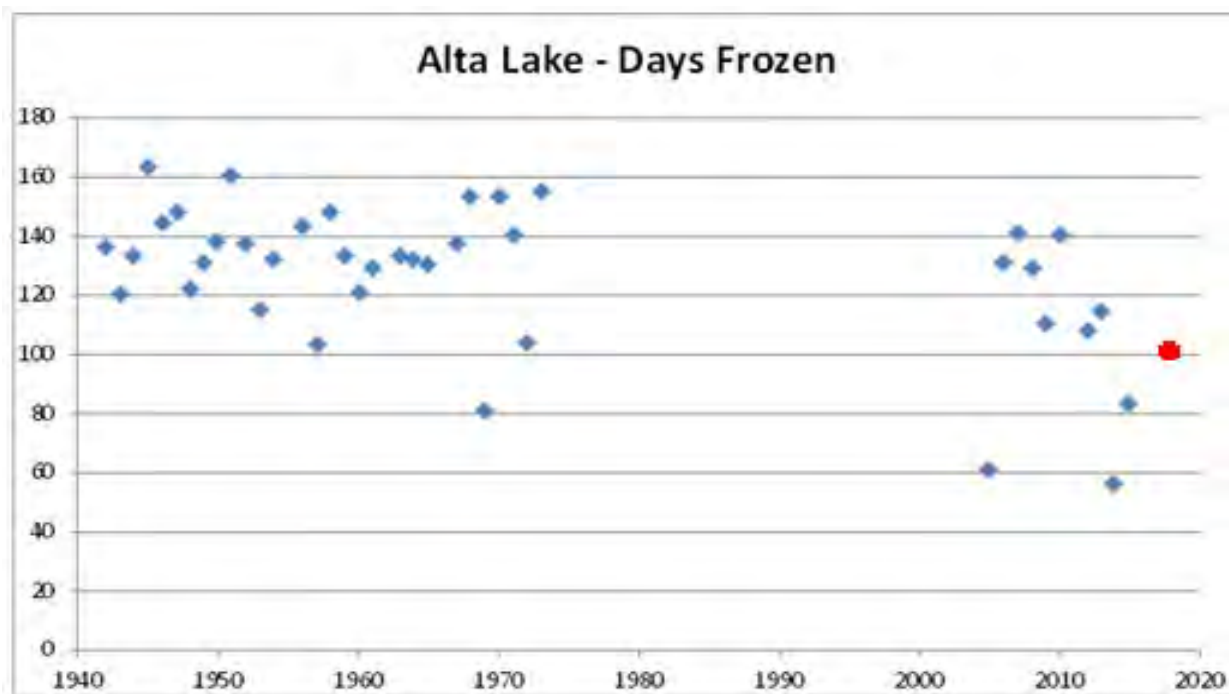


Figure 8-1. Ice records from Alta Lake from two datasets, 1942-1975 and 2002-2019. No data was recorded between those two periods.

Table 8-1. Ice records from Alta Lake from two datasets, 1942-1975 and 2002-2019. No data was recorded between these two periods. Some years did not record all data.

| | | Dataset | | | | Recent vs. Early Records |
|--------------------|----------|------------------|-----------|------------------|-----------|--------------------------|
| | | <u>1942-1975</u> | | <u>2002-2018</u> | | |
| | | Date | Day Count | Date | Day Count | |
| <u>Ice-On</u> | Records | n/a | 31 | n/a | 11 | 20 records fewer |
| | Earliest | 8-Nov-45 | 312 | 30-Nov-06 | 334 | 22 days later |
| | Latest | 15-Jan-70 | 380 | 6-Jan-06 | 371 | 9 days earlier |
| | Median | Dec. 12th | 346 | Dec. 18th | 353 | 6 days later |
| | Average | Dec. 12th | 346 | Dec. 16th | 351 | 5 days later |
| <u>Ice-Off</u> | | | | | | |
| | Records | n/a | 31 | n/a | 17 | 14 records fewer |
| | Earliest | 23-Mar-63 | 82 | 20-Feb-15 | 51 | 31 days earlier |
| | Latest | 21-May-52 | 142 | 29-Apr-08 | 120 | 22 days earlier |
| | Median | April 23rd | 113 | April 10th | 100 | 13 days earlier |
| | Average | April 23rd | 113 | April 5th | 95 | 18 days earlier |
| <u>Days Frozen</u> | | | | | | |
| | Records | n/a | 29 | n/a | 11 | 18 records fewer |
| | Median | n/a | 133 | n/a | 110 | 23 days shorter |
| | Average | n/a | 134 | n/a | 107 | 27 days shorter |

9. Conclusions

9.1 Plain Language Summary

The Resort Municipality of Whistler (RMOW) is located in the southern Coast Mountains of British Columbia, approximately 100 km north of the city of Vancouver. For many years the RMOW has been concerned about describing and conserving biodiversity within the urban development footprint of the RMOW. To help address those concerns, an Ecosystem Monitoring Program (the Program) was initiated by the RMOW in 2013. The objective of the Program was to describe both the waters and lands within the RMOW and to find out if there have been changes over time that might indicate a loss, or possible loss, of biodiversity.

Because it is not possible to look at everything in an ecosystem, the Program design was based on the use of indicators to describe and look for changes in the waters and lands over time. These indicators have included plants, animals, and specific parts of the environment. For 2019, the most important indicators in the Program included surface water, sediment, benthic invertebrates, and the fish community in streams, and Coastal Tailed Frogs (*Ascaphus truei*) and beavers (*Castor canadensis*) in areas within streams and ponds. Additional indicators for 2019 included Northern Goshawks (*Accipiter gentilis laingi*), Western Toads (*Anaxyrus boreas*), Western Screech Owls (*Megascops kennicottii*), black cottonwoods (*Populus trichocarpa*), and the timing and duration of ice on Alta Lake.

The 2019 stream Program looked at the main indicators listed above for four of the streams located within the RMOW: Jordan Creek, Crabapple Creek, The River of Golden Dreams, and 21-Mile Creek. The Program included collection and description of surface water and sediment samples, benthic invertebrate samples, and fish in the four streams.

9.1.1 Surface Water and Sediment

For surface water samples both the water chemistry, physical conditions (pH, temperature, conductivity, oxygen), and sediment chemistry were described and compared with the British Columbia Water Quality Guidelines (BCWQG). The BCWQG are guidelines that have been established by the BC Government so that if guidelines are met, then the plants and animals that live in the water will be protected, but if guidelines are not met, then there is an increased risk to the health of the plants and animals that live in the water.

The 2019 results for surface waters showed that BCWQG were met for all the chemical and physical measurements except for aluminum in Crabapple Creek. For sediment there were a few measurements that did not meet the BCWQG, including copper in Jordan Creek, and arsenic and copper in Crabapple Creek. Overall, the 2019 Program showed that the chemical and physical conditions in Jordan Creek, Crabapple Creek, The River of Golden Dreams, and 21-Mile Creek were suitable for supporting the continued health of the animal and plant communities living in these streams.

9.1.2 Benthic Invertebrates

Benthic invertebrates are animals that live in the sediment of streams, rivers, ponds, and lakes. These animals are useful for describing the health and biodiversity of communities that live in surface waters because they are numerous, are found in almost all habitats, do not migrate, are sensitive to pollution, are easy to collect, and can be easily identified. For the benthic invertebrate samples collected in 2019, the benthic invertebrate community was described and then assessed using a weight-of-evidence approach using three different methods.

For the benthic invertebrate samples collected in 2019, the benthic invertebrate community was described and then compared with the Fraser Basin 2014 Reference Model as developed through the Canadian Biomonitoring Network (CABIN) Program within Environment Canada. The CABIN Program collects benthic invertebrate samples from many sites within a region and over time develops an understanding of what the benthic invertebrate community should look like in a pristine, reference stream or river. Benthic invertebrate samples from test sites, such as at Whistler, can then be compared with the reference samples. If the benthic invertebrate communities from a test site are comparable with the community from the reference sites, then the test sites are said to be in 'reference condition' and in good health. If the test sites are slightly unusual in comparison reference sites, they are assessed as 'mildly divergent', and if the tests sites are highly unusual then they are assessed as 'divergent' to 'highly divergent'.

The 2019 results showed that the benthic invertebrate community was in reference condition for The River of Golden Dreams and Jordan Creek, and mildly divergent for Crabapple Creek and 21-Mile Creek. Results from previous years were also variable, with all creeks either in reference or mildly divergent most of the time. A closer look at the benthic invertebrate community structure also showed that the benthic invertebrate communities in the four streams appeared healthy, with pollution-sensitive species present and abundant at all sites. The weight-of-evidence therefore indicated that the four creeks were unimpaired, although Jordan Creek was relatively borderline.

For the benthic invertebrate samples collected in 2019, the benthic invertebrate community was also described using the Hilsenhoff Index of Biotic Integrity (HIBI). The HIBI is calculated using tolerance scores, which have been developed over time by experts and relate to the response of benthic invertebrates to organic pollution. The HIBI scores range from 0 to 10, with a score of 0 indicating that a site is dominated by pollution-sensitive benthic invertebrates and that there is no apparent organic pollution at the site, and a score of 10 indicating that a site is dominated by pollution-tolerant benthic invertebrates. The HIBI is of interest because of the potential for organic pollution in an urban setting, including from stormwater runoff, septic tank leakage, fertilizer runoff, and/or wildlife waste.

The 2019 results indicated that the benthic invertebrate community was in good to very good condition within the four streams that were assessed, with the potential for some slight amount organic pollution. Including the CABIN assessment, these results will be tested again in 2020 to find out if there have been any consistent changes over time that might be of concern.

9.1.3 Fish

A total of 179 fish were captured during the 2019 electrofishing and minnow trap efforts. As with previous years, three species of fish were captured in 2019: Threespine Stickleback (*Gasterosteus aculeatus*), undifferentiated trout fry from resident populations of Rainbow Trout (*Oncorhynchus mykiss*) and Coastal Cutthroat Trout (*O. clarkii clarkii*), and sculpin (*Cottus* sp.). The fish communities within the sampled creeks were inhabited by 0+ year fry and juvenile trout, demonstrating the importance of the four creeks as rearing and feeding habitat. The condition of the fish was generally good in 2019, in keeping with previous years: there were no differences in the length/weight relationship between sites.

The results of the 2019 Ecosystem Monitoring Program showed that Jordan Creek, Crabapple Creek, The River of Golden Dreams, and 21-Mile Creek had healthy benthic invertebrate and fish communities and that the 2019 results were consistent with earlier results collected in 2016, 2017, and 2018.

9.1.4 Coastal Tailed Frogs

Fifteen sites were surveyed for Coastal Tailed Frogs (*Ascaphus truei*) in 2019 with a continued emphasis on previously unsurveyed creeks on the west side of Whistler Valley (Van West Creek, Sproatt Creek, and “FJ West Creek”). No evidence of negative impacts was detected at any creek, including the two on Whistler Mountain which could potentially be impacted by ski and mountain bike activities (Whistler Creek and Archibald Creek).

Mapping irregularities were discovered at lower elevations of the west-side creeks during 2018 tailed frog surveys. Further investigation in 2019 located where Sproatt Creek is diverted underground upstream of the CN Rail tracks (which is why it is dry downstream in low flows). It also confirmed that a branch mapped southeast to Alpha Lake is currently dry since there is no connection with the main stream. Evidence of extensive flooding and stream diversion on FJ West Creek was traced to a storm in November 2017. As a result, the RMOW’s stream mapping in that area is no longer correct.

9.1.5 Beavers

The primary goal for beaver (*Castor canadensis*) monitoring since 2016 has been to establish a full survey of Whistler’s active lodges. Work in 2019 built on past years and resulted in the most comprehensive beaver survey to date. More active lodges were recorded than ever before (27), mainly due to determined efforts to fully access the Miller Creek Wetlands. A total of eight lodges were detected (compared to two in 2018) which confirmed for the first time that the Miller Creek Wetlands provide beaver habitat as significant as the River of Golden Dreams (where seven lodges were found).

Based on number of beavers per lodge, the beaver population is now estimated to be 157 in Whistler Valley (low to high estimates range from 113 to 173 beavers). There was no direct evidence that lodges were lost due to development or other human activities in the past year. Valley Trail construction beside the Millar Creek Wetlands has apparently not prevented the large beaver population from thriving. Lowering of the water level in Alta Vista Pond by the RMOW Roads Department reduced the area of beaver habitat but that lodge nonetheless remains active. Beavers were detected near the western entrance of Lost Lake Park in 2018 but not in 2019. It possible that the new Muffin Man bike trail had a negative impact on beavers, but

no direct evidence of this was detected. Based on 2019 observations, Whistler's beaver therefore appears to be stable.

The total area of beaver-affected wetlands in 2019 was estimated at 100.3 hectares, or approximately two-thirds the total area of Whistler's wetlands. This total includes six hectares added in 2019; field surveys revealed more beaver-caused flooding in the Miller Creek Wetlands than was mapped in 2018 from air photos. The total area of beaver-affected wetlands in Whistler can therefore be considered stable or possibly expanding slightly.

9.1.6 Other Indicators

Three exploratory indicators added to the program in 2018 were again included in 2019: Northern Goshawk (*Accipiter gentilis laingi*), Western Toad (*Anaxyrus boreas*), and black cottonwood (*Populus trichocarpa*). Recent breeding of Northern Goshawks has been sporadically documented since 2011 in Whistler's unlogged forests at low elevations. There were 11 records of goshawks, including evidence of breeding near Comfortably Numb Trail. Based on a concentration of visual records between Whistler Creekside and Kadenwood, it is possible there was a second pair in 2019 there but no evidence of breeding was reported. No evidence of breeding of Western Toads has been recently found south of Lost Lake, including in 2019. Several new ponds adjacent to Highway 99 in the Callaghan Creek area were confirmed to be suitable for inclusion in 2020 surveys. A plan to improve the mapping of black cottonwoods in Whistler Valley using the RMOW's new LIDAR layer was unsuccessful which means field surveys will be needed instead.

10. Certification

This report was prepared and reviewed by the undersigned:

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Appendix A

Daily Stream Temperature Data

| | | | | River of Golden | | | | |
|---------|-----------|------|------------|-----------------|--------------|--------|--------------|-------------|
| | Month | Year | month-year | BRB-DS_AQ010 | Jordan Creek | Dreams | Scotia Creek | Alpha Creek |
| Average | August | 2018 | Aug-18 | 14.229 | 17.012 | 11.939 | 11.364 | 10.499 |
| | September | 2018 | Sep-18 | 9.775 | 11.565 | 7.734 | 7.162 | 6.36 |
| | October | 2018 | Oct-18 | 5.811 | 7.597 | 5.619 | 5.265 | 4.796 |
| | November | 2018 | Nov-18 | 3.921 | 3.332 | 2.554 | 0.842 | 1.811 |
| | December | 2018 | Dec-18 | 1.622 | 2.04 | 1.365 | 0.189 | 1.321 |
| | January | 2019 | Jan-19 | 1.601 | 1.494 | 0.977 | -2.994 | 0.746 |
| | February | 2019 | Feb-19 | 0.325 | 1.079 | 0.958 | -1.327 | 0.55 |
| | March | 2019 | Mar-19 | 1.647 | 2.703 | 2.89 | 1.953 | 1.718 |
| | April | 2019 | Apr-19 | 4.176 | 7.063 | 4.491 | 3.905 | 2.871 |
| | May | 2019 | May-19 | 7.542 | 9.731 | 5.609 | 6.579 | 6.438 |
| | June | 2019 | Jun-19 | 11.143 | 13.664 | 9.751 | 10.544 | 9.621 |
| | July | 2019 | Jul-19 | 12.848 | 16.381 | 12.524 | 12.46 | 10.574 |
| MIN | August | 2018 | Aug-18 | 10.761 | 14.553 | 9.287 | 8.469 | 7.72 |
| | September | 2018 | Sep-18 | 6.687 | 8.369 | 4.272 | 4.037 | 3.221 |
| | October | 2018 | Oct-18 | 3.906 | 5.642 | 3.221 | -0.088 | 1.967 |
| | November | 2018 | Nov-18 | 2.021 | 0.88 | 0.522 | -3.986 | 0.051 |
| | December | 2018 | Dec-18 | 0.051 | 1.588 | 0.246 | -2.859 | 0.051 |
| | January | 2019 | Jan-19 | -0.032 | 0.825 | 0.024 | -10.898 | 0.024 |
| | February | 2019 | Feb-19 | -0.032 | 0.687 | 0.19 | -4.987 | 0.218 |
| | March | 2019 | Mar-19 | -0.032 | 1.18 | 0.66 | 0.135 | 0.825 |
| | April | 2019 | Apr-19 | 2.584 | 4.011 | 2.744 | -0.255 | 1.534 |
| | May | 2019 | May-19 | 3.722 | 7.167 | 3.354 | 3.063 | 2.903 |
| | June | 2019 | Jun-19 | 7.242 | 11.516 | 5.565 | 5.308 | 6.94 |
| | July | 2019 | Jul-19 | 10.59 | 14.146 | 9.854 | 6.763 | 8.02 |
| MAX | August | 2018 | Aug-18 | 17.272 | 20.103 | 15.008 | 13.786 | 13.137 |
| | September | 2018 | Sep-18 | 13.137 | 14.984 | 10.418 | 9.731 | 8.99 |
| | October | 2018 | Oct-18 | 7.77 | 9.509 | 7.217 | 7.494 | 6.839 |
| | November | 2018 | Nov-18 | 7.745 | 5.745 | 4.766 | 5.05 | 4.089 |
| | December | 2018 | Dec-18 | 3.354 | 2.53 | 2.047 | 2.744 | 2.155 |
| | January | 2019 | Jan-19 | 2.956 | 2.021 | 2.101 | 1.044 | 1.967 |

| Month | Year | month-year | BRB-DS_AQ010 | Jordan Creek | River of Golden Dreams | Scotia Creek | Alpha Creek |
|----------|------|------------|--------------|--------------|---------------------------|--------------|-------------|
| February | 2019 | Feb-19 | 2.503 | 1.967 | 2.047 | 0.273 | 0.907 |
| March | 2019 | Mar-19 | 5.024 | 4.688 | 5.949 | 4.402 | 2.69 |
| April | 2019 | Apr-19 | 7.192 | 10.956 | 7.116 | 7.77 | 5.076 |
| May | 2019 | May-19 | 12.823 | 12.727 | 9.336 | 11.005 | 10.98 |
| June | 2019 | Jun-19 | 14.266 | 16.915 | 13.666 | 14.96 | 11.516 |
| July | 2019 | Jul-19 | 14.984 | 19.056 | 14.481 | 17.938 | 11.953 |

Appendix B

Benthic Invertebrate
Taxonomy Results and
CABIN Outputs

Methods and QC Report 2019

Project ID: Whistler 160255

Client: Palmer Environmental

The logo for Cordillera Consulting, featuring the company name in white text on a green rectangular background.

Cordillera
Consulting

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Sample Reception

On August 21, 2019, Cordillera Consulting received 6 benthic samples from palmer Environmental. When samples arrived to Cordillera Consulting, exterior packaging was initially inspected for damage or wet spots that would have indicated damage to the interior containers.

Samples were logged into a proprietary software database (INSTAR1) where the clients assigned sample name was recorded along with a Cordillera Consulting (CC) number for cross-reference. Each sample was checked to ensure that all sites and replicates recorded on field sheets or packing lists were delivered intact and with adequate preservative. Any missing, mislabelled or extra samples were reported to the client immediately to confirm the total numbers and correct names on the sample jars. The client representative was notified of the arrival of the shipment and provided a sample inventory once intake was completed.

See table below for sample inventory:

Table 1: Summary of sample information including Cordillera Consulting (CC) number

| Sample | Site Code | CC# | Date | Size | # of Jars |
|---------------|-------------|----------|-----------|-------|-----------|
| RGD-AQ11 | RGD-AQ11 | CC200370 | 7/30/2019 | 400µM | 1 |
| RGD-AQ11QA/QC | RGD-AQ11 | CC200371 | 7/30/2019 | 400µM | 2 |
| CRB-DS-AQ01 | CRB-DS-AQ01 | CC200372 | 7/30/2019 | 400µM | 2 |
| RGD-DS-AQ12 | RGD-DS-AQ12 | CC200373 | 7/31/2019 | 400µM | 1 |
| JOR-DS-AQ31 | JOR-DS-AQ31 | CC200374 | 7/31/2019 | 400µM | 1 |
| 21M-DS-AQ21 | 21M-DS-AQ21 | CC200375 | 7/30/2019 | 400µM | 1 |

Sample Sorting

- Using a gridded Petri dish, fine forceps and a low power stereo-microscope (Olympus, Nikon, Leica) the sorting technicians removed the invertebrates and sorted them into family/orders.
- The sorting technician kept a running tally of total numbers excluding organisms from Porifera, Nemata, Platyhelminthes, Ostracoda, Copepoda, Cladocera and terrestrial drop-ins such as aphids. These organisms were marked for their presence (given a value of 1) only and left in the sample. They were not included towards the 300-organism subsample count.
- Where specimens are broken or damaged, only heads were counted.
- Subsampling was conducted with the use of a Marchant Box.
- When using the Marchant box, cells were extracted at the same time in the order indicated by a random number table. If the 300th organism was found part way into sorting a cell then the balance of that cell was sorted. If the organism count had not reached 300 by the 50th cell then the entire sample was sorted.
- The total number of cells sorted and the number of organisms removed were recorded manually on a bench sheet and then recorded into INSTAR1
- Organisms were stored in vials containing 80% ethanol and an interior label indicating the site names, date of sampling, site code numbers and portion subsampled. This information was also recorded on the laboratory bench sheet and on INSTAR1.
- The sorted portion of the debris was preserved and labeled separately from the unsorted portion and was tested for sorting efficiency (Sorting Quality Control – Sorting Efficiency). The unsorted portion was also labeled and preserved in separate jars.

Percent sub-sampled and total countable invertebrates pulled from the samples were summarized in the table below.

Table 2: Percent sub-sample and invertebrate count for each sample

| Sample | Date | CC# | 400 micron fraction | |
|---------------|-----------|----------|---------------------|-----------------|
| | | | % Sampled | # Invertebrates |
| RGD-AQ11 | 30-Jul-19 | CC200370 | 27% | 320 |
| RGD-AQ11QA/QC | 30-Jul-19 | CC200371 | 100% | 6 |
| CRB-DS-AQ01 | 30-Jul-19 | CC200372 | 10% | 350 |
| RGD-DS-AQ12 | 31-Jul-19 | CC200373 | 23% | 313 |
| JOR-DS-AQ31 | 31-Jul-19 | CC200374 | 35% | 328 |
| 21M-DS-AQ21 | 30-Jul-19 | CC200375 | 18% | 323 |

Sorting Quality Control - Sorting Efficiency

As a part of Cordillera's laboratory policy, all projects undergo sorting efficiency checks.

- As sorting progresses, 10% of samples were randomly chosen by senior members of the sorting team for resorting.
- All sorters working on a project had at least 1 sample resorted by another sorter.
- An efficiency of 90 % was expected (95% for CABIN samples).
- If 90/95% efficiency was not met, samples from that sorter were resorted.
- To calculate sorting efficiency the following formula was used:

$$\frac{\text{\#Organisms Missed}}{\text{Total Organisms Found}} * 100 = \%OM$$

Table 3: Summary of sorting efficiency

| | | | | Total from Sample | Percent Efficiency |
|--|--|----------|--|-------------------|--------------------|
| | | | | | |
| Site - QC, Sample - QC1, CC# - CC200375, Percent sampled = 18%, Sieve size = 400 | | | | | |
| Ephemeroptera | | 1 | | | |
| Total: | | 1 | | 323 | 100% |

Taxonomic Effort

The next procedure was the identification to genus-species level where possible of all the organisms in the sample.

- Identifications were made at the genus/species level for all insect organisms found including Chironomidae (Based on CABIN protocol).
- Non-insect organisms (except those not included in CABIN count) were identified to genus/species where possible and to a minimum of family level with intact and mature specimens.
- The Standard Taxonomic Effort lists compiled by the CABIN manual¹, SAFIT², and PNAMP³ were used as a guide line for what level of identification to achieve where the condition and maturity of the organism enabled.
- Organisms from the same families/order were kept in separate vials with 80% ethanol and an interior label of printed laser paper.
- Chironomidae was identified to genus/species level where possible and was aided by slide mounts. CMC-10 was used to clear and mount the slide.
- Oligochaetes was identified to family/genus level with the aid of slide mounts. CMC-10 was used to clear and mount the slide.
- Other Annelida (leeches, polychaetes) were identified to the family/genus/species level with undamaged, mature specimens.

- Mollusca was identified to family and genus/species where possible
- Decapoda, Amphipoda and Isopoda were identified at family/genus/species level where possible.
- Bryozoans and Nemata remained at the phylum level
- Hydrachnidae and Cnidaria were identified at the family/genus level where possible.
- When requested, reference collections were made containing at least one individual from each taxa listed. Organisms represented will have been identified to the lowest practical level.
- Reference collection specimens were stored in 55 mm glass vials with screw-cap lids with polyseal inserts (museum quality). They were labeled with taxa name, site code, date identified and taxonomist name. The same information was applied to labels on the slide mounts.

Taxonomy Notes: *Baetis tricaudatus* group has now been renamed to *Baetis rhodani* group. There has been no change in the determination of the taxa. See Webb 2017 in the taxonomy keys.

Taxonomists

The taxonomists for this project were certified by the Society of Freshwater Science (SFS) Taxonomic Certification Program at level 2 which is the required certification for CABIN projects:

Scott Finlayson: Group 1 General Arthropods (East/West); Group 2 EPT (East/West); Group 3 Chironomidae (East/West); Group 4 Oligochaeta

Adam Bliss: Group 1 General Arthropods (East/West); Group 2 EPT (East/West); Group 3 Chironomidae

Rita Avery: Group 1 General Arthropods (East/West); Group 2 EPT (East/West)

Taxonomic QC

Taxonomic QC was performed in house by someone other than the original taxonomist.

- Quality control protocol involved complete, blind re-identification and re-enumeration of at least 10% of samples by a second SFS-certified taxonomist.
- Samples for taxonomic quality control were randomly selected and quality control procedures were conducted as the project progresses through the laboratories.
- The second (QC) taxonomist will calculate and record four types of errors:
 1. Misidentification error
 2. Enumeration error
 3. Questionable taxonomic resolution error
 4. Insufficient taxonomic resolution error

The QC coordinator then calculates the following estimates of taxonomic precision.

1. The percent total identification error rate is calculated as:

$$\frac{\text{Sum of incorrect identifications}}{\text{total organisms counted in audit}} * (100)$$

The average total identification error rate of audited samples did not exceed 5%. All samples that exceed a 5% error rate were re-evaluated to determine whether repeated errors or patterns in error contributed.

2. The percent difference in enumeration (PDE) to quantify the consistency of specimen counts.

$$PDE = \frac{|n_1 - n_2|}{n_1 + n_2} \times 100$$

3. The percent taxonomic disagreement (PTD) to quantify the shared precision between two sets of identifications.

$$PTD = \left(1 - \left[\frac{a}{N}\right]\right) \times 100$$

4. Bray Curtis dissimilarity Index to quantify the differences in identifications.

$$BC_{ij} = 1 - \frac{2C_{ij}}{S_j + S_i}$$

Error Summary

All samples report errors within the acceptable limits for CABIN Laboratory methods (less than 5% error).

Table 4: Summary of taxonomic error following QC

| Site | Taxa Identified | % Error | PDE | PTD | Bray - Curtis Dissimilarity index |
|---|-----------------|---------|------------|--------|-----------------------------------|
| Site - 2019, Sample - RGD-AQ11, CC# - CC200370, Percent sampled = 27%, Sieve size = 400 | 319 | 0.31 | 0.15649452 | 2.1875 | 0.02034429 |

There will always be disagreements between taxonomists regarding the degree of taxonomic resolution in immature specimens and when laboratories make use of different keys for certain groups (Mollusks is an especially disputed group). It is always possible that some taxa found by the original taxonomist were overlooked in QC.

All of the Taxonomic QC samples that were observed passed testing according to the CABIN misidentification protocols. See the tables below for results from taxonomic QC audit.

Error Rationale

| Site - 2019, Sample - RGD-AQ11, CC# - CC200370, Percent sampled = 27%, Sieve size = 400 | Laboratory Count | QC Audit Count | Agreement | Misidentification | Questionable Taxonomic Resolution | Enumeration | Insufficient Taxonomic Resolution | Comments |
|---|------------------|----------------|-----------|-------------------|-----------------------------------|-------------|-----------------------------------|-------------------|
| Oreodytes | 2 | 2 | | | | | | |
| Chironomidae | 1 | 1 | | | | | | |
| Microtendipes | 1 | 1 | | | | | | |
| Polypedilum | 1 | 1 | | | | | | |
| Micropsectra | 6 | 5 | No | | | X | | |
| Tanytarsus | 1 | 2 | No | 1 | | X | | From Micropsectra |
| Eukiefferiella | 4 | 4 | | | | | | |
| Tvetenia | 1 | 1 | | | | | | |
| Tanypodinae | 1 | 1 | | | | | | |
| Thienemannimyia group | 1 | 1 | | | | | | |
| Simuliidae | 1 | 1 | | | | | | |
| Helodon | 1 | 1 | | | | | | |
| Simulium | 52 | 52 | | | | | | |
| Ameletus | 14 | 14 | | | | | | |
| Baetidae | 17 | 17 | | | | | | |
| Dicranota | 1 | 1 | | | | | | |
| Baetis | 21 | 22 | No | | | X | | |
| Drunella grandis group | 1 | 1 | | | | | | |
| Drunella spinifera | 1 | 1 | | | | | | |
| Heptageniidae | 13 | 17 | No | | | X | | |
| Cinygmula | 20 | 16 | No | | | X | | |
| Epeorus | 29 | 29 | | | | | | |
| Rhithrogena | 7 | 7 | | | | | | |
| Leptophlebiidae | 2 | 2 | | | | | | |
| Capniidae | 5 | 5 | | | | | | |
| Chloroperlidae | 6 | 6 | | | | | | |
| Sweltsa | 22 | 22 | | | | | | |
| Malenka | 1 | 1 | | | | | | |
| Zapada | 3 | 3 | | | | | | |
| Zapada columbiana | 1 | 1 | | | | | | |
| Perlidae | 3 | 3 | | | | | | |
| Calineuria californica | 1 | 1 | | | | | | |

| | | | | | | | | |
|----------------------------------|--------------------|------------|------|------|---|---|---|--|
| Megarcys | 1 | 1 | | | | | | |
| Perlodidae | 2 | 2 | | | | | | |
| Rhyacophila angelita group | 1 | 1 | | | | | | |
| Crangonyx | 4 | 4 | | | | | | |
| Atractides | 3 | 3 | | | | | | |
| Sperchon | 6 | 5 | No | | | X | | |
| Doroneuria | 1 | 1 | | | | | | |
| Baetis rhodani group | 59 | 58 | No | | | X | | |
| Tubificinae with hair chaetae | 1 | 1 | | | | | | |
| Tubificinae without hair chaetae | 1 | 1 | | | | | | |
| | | | | | | | | |
| | | | | | | | | |
| Total: | 320 | 319 | | | | | | |
| | | | | | 0 | 7 | 0 | |
| % Total Misidentification Rate = | misidentifications | x100 = | 0.31 | Pass | | | | |
| | total number | | | | | | | |

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¹ McDermott, H., Paull, T., Strachan, S. (May 2014). Laboratory Methods: Processing, Taxonomy, and Quality Control of Benthic Macroinvertebrate Samples, Environment Canada. ISBN: 978-1-100-25417-3

² Southwest Association of Freshwater Invertebrate Taxonomists. (2015). www.safit.org

³ Pacific Northwest Aquatic Monitoring Partnership (Accessed 2015). www.pnamp.org

Taxonomic Keys

Below is a reference list of taxonomic keys utilized by taxonomists at Cordillera Consulting. Cordillera taxonomists routinely seek out new literature to ensure the most accurate identification keys are being utilized. This is not reflective of the exhaustive list of resources that we use for identification. A more complete list of taxonomic resources can be found at Southwest Association of Freshwater Invertebrate Taxonomists. (2015).

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Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | 21M-DS-AQ21 |
| Sampling Date | Aug 03 2016 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12758 N, 122.97288 W |
| Altitude | 632 |
| Local Basin Name | Twenty-One Mile Creek |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 6.1% | 4.0% | 28.9% | 28.9% | 20.6% | 11.5% |
| CABIN Assessment of 21M-DS-AQ21 on Aug 03, 2016 | Mildly Divergent | | | | | |

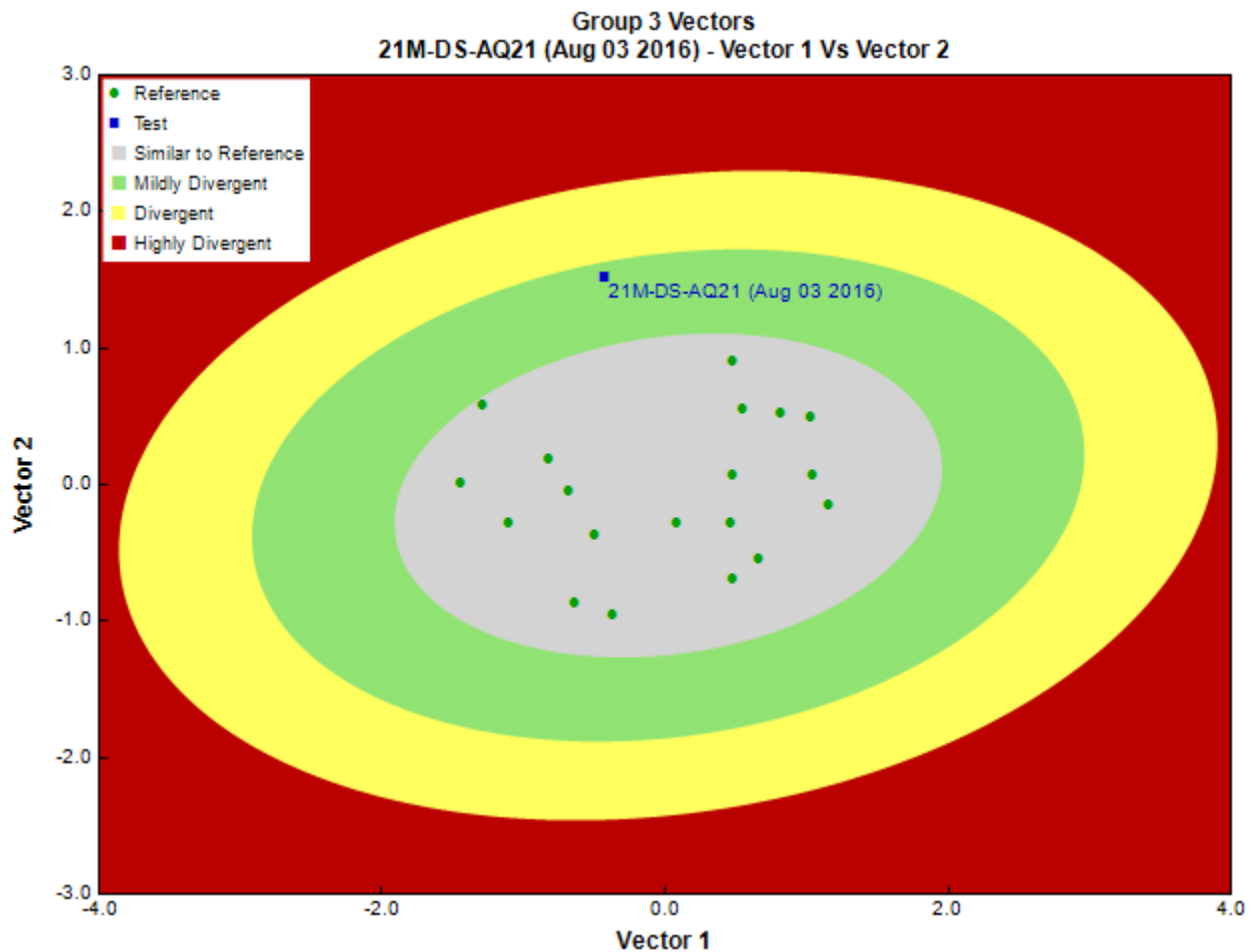


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | October 11, 2016 |
| | Marchant Box |
| Sub-Sample Proportion | 20/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 8 | 40.0 |
| | Insecta | Diptera | Ceratopogonidae | 2 | 10.0 |
| | | | Chironomidae | 22 | 110.0 |
| | | | Empididae | 1 | 5.0 |
| | | | Simuliidae | 29 | 145.0 |
| | | | Tipulidae | 1 | 5.0 |
| | | Ephemeroptera | Baetidae | 103 | 515.0 |
| | | | Ephemerellidae | 4 | 20.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | | Heptageniidae | 48 | 240.0 |
| | | Plecoptera | Chloroperlidae | 12 | 60.0 |
| | | | Nemouridae | 65 | 325.0 |
| | | | Perlodidae | 2 | 10.0 |
| | | Trichoptera | Rhyacophilidae | 7 | 35.0 |
| | | | Total | 304 | 1,520.0 |

Metrics

| Name | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.74 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 39.8 | 55.3 \pm 17.5 |
| % Predatores | 22.7 | 22.0 \pm 15.6 |
| % Scrapers | 59.2 | 53.5 \pm 23.0 |
| % Shredder | 21.7 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 8.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 20.7 | 20.6 \pm 17.1 |
| % EPT Individuals | 79.3 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 87.8 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.8 \pm 0.2 |
| Total Abundance | 1520.0 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 0.0 | 0.2 \pm 0.4 |
| Diptera taxa | 5.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 3.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 1205.0 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 7.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 1.9 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.4 | 0.3 \pm 0.1 |
| Total No. of Taxa | 13.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 1.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.32 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.08 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.23 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.08 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.51 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.24 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.97 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.77 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.21 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.45 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.25 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.19 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.88 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.42 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.14 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.37 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.26 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.15 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.24 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.22 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.21 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.18 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.17 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.03 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.40 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.59 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecchynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.19 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.66 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.19 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.07 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.10 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.07 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.23 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.38 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.59 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.53 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.20 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.18 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 7.14 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 1.12 |
| RIVPACS : Expected taxa P>0.70 | 4.26 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.17 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-Avg (cm) | 10.7 | 28.5 \pm 10.6 |
| Depth-BankfullMinusWetted (cm) | 100.00 | 163.00 |
| Depth-Max (cm) | 29.0 | 44.5 \pm 18.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 0 \pm 1 |
| Slope (m/m) | 0.0300000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.58 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 0.93 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 11.2 | 85.0 \pm 66.5 |
| Width-Wetted (m) | 9.6 | 23.1 \pm 31.8 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.24000 | 16.49843 \pm 2.42987 |
| Landcover | | |
| Natl-SnowIce (%) | 26.43000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| %Cobble (%) | 18 | 63 \pm 4 |
| %Gravel (%) | 3 | 3 \pm 4 |
| %Pebble (%) | 79 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 4.00 | 6.67 \pm 3.25 |
| Dg (cm) | 4.1 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.45000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-DO (mg/L) | 9.3900000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 6.3 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 40.5000000 | 74.4000000 \pm 44.3472660 |
| General-TempAir (Degrees Celsius) | 16.3 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 12.0000000 | 5.7731579 \pm 1.9704316 |
| General-Turbidity (NTU) | 2.6300000 | 1.3000000 \pm 0.9899495 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | 21M-DS-AQ21 |
| Sampling Date | Jul 25 2017 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12750 N, 122.97278 W |
| Altitude | 650 |
| Local Basin Name | Twenty-One Mile Creek |
| | River of Golden Dreams |
| Stream Order | 3 |

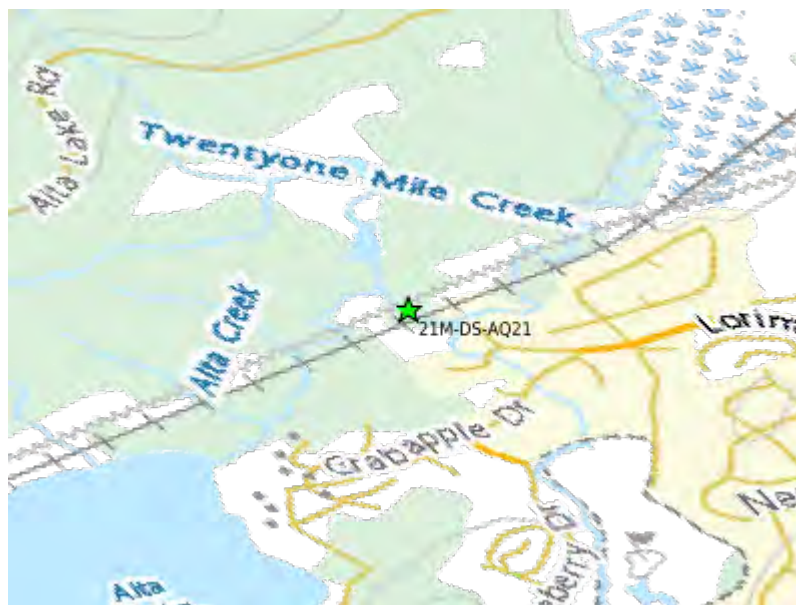


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st NatI-SnowIce NatI-Water NatI-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 1 2 3 4 5 6 |

| | | | | | | |
|--|-----------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 10.2% | 5.3% | 33.2% | 24.3% | 17.2% | 9.8% |
| CABIN Assessment of 21M-DS-AQ21 on Jul 25, 2017 | Divergent | | | | | |

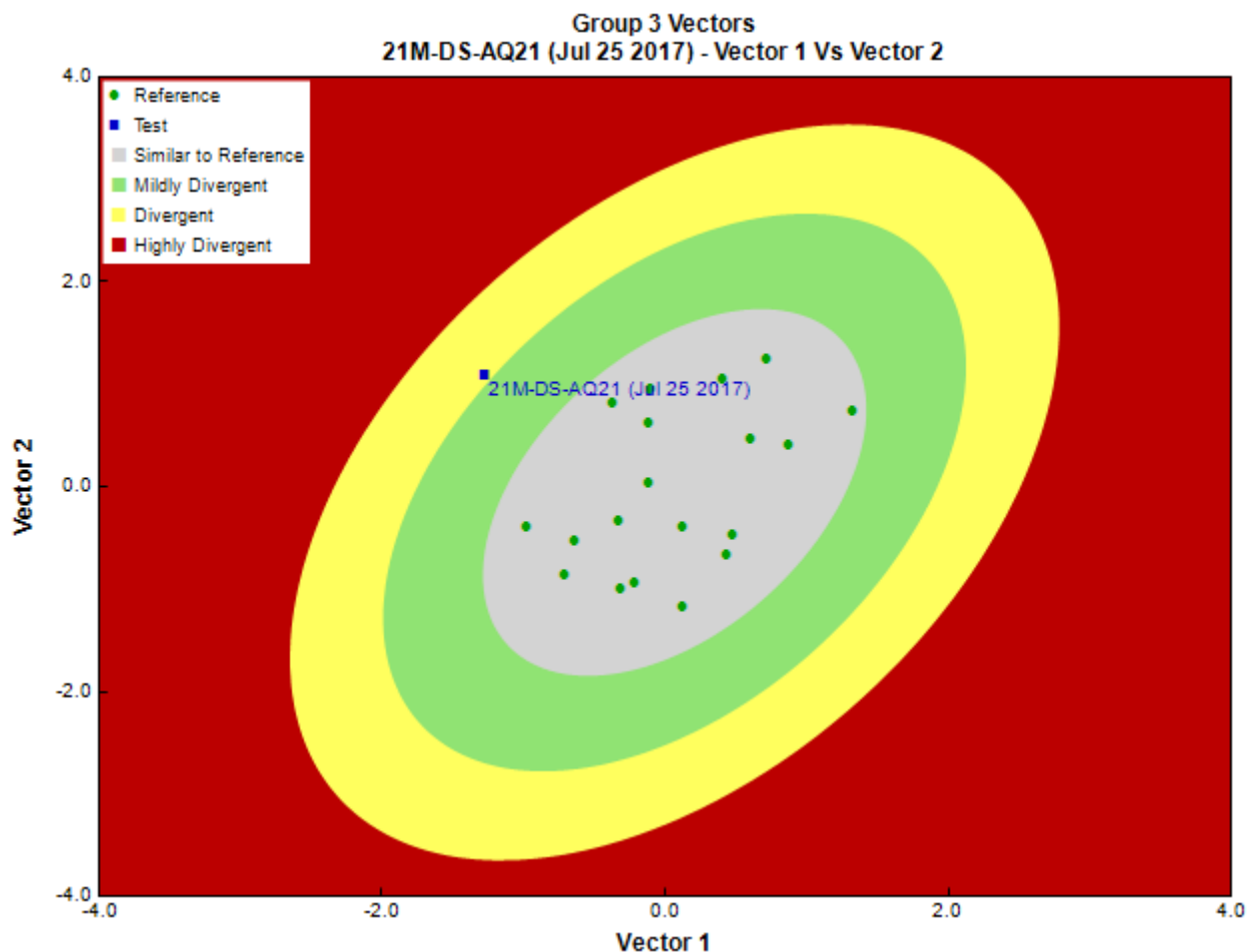


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | November 02, 2017 |
| | Marchant Box |
| Sub-Sample Proportion | 24/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 4 | 16.7 |
| | Insecta | Diptera | Ceratopogonidae | 7 | 29.2 |
| | | | Chironomidae | 11 | 45.8 |
| | | | Empididae | 1 | 4.2 |
| | | | Simuliidae | 50 | 208.3 |
| | | | Tipulidae | 2 | 8.3 |
| | | Ephemeroptera | Baetidae | 86 | 358.3 |
| | | | Heptageniidae | 209 | 870.8 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|------------|----------------|-----------|-------------|
| | | Plecoptera | Chloroperlidae | 15 | 62.5 |
| | | | Nemouridae | 2 | 8.3 |
| | | | Perlidae | 3 | 12.5 |
| | | | Perlodidae | 1 | 4.2 |
| | | | Total | 391 | 1,629.1 |

Metrics

| Name | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.78 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 16.6 | 55.3 \pm 17.5 |
| % Predatores | 17.9 | 22.0 \pm 15.6 |
| % Scrapers | 88.2 | 53.5 \pm 23.0 |
| % Shredder | 1.0 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 7.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 19.2 | 20.6 \pm 17.1 |
| % EPT Individuals | 80.8 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 94.9 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 1.0 | 0.8 \pm 0.2 |
| Total Abundance | 1629.2 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 0.0 | 0.2 \pm 0.4 |
| Diptera taxa | 5.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 2.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 1316.7 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 6.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.7 \pm 0.1 |
| Plecoptera taxa | 4.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 1.4 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.6 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 12.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 0.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.31 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.23 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.07 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.51 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.23 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.96 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.79 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.22 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.46 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.23 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.19 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halipidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.90 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.41 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.14 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.35 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.25 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.13 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.25 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.22 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.16 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.15 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.36 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.60 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.19 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.67 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.18 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.06 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.10 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.06 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.24 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.36 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.62 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.53 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.20 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.19 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 7.24 |
| RIVPACS : Observed taxa P>0.50 | 7.00 |
| RIVPACS : O:E (p > 0.5) | 0.97 |
| RIVPACS : Expected taxa P>0.70 | 4.30 |
| RIVPACS : Observed taxa P>0.70 | 4.00 |
| RIVPACS : O:E (p > 0.7) | 0.93 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-Avg (cm) | 25.2 | 28.5 \pm 10.6 |
| Depth-BankfullMinusWetted (cm) | 48.00 | 163.00 |
| Depth-Max (cm) | 38.0 | 44.5 \pm 18.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.66 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 1.02 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 11.5 | 85.0 \pm 66.5 |
| Width-Wetted (m) | 48.0 | 23.1 \pm 31.8 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.24000 | 16.49843 \pm 2.42987 |
| Landcover | | |
| MNP-WetlandHerb (%) | 0.00000 | 0.00000 \pm 0.00000 |
| Natl-SnowIce (%) | 26.43000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |
| %Cobble (%) | 7 | 63 \pm 4 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| %Gravel (%) | 30 | 3 \pm 4 |
| %Pebble (%) | 63 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 3.00 | 6.67 \pm 3.25 |
| Dg (cm) | 2.3 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 3 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.45000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-DO (mg/L) | 11.3300000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 7.1 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 40.0000000 | 74.4000000 \pm 44.3472660 |
| General-TempAir (Degrees Celsius) | 31.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 11.6000000 | 5.7731579 \pm 1.9704316 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | 21M-DS-AQ21 |
| Sampling Date | Jul 31 2018 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12767 N, 122.97298 W |
| Altitude | 645 |
| Local Basin Name | Twenty-One Mile Creek |
| | River of Golden Dreams |
| Stream Order | 3 |

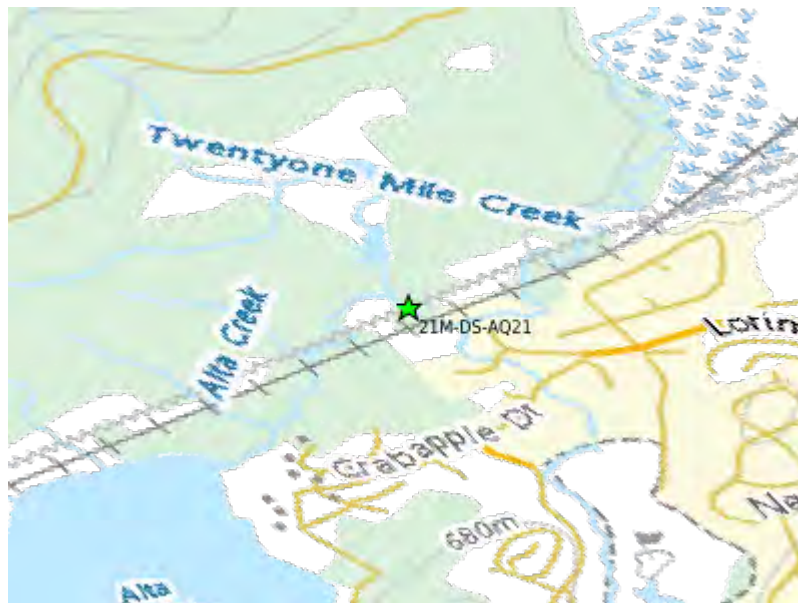


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st NatI-SnowIce NatI-Water NatI-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 10.4% | 4.7% | 22.1% | 17.5% | 38.9% | 6.5% |
| CABIN Assessment of 21M-DS-AQ21 on Jul 31, 2018 | Mildly Divergent | | | | | |

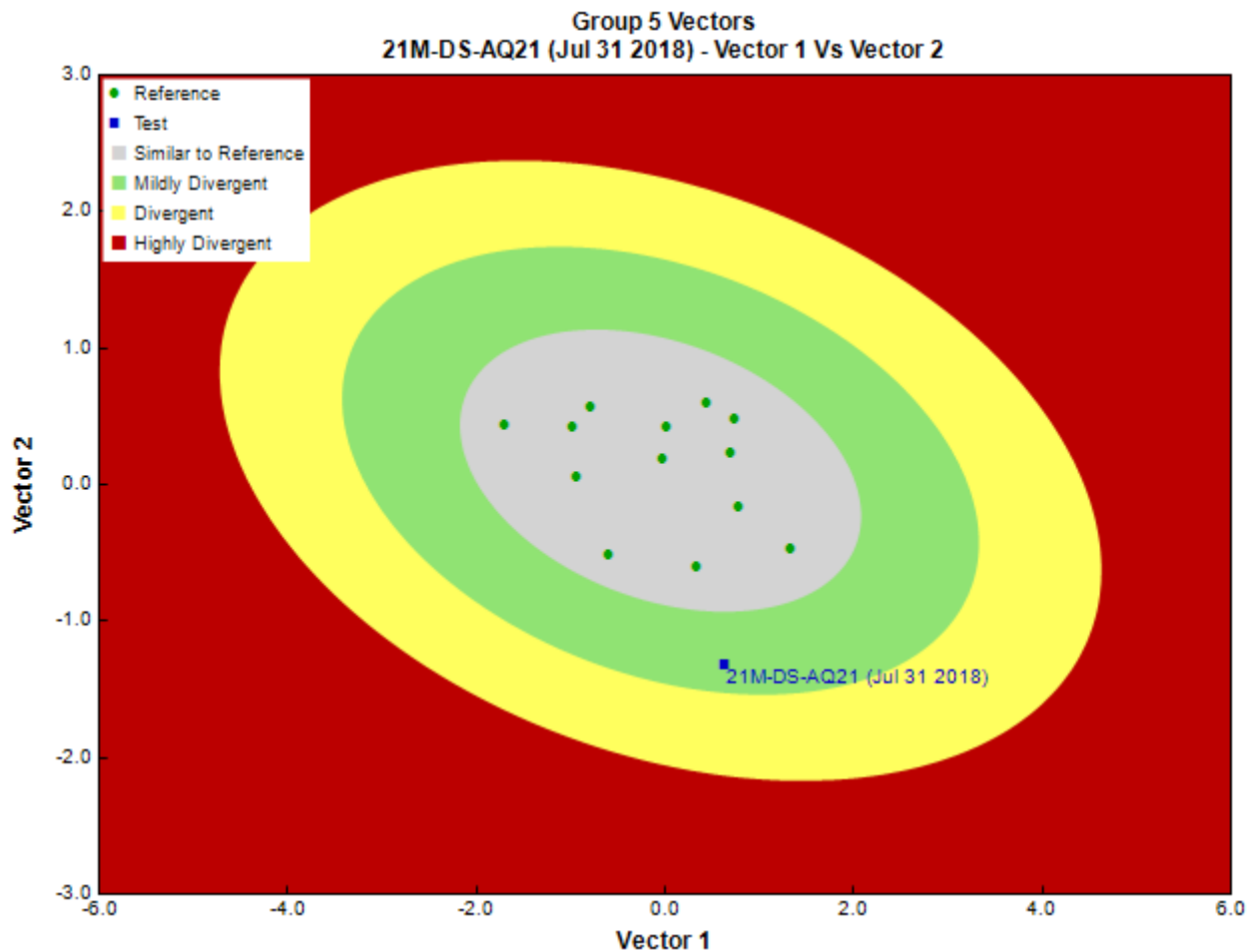


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 100/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|-------------------|-----------|-------------|
| Annelida | Clitellata | Lumbriculida | Lumbriculidae | 28 | 27.8 |
| Arthropoda | Arachnida | Trombidiformes | Hygrobatidae | 33 | 33.3 |
| | | | Lebertiidae | 11 | 11.1 |
| | | | Sperchontidae | 22 | 22.2 |
| | | | Torrenticolidae | 6 | 5.6 |
| | Insecta | Diptera | Ceratopogonidae | 45 | 44.5 |
| | | | Chironomidae | 145 | 144.5 |
| | | | Deuterophlebiidae | 6 | 5.6 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|---------------|-----------------|-----------|-------------|
| | | | Simuliidae | 489 | 488.9 |
| | | | Tipulidae | 1 | 1.0 |
| | | Ephemeroptera | Ameletidae | 28 | 27.8 |
| | | | Baetidae | 361 | 361.1 |
| | | | Ephemerellidae | 50 | 50.0 |
| | | | Heptageniidae | 412 | 411.2 |
| | | | Leptophlebiidae | 6 | 5.6 |
| | | Plecoptera | Capniidae | 6 | 5.6 |
| | | | Chloroperlidae | 128 | 127.8 |
| | | | Leuctridae | 6 | 5.6 |
| | | | Nemouridae | 23 | 22.3 |
| | | | Perlidae | 95 | 94.5 |
| | | | Perlodidae | 39 | 38.9 |
| | | Trichoptera | | 1 | 1.0 |
| | | | Hydroptilidae | 6 | 5.6 |
| | | | Rhyacophilidae | 44 | 44.4 |
| | | | Total | 1,991 | 1,985.9 |

Metrics

| Name | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.87 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 38.9 | 67.6 \pm 30.3 |
| % Predatores | 44.5 | 41.1 \pm 20.2 |
| % Scrapers | 64.1 | 34.3 \pm 21.0 |
| % Shredder | 1.7 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 25.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 39.5 | 47.4 \pm 26.3 |
| % EPT Individuals | 60.5 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 77.3 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.6 \pm 0.2 |
| Total Abundance | 1985.3 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.5 \pm 0.5 |
| Diptera taxa | 5.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 5.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 1200.0 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 13.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.6 \pm 0.1 |
| Plecoptera taxa | 6.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 2.3 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 23.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 2.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.23 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.06 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.05 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.78 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.22 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.15 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.47 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.31 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.98 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.66 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.03 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.22 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.36 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.25 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.77 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.06 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.15 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.80 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.37 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.17 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.05 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.41 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.26 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.26 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.00 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.24 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.19 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.17 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.24 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.23 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.04 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.49 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.50 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.06 |
| Pelecophlebiidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.03 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.17 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.57 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.28 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.12 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.09 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.09 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.34 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.20 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.32 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.03 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.49 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.50 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.22 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.13 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.06 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.04 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 4.56 |
| RIVPACS : Observed taxa P>0.50 | 6.00 |
| RIVPACS : O:E (p > 0.5) | 1.32 |
| RIVPACS : Expected taxa P>0.70 | 3.33 |
| RIVPACS : Observed taxa P>0.70 | 4.00 |
| RIVPACS : O:E (p > 0.7) | 1.20 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 18.5 | 40.5 \pm 22.4 |
| Macrophyte (PercentRange) | 1 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.23 \pm 0.44 |
| Reach-%Logging (PercentRange) | 0 | 0 \pm 0 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 |
| Reach-Pools (Binary) | 0 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 0 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.60 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 0.77 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 11.7 | 75.1 \pm 72.8 |
| Width-Wetted (m) | 10.9 | 50.6 \pm 60.4 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 171.50745 \pm 107.47690 |
| Temp07_JULmax (Degrees Celsius) | 18.24000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 26.43000 | 3.62533 \pm 10.17162 |
| Natl-Water (%) | 2.82000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 1 | 0 |
| %Cobble (%) | 3 | 58 |
| %Gravel (%) | 12 | 1 |
| %Pebble (%) | 84 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 0 | 0 |
| D50 (cm) | 3.00 | 3.30 |
| Dg (cm) | 2.8 | 6.6 |
| Dominant-1st (Category(0-9)) | 4 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 5 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 3 |
| SurroundingMaterial (Category(0-9)) | 3 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 39.45000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-DO (mg/L) | 14.6000000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 6.2 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 38.1000000 | 176.1000000 |
| General-TempWater (Degrees Celsius) | 19.9000000 | 13.2730769 \pm 4.7663725 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | 21M-DS-AQ21 |
| Sampling Date | Jul 30 2019 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12761 N, 122.97293 W |
| Altitude | 643 |
| Local Basin Name | Twenty-One Mile Creek |
| | River of Golden Dreams |
| Stream Order | 3 |

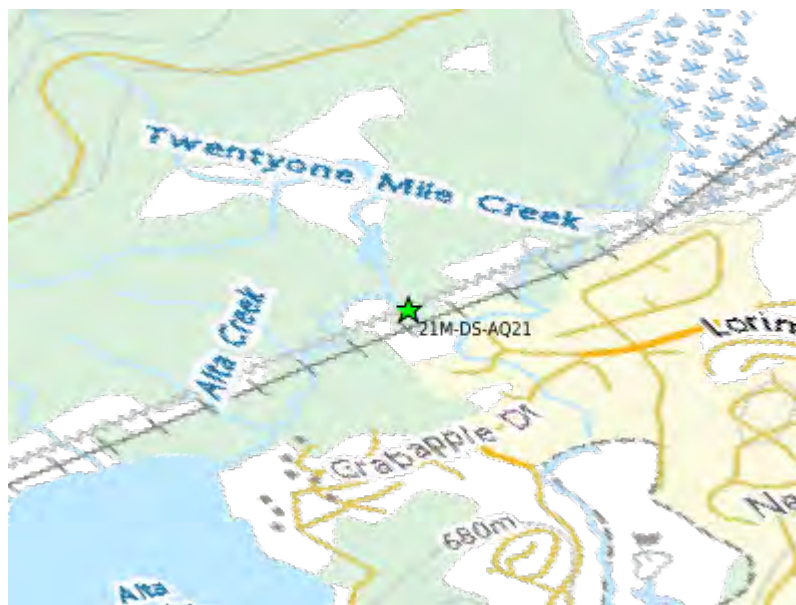


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 10.3% | 5.3% | 33.1% | 24.4% | 17.2% | 9.8% |
| CABIN Assessment of 21M-DS-AQ21 on Jul 30, 2019 | Mildly Divergent | | | | | |

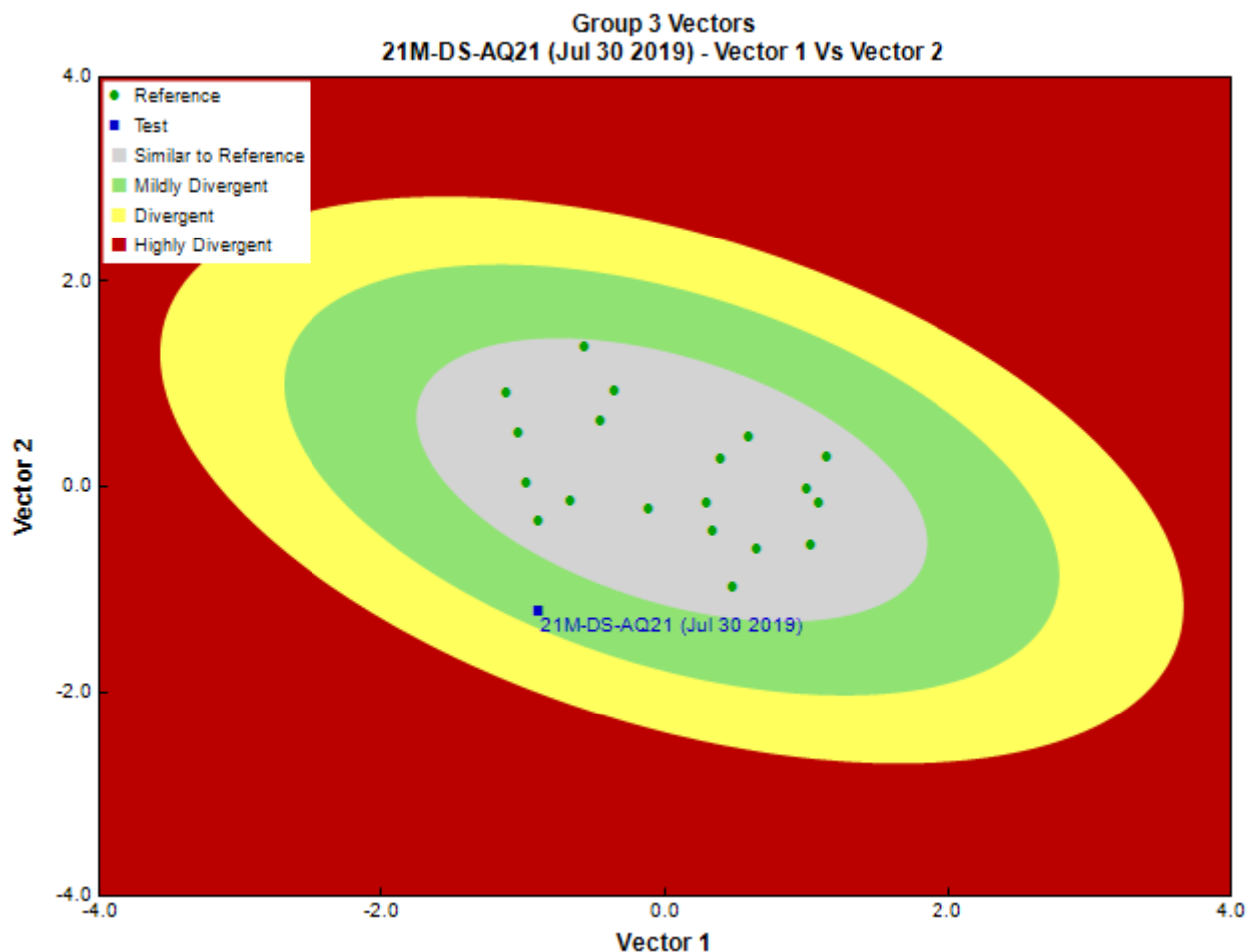


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 18/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|-----------------|-----------|-------------|
| Annelida | Clitellata | Tubificida | Naididae | 6 | 33.3 |
| Arthropoda | Arachnida | Sarcoptiformes | | 1 | 5.6 |
| | | Trombidiformes | Hydryphantidae | 1 | 5.6 |
| | | | Hygrobatidae | 2 | 11.1 |
| | | | Lebertiidae | 2 | 11.1 |
| | | | Sperchontidae | 1 | 5.6 |
| | | | Torrenticolidae | 1 | 5.6 |
| | Collembola | Collembola | | 1 | 5.6 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|---------|---------------|-----------------|-----------|-------------|
| | Insecta | Diptera | Ceratopogonidae | 2 | 11.1 |
| | | | Chironomidae | 21 | 116.8 |
| | | | Empididae | 2 | 11.2 |
| | | | Simuliidae | 41 | 227.8 |
| | | | Tipulidae | 1 | 5.6 |
| | | Ephemeroptera | Ameletidae | 16 | 88.9 |
| | | | Baetidae | 82 | 455.6 |
| | | | Ephemerellidae | 11 | 61.2 |
| | | | Heptageniidae | 89 | 494.5 |
| | | | Leptophlebiidae | 2 | 11.1 |
| | | Plecoptera | Capniidae | 2 | 11.1 |
| | | | Chloroperlidae | 14 | 77.8 |
| | | | Nemouridae | 8 | 44.5 |
| | | | Perlidae | 5 | 27.8 |
| | | | Perlodidae | 6 | 33.4 |
| | | Trichoptera | Hydropsychidae | 1 | 5.6 |
| | | | Limnephilidae | 2 | 11.1 |
| | | | Rhyacophilidae | 3 | 16.7 |
| | | | Total | 323 | 1,795.3 |

Metrics

| Name | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.75 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 35.0 | 55.3 \pm 17.5 |
| % Predatores | 26.6 | 22.0 \pm 15.6 |
| % Scrapers | 66.6 | 53.5 \pm 23.0 |
| % Shredder | 4.0 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 27.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 24.9 | 20.6 \pm 17.1 |
| % EPT Individuals | 75.1 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 77.6 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.8 \pm 0.2 |
| Total Abundance | 1794.4 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 0.0 | 0.2 \pm 0.4 |
| Diptera taxa | 5.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 5.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 1338.9 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 13.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.7 \pm 0.1 |
| Plecoptera taxa | 5.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 2.2 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 24.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 3.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.31 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.23 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.07 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.52 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.23 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.96 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.79 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.22 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.46 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.23 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.19 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halipidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.90 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.41 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.14 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.35 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.25 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.13 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.25 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.22 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.16 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.15 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.36 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.60 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecchynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.19 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at 21M-DS-AQ21 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.67 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.18 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.06 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.10 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.06 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.24 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.36 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.62 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.53 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.20 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.19 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 7.24 |
| RIVPACS : Observed taxa P>0.50 | 9.00 |
| RIVPACS : O:E (p > 0.5) | 1.24 |
| RIVPACS : Expected taxa P>0.70 | 4.30 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.16 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-BankfullMinusWetted (cm) | 20.00 | 163.00 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 0 \pm 1 |
| Slope (m/m) | 0.0100000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 0 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 0 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.70 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 0.89 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 11.4 | 85.0 \pm 66.5 |

Habitat Description

| Variable | 21M-DS-AQ21 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Width-Wetted (m) | 11.3 | 23.1 \pm 31.8 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.24000 | 16.49843 \pm 2.42987 |
| Landcover | | |
| Natl-SnowIce (%) | 26.43000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |
| %Cobble (%) | 3 | 63 \pm 4 |
| %Gravel (%) | 14 | 3 \pm 4 |
| %Pebble (%) | 82 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 3.80 | 6.67 \pm 3.25 |
| Dg (cm) | 2.9 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.45000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-Conductivity (μ S/cm) | 36.3000000 | 62.9529406 \pm 33.2341330 |
| General-DO (mg/L) | 9.7800000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 7.0 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 51.8000000 | 74.4000000 \pm 44.3472660 |
| General-TempWater (Degrees Celsius) | 13.3000000 | 5.7731579 \pm 1.9704316 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | CRB-DS-AQ01 |
| Sampling Date | Aug 02 2016 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12660 N, 122.97170 W |
| Altitude | 660 |
| Local Basin Name | Crabapple Creek |
| | River of Golden Dreams |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 43.6% | 26.5% | 0.1% | 19.1% | 8.6% | 2.2% |
| CABIN Assessment of CRB-DS-AQ01 on Aug 02, 2016 | Mildly Divergent | | | | | |

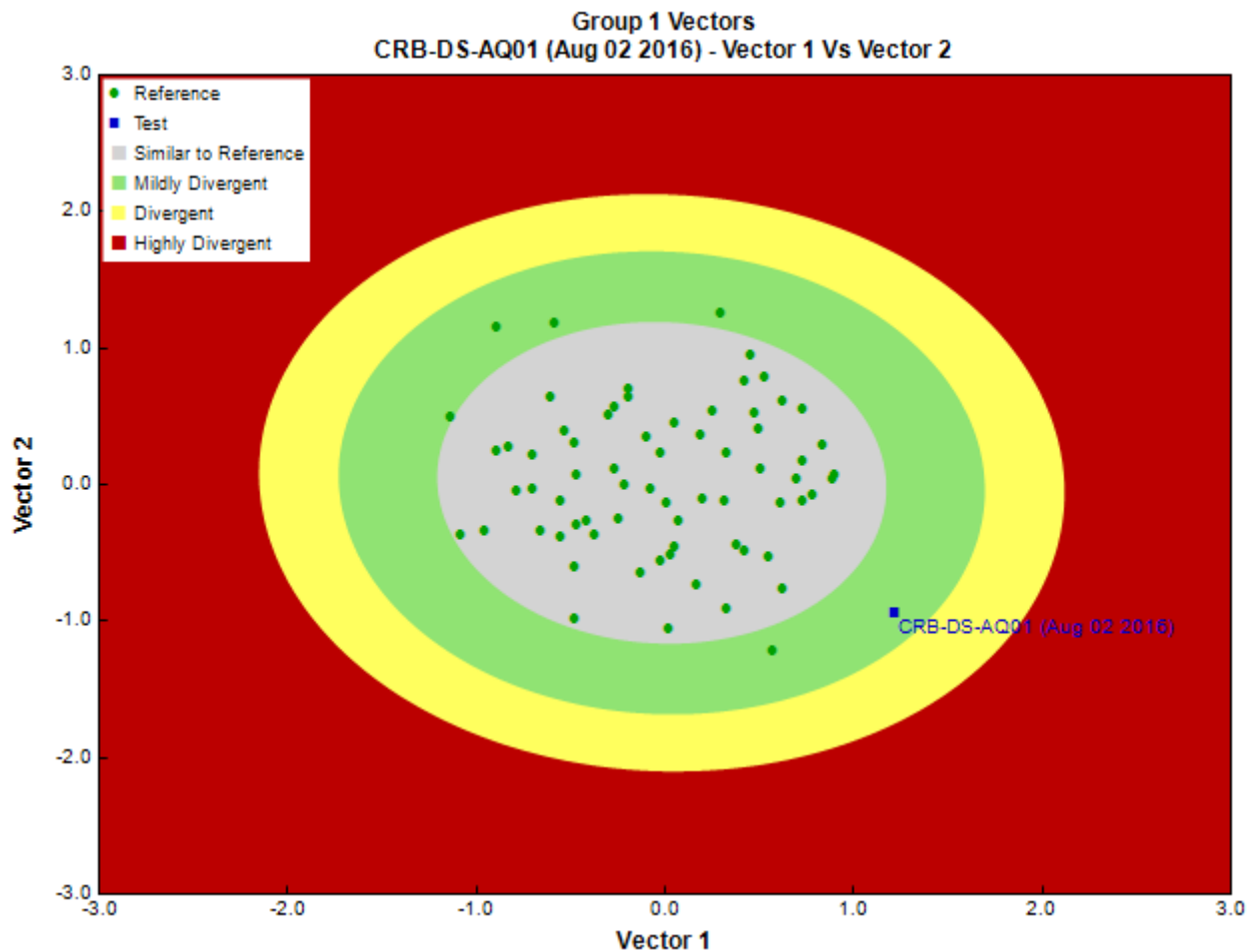


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | September 27, 2016 |
| | Marchant Box |
| Sub-Sample Proportion | 10/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 10 | 100.0 |
| | Collembola | Collembola | | 1 | 10.0 |
| | Insecta | Diptera | Chironomidae | 18 | 180.0 |
| | | | Empididae | 5 | 50.0 |
| | | | Simuliidae | 17 | 170.0 |
| | | | Tipulidae | 2 | 20.0 |
| | | Ephemeroptera | Baetidae | 41 | 410.0 |
| | | | Ephemerellidae | 5 | 50.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | Megaloptera | Sialidae | 1 | 10.0 |
| | | Plecoptera | Chloroperlidae | 55 | 550.0 |
| | | | Leuctridae | 1 | 10.0 |
| | | | Nemouridae | 159 | 1,590.0 |
| | | | Perlodidae | 1 | 10.0 |
| | | Trichoptera | Limnephilidae | 1 | 10.0 |
| | | | Rhyacophilidae | 1 | 10.0 |
| | | | Total | 318 | 3,180.0 |

Metrics

| Name | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.71 | 0.5 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.8 \pm 2.9 |
| % Gatherers | 63.2 | 43.7 \pm 17.3 |
| % Predatores | 16.7 | 20.9 \pm 13.4 |
| % Scrapers | 18.6 | 54.8 \pm 18.3 |
| % Shredder | 51.3 | 21.3 \pm 13.9 |
| No. Clinger Taxa | 9.0 | 16.0 \pm 5.6 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 16.4 | 19.1 \pm 14.0 |
| % EPT Individuals | 83.3 | 79.0 \pm 14.8 |
| % of 5 dominant taxa | 91.5 | 83.1 \pm 9.3 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.9 \pm 0.1 |
| Total Abundance | 3180.0 | 5010.8 \pm 6541.9 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.4 \pm 0.5 |
| Diptera taxa | 4.0 | 3.1 \pm 1.3 |
| Ephemeroptera taxa | 2.0 | 3.6 \pm 1.0 |
| EPT Individuals (Sum) | 2640.0 | 3855.4 \pm 5103.0 |
| EPT taxa (no) | 8.0 | 11.0 \pm 2.8 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.7 \pm 0.1 |
| Plecoptera taxa | 4.0 | 4.5 \pm 1.3 |
| Shannon-Wiener Diversity | 1.6 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.7 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 14.0 | 16.8 \pm 4.7 |
| Trichoptera taxa | 2.0 | 2.9 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.30 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.03 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.02 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.05 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.92 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.41 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.03 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.68 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.26 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.76 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.03 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.47 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.53 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.26 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.85 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.05 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.26 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.91 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.01 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.54 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.24 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.09 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.06 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.41 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.42 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.11 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.51 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.31 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.28 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.13 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.12 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.03 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.02 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.41 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.78 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.02 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.02 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.29 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.62 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.02 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.28 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.04 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.24 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.48 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.38 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.29 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.43 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.60 |
| Tormenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.38 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.03 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.09 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 8.70 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 0.92 |
| RIVPACS : Expected taxa P>0.70 | 5.21 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.96 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 0.33000 | 25.53190 \pm 36.87363 |
| Channel | | |
| Depth-Avg (cm) | 8.6 | 20.6 \pm 10.3 |
| Depth-BankfullMinusWetted (cm) | 58.00 | 37.41 \pm 19.51 |
| Depth-Max (cm) | 12.5 | 30.1 \pm 17.1 |
| Macrophyte (PercentRange) | 0 | 0 \pm 1 |
| Reach-%CanopyCoverage (PercentRange) | 2.00 | 0.86 \pm 1.10 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 \pm 1 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0139981 \pm 0.0172321 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.33 | 0.39 \pm 0.19 |
| Velocity-Max (m/s) | 0.62 | 0.58 \pm 0.28 |
| Width-Bankfull (m) | 5.2 | 18.4 \pm 20.0 |
| Width-Wetted (m) | 3.0 | 7.4 \pm 6.3 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 \pm 0 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 \pm 0 |
| Climate | | |
| Precip02_FEB (mm) | 163.00000 | 57.95789 \pm 41.70288 |
| Temp07_JULmax (Degrees Celsius) | 20.48000 | 17.55944 \pm 2.17158 |
| Landcover | | |
| Natl-SnowIce (%) | 0.00000 | 0.55339 \pm 1.25503 |
| Natl-Water (%) | 0.00000 | 1.38766 \pm 2.38578 |
| Natl-WetlandHerb (%) | 0.00000 | 0.46466 \pm 1.02141 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 1 \pm 3 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| %Boulder (%) | 1 | 6 \pm 5 |
| %Cobble (%) | 68 | 52 \pm 17 |
| %Gravel (%) | 6 | 6 \pm 5 |
| %Pebble (%) | 25 | 34 \pm 16 |
| %Sand (%) | 0 | 0 \pm 1 |
| %Silt+Clay (%) | 0 | 1 \pm 2 |
| D50 (cm) | 8.00 | 7.79 \pm 2.83 |
| Dg (cm) | 6.7 | 7.0 \pm 2.2 |
| Dominant-1st (Category(0-9)) | 6 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 7 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 26.12000 | 19.23143 \pm 15.15733 |
| Water Chemistry | | |
| General-DO (mg/L) | 9.3500000 | 11.6403031 \pm 1.0007120 |
| General-pH (pH) | 7.6 | 7.6 \pm 0.5 |
| General-SpCond (μ S/cm) | 217.8000000 | 127.8461538 \pm 102.3985239 |
| General-TempAir (Degrees Celsius) | 12.2 | 11.6 \pm 4.1 |
| General-TempWater (Degrees Celsius) | 12.7000000 | 5.9833333 \pm 2.8160802 |
| General-Turbidity (NTU) | 1.5500000 | 0.5285714 \pm 0.3093773 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | CRB-DS-AQ01 |
| Sampling Date | Jul 25 2017 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12639 N, 122.97167 W |
| Altitude | 643 |
| Local Basin Name | Crabapple Creek |
| | River of Golden Dreams |
| Stream Order | 2 |

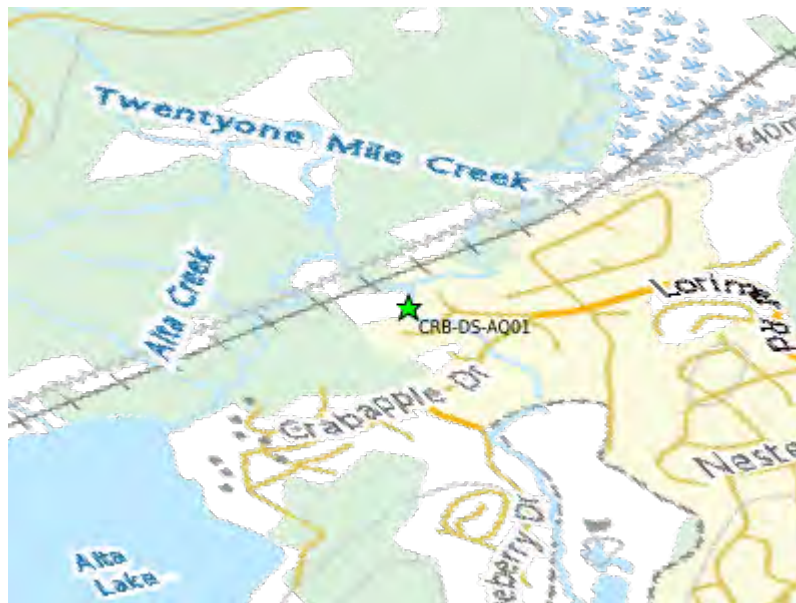


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 43.7% | 26.6% | 0.1% | 19.1% | 8.5% | 2.1% |
| CABIN Assessment of CRB-DS-AQ01 on Jul 25, 2017 | Similar to Reference | | | | | |

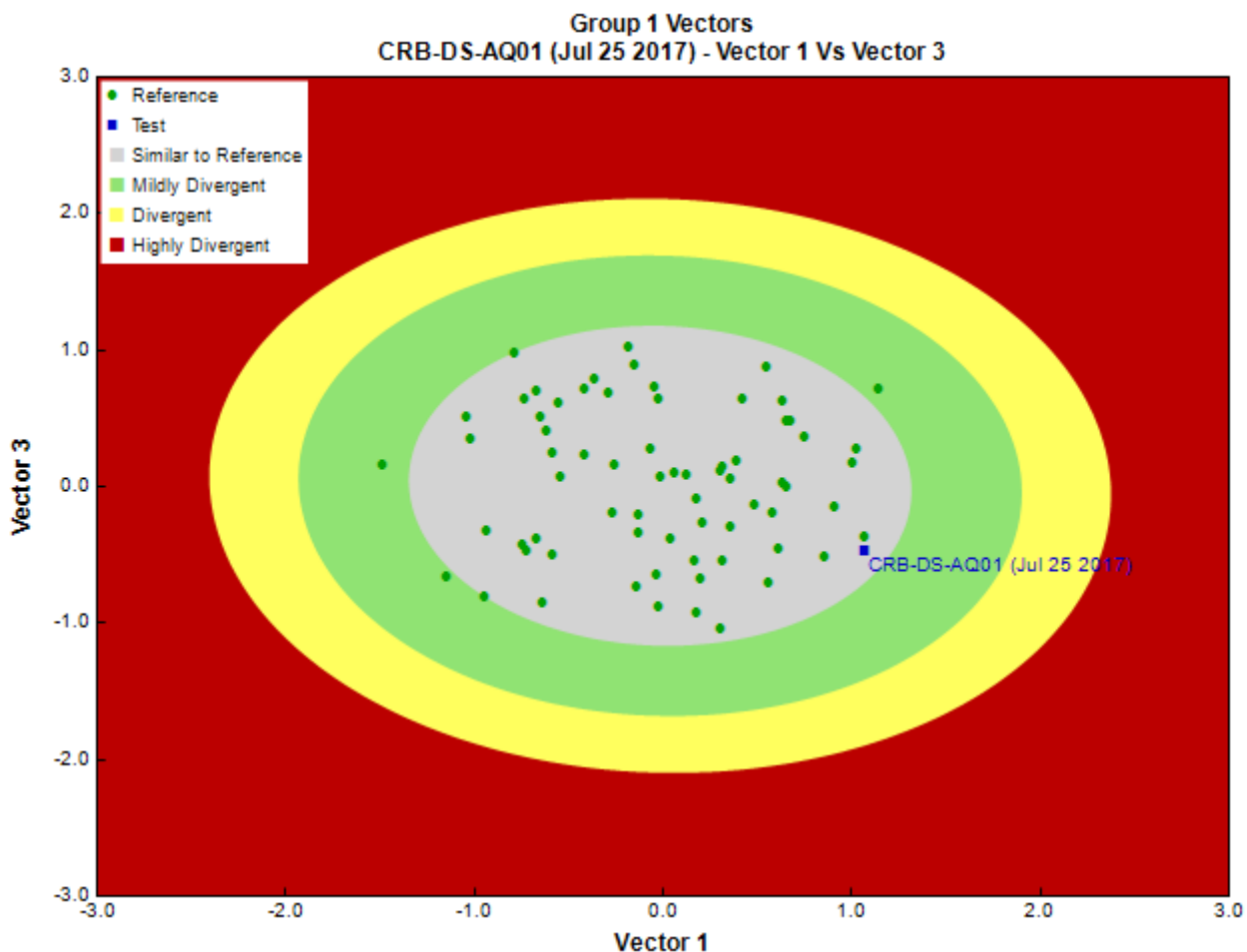


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | November 01, 2017 |
| | Marchant Box |
| Sub-Sample Proportion | 16/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 5 | 31.3 |
| | Insecta | Diptera | Chironomidae | 34 | 212.5 |
| | | | Simuliidae | 15 | 93.8 |
| | | Ephemeroptera | Baetidae | 302 | 1,887.5 |
| | | | Ephemerellidae | 9 | 56.3 |
| | | | Leptophlebiidae | 1 | 6.3 |
| | | Plecoptera | Chloroperlidae | 12 | 75.0 |
| | | | Nemouridae | 25 | 156.3 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | Trichoptera | Limnephilidae | 1 | 6.3 |
| | | | Rhyacophilidae | 8 | 50.0 |
| | | | Total | 412 | 2,575.3 |

Metrics

| Name | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.37 | 0.5 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.8 \pm 2.9 |
| % Gatherers | 20.4 | 43.7 \pm 17.3 |
| % Predatores | 15.0 | 20.9 \pm 13.4 |
| % Scrapers | 77.2 | 54.8 \pm 18.3 |
| % Shredder | 6.3 | 21.3 \pm 13.9 |
| No. Clinger Taxa | 8.0 | 16.0 \pm 5.6 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 13.1 | 19.1 \pm 14.0 |
| % EPT Individuals | 86.9 | 79.0 \pm 14.8 |
| % of 5 dominant taxa | 94.2 | 83.1 \pm 9.3 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.9 \pm 0.1 |
| Total Abundance | 2575.0 | 5010.8 \pm 6541.9 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.4 \pm 0.5 |
| Diptera taxa | 2.0 | 3.1 \pm 1.3 |
| Ephemeroptera taxa | 3.0 | 3.6 \pm 1.0 |
| EPT Individuals (Sum) | 2237.5 | 3855.4 \pm 5103.0 |
| EPT taxa (no) | 7.0 | 11.0 \pm 2.8 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.5 | 0.7 \pm 0.1 |
| Plecoptera taxa | 2.0 | 4.5 \pm 1.3 |
| Shannon-Wiener Diversity | 1.1 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.4 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 10.0 | 16.8 \pm 4.7 |
| Trichoptera taxa | 2.0 | 2.9 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.30 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.03 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.02 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.05 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.92 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.41 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.03 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.68 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.26 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.76 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.03 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.47 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.53 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.26 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.85 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.05 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.26 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.91 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.01 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.54 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.24 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.09 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.06 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.41 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.42 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.11 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.51 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.31 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.28 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.13 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.12 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.03 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.02 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.41 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.78 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.02 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.02 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.29 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.62 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.02 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.28 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.04 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.24 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.48 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.38 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.29 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.43 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.60 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.38 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.03 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.09 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 8.70 |
| RIVPACS : Observed taxa P>0.50 | 6.00 |
| RIVPACS : O:E (p > 0.5) | 0.69 |
| RIVPACS : Expected taxa P>0.70 | 5.21 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.96 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 0.33000 | 25.53190 \pm 36.87363 |
| Channel | | |
| Depth-Avg (cm) | 7.0 | 20.6 \pm 10.3 |
| Depth-BankfullMinusWetted (cm) | 30.00 | 37.41 \pm 19.51 |
| Depth-Max (cm) | 10.0 | 30.1 \pm 17.1 |
| Macrophyte (PercentRange) | 0 | 0 \pm 1 |
| Reach-%CanopyCoverage (PercentRange) | 3.00 | 0.86 \pm 1.10 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 2 \pm 1 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0139981 \pm 0.0172321 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.30 | 0.39 \pm 0.19 |
| Velocity-Max (m/s) | 0.40 | 0.58 \pm 0.28 |
| Width-Bankfull (m) | 4.4 | 18.4 \pm 20.0 |
| Width-Wetted (m) | 3.3 | 7.4 \pm 6.3 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 \pm 0 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 \pm 0 |
| Climate | | |
| Precip02_FEB (mm) | 163.00000 | 57.95789 \pm 41.70288 |
| Temp07_JULmax (Degrees Celsius) | 20.48000 | 17.55944 \pm 2.17158 |
| Landcover | | |
| Natl-SnowIce (%) | 0.00000 | 0.55339 \pm 1.25503 |
| Natl-Water (%) | 0.00000 | 1.38766 \pm 2.38578 |
| Natl-WetlandHerb (%) | 0.00000 | 0.46466 \pm 1.02141 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 1 \pm 3 |
| %Boulder (%) | 0 | 6 \pm 5 |
| %Cobble (%) | 47 | 52 \pm 17 |
| %Gravel (%) | 11 | 6 \pm 5 |
| %Pebble (%) | 39 | 34 \pm 16 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pmSD |
|--|--------------------|--|
| %Sand (%) | 0 | 0 \pm 1 |
| %Silt+Clay (%) | 0 | 1 \pm 2 |
| D50 (cm) | 6.00 | 7.79 \pm 2.83 |
| Dg (cm) | 4.6 | 7.0 \pm 2.2 |
| Dominant-1st (Category(0-9)) | 6 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 5 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 26.12000 | 19.23143 \pm 15.15733 |
| Water Chemistry | | |
| General-DO (mg/L) | 11.6000000 | 11.6403031 \pm 1.0007120 |
| General-pH (pH) | 7.4 | 7.6 \pm 0.5 |
| General-SpCond (μS/cm) | 336.3000000 | 127.8461538 \pm 102.3985239 |
| General-TempAir (Degrees Celsius) | 17.5 | 11.6 \pm 4.1 |
| General-TempWater (Degrees Celsius) | 12.0000000 | 5.9833333 \pm 2.8160802 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | CRB-DS-AQ01 |
| Sampling Date | Aug 01 2018 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12648 N, 122.97171 W |
| Altitude | 645 |
| Local Basin Name | Crabapple Creek |
| | River of Golden Dreams |
| Stream Order | 2 |

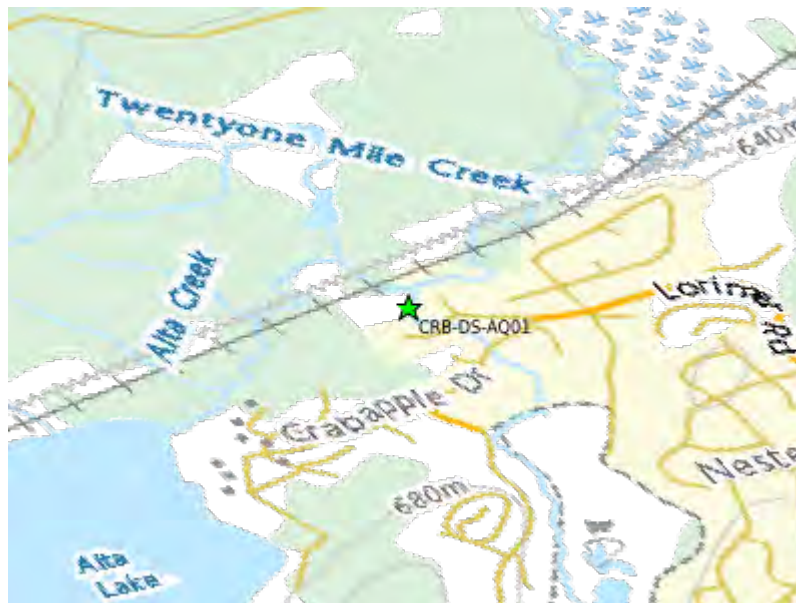


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 43.7% | 26.5% | 0.1% | 19.1% | 8.5% | 2.1% |
| CABIN Assessment of CRB-DS-AQ01 on Aug 01, 2018 | Similar to Reference | | | | | |

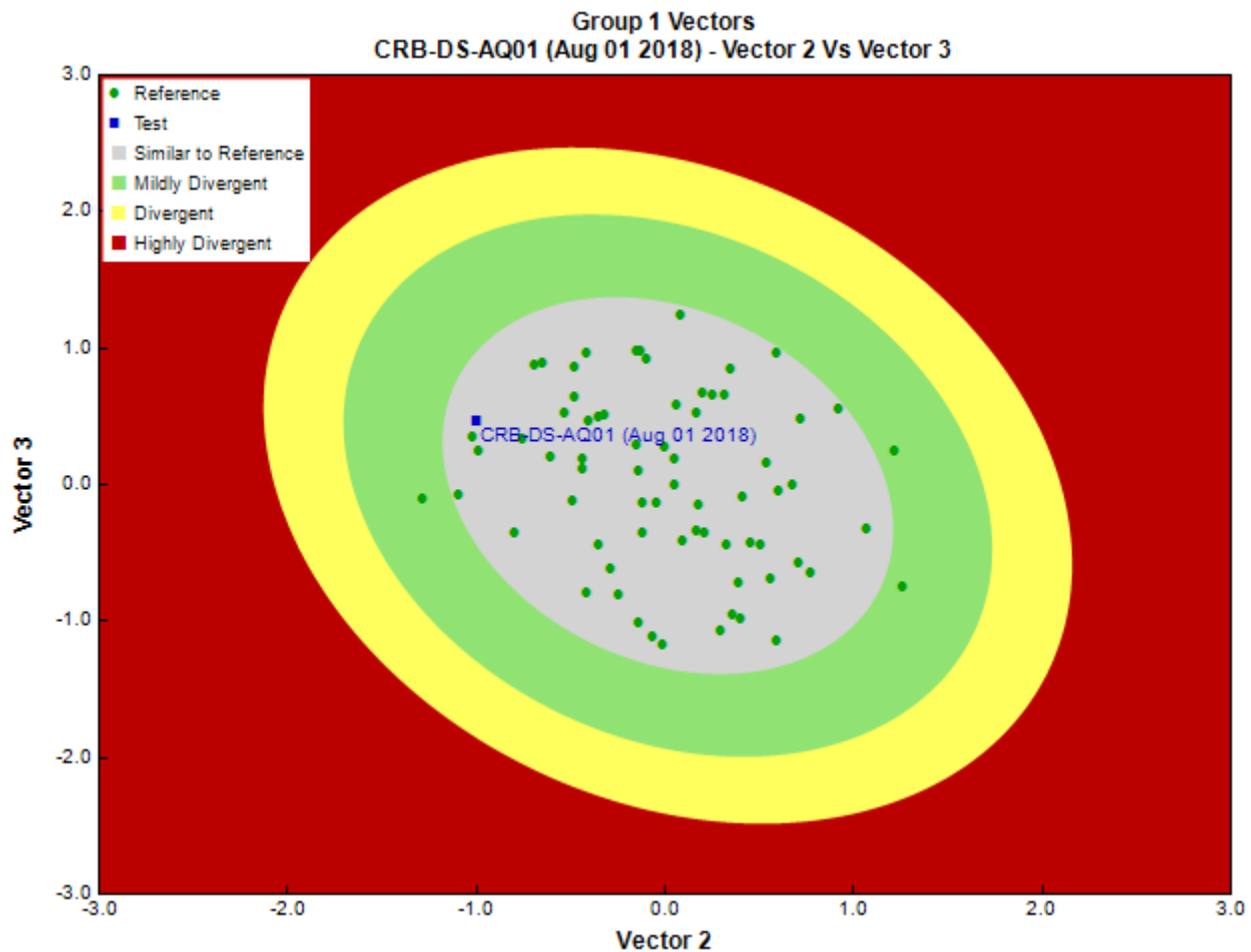


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 10/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|-----------------|-----------|-------------|
| Annelida | Clitellata | Tubificida | Naididae | 1 | 10.0 |
| Arthropoda | Arachnida | Trombidiformes | Aturidae | 1 | 10.0 |
| | | | Hydryphantidae | 1 | 10.0 |
| | | | Hygrobatidae | 3 | 30.0 |
| | | | Lebertiidae | 1 | 10.0 |
| | | | Sperchontidae | 1 | 10.0 |
| | | | Torrenticolidae | 1 | 10.0 |
| | Insecta | Diptera | Chironomidae | 44 | 440.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|---------------|-----------------|-----------|-------------|
| | | | Empididae | 4 | 40.0 |
| | | | Simuliidae | 14 | 140.0 |
| | | | Tipulidae | 4 | 40.0 |
| | | Ephemeroptera | Baetidae | 116 | 1,160.0 |
| | | | Ephemerellidae | 1 | 10.0 |
| | | | Heptageniidae | 4 | 40.0 |
| | | | Leptophlebiidae | 24 | 240.0 |
| | | Plecoptera | Chloroperlidae | 10 | 100.0 |
| | | | Nemouridae | 78 | 780.0 |
| | | | Perlodidae | 1 | 10.0 |
| | | Trichoptera | Limnephilidae | 5 | 50.0 |
| | | | Rhyacophilidae | 5 | 50.0 |
| | | | Total | 319 | 3,190.0 |

Metrics

| Name | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.43 | 0.5 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.8 \pm 2.9 |
| % Gatherers | 52.4 | 43.7 \pm 17.3 |
| % Predatores | 23.8 | 20.9 \pm 13.4 |
| % Scrapers | 43.6 | 54.8 \pm 18.3 |
| % Shredder | 27.3 | 21.3 \pm 13.9 |
| No. Clinger Taxa | 16.0 | 16.0 \pm 5.6 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 23.5 | 19.1 \pm 14.0 |
| % EPT Individuals | 76.5 | 79.0 \pm 14.8 |
| % of 5 dominant taxa | 86.5 | 83.1 \pm 9.3 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.8 | 0.9 \pm 0.1 |
| Total Abundance | 3190.0 | 5010.8 \pm 6541.9 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.4 \pm 0.5 |
| Diptera taxa | 4.0 | 3.1 \pm 1.3 |
| Ephemeroptera taxa | 4.0 | 3.6 \pm 1.0 |
| EPT Individuals (Sum) | 2440.0 | 3855.4 \pm 5103.0 |
| EPT taxa (no) | 9.0 | 11.0 \pm 2.8 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.5 \pm 1.3 |
| Shannon-Wiener Diversity | 1.9 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 20.0 | 16.8 \pm 4.7 |
| Trichoptera taxa | 2.0 | 2.9 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.30 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.03 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.02 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.05 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.92 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.41 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.03 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.68 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.26 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.76 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.03 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.47 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.53 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.26 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.85 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.05 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.26 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.91 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.01 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.54 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.24 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.09 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.06 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.41 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.42 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.11 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.51 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.31 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.28 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.13 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.12 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.03 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.02 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.41 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.78 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.02 |
| Pelecoryhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.02 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.29 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.62 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.02 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.28 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.04 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.24 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.48 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.38 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.29 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.43 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.60 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.38 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.03 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.09 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|-------|
| RIVPACS : Expected taxa P>0.50 | 8.70 |
| RIVPACS : Observed taxa P>0.50 | 10.00 |
| RIVPACS : O:E (p > 0.5) | 1.15 |
| RIVPACS : Expected taxa P>0.70 | 5.21 |
| RIVPACS : Observed taxa P>0.70 | 6.00 |
| RIVPACS : O:E (p > 0.7) | 1.15 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 0.33000 | 25.53190 \pm 36.87363 |
| Channel | | |
| Depth-Avg (cm) | 14.2 | 20.6 \pm 10.3 |
| Depth-BankfullMinusWetted (cm) | 37.00 | 37.41 \pm 19.51 |
| Macrophyte (PercentRange) | 0 | 0 \pm 1 |
| Reach-%CanopyCoverage (PercentRange) | 3.00 | 0.86 \pm 1.10 |
| Reach-%Logging (PercentRange) | 0 | 0 \pm 0 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 \pm 1 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0139981 \pm 0.0172321 |
| Veg-Coniferous (Binary) | 0 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.34 | 0.39 \pm 0.19 |
| Velocity-Max (m/s) | 0.44 | 0.58 \pm 0.28 |
| Width-Bankfull (m) | 4.5 | 18.4 \pm 20.0 |
| Width-Wetted (m) | 3.1 | 7.4 \pm 6.3 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 \pm 0 |
| Climate | | |
| Precip02_FEB (mm) | 163.00000 | 57.95789 \pm 41.70288 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Temp07_JULmax (Degrees Celsius) | 20.48000 | 17.55944 \pm 2.17158 |
| Landcover | | |
| Natl-SnowIce (%) | 0.00000 | 0.55339 \pm 1.25503 |
| Natl-Water (%) | 0.00000 | 1.38766 \pm 2.38578 |
| Natl-WetlandHerb (%) | 0.00000 | 0.46466 \pm 1.02141 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 1 \pm 3 |
| %Boulder (%) | 1 | 6 \pm 5 |
| %Cobble (%) | 50 | 52 \pm 17 |
| %Gravel (%) | 18 | 6 \pm 5 |
| %Pebble (%) | 26 | 34 \pm 16 |
| %Sand (%) | 0 | 0 \pm 1 |
| %Silt+Clay (%) | 5 | 1 \pm 2 |
| D50 (cm) | 6.80 | 7.79 \pm 2.83 |
| Dg (cm) | 3.9 | 7.0 \pm 2.2 |
| Dominant-1st (Category(0-9)) | 6 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 3 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 3 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 26.12000 | 19.23143 \pm 15.15733 |
| Water Chemistry | | |
| General-DO (mg/L) | 7.5300000 | 11.6403031 \pm 1.0007120 |
| General-pH (pH) | 7.5 | 7.6 \pm 0.5 |
| General-SpCond (μ S/cm) | 194.4000000 | 127.8461538 \pm 102.3985239 |
| General-TempAir (Degrees Celsius) | 19.0 | 11.6 \pm 4.1 |
| General-TempWater (Degrees Celsius) | 16.0000000 | 5.9833333 \pm 2.8160802 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | CRB-DS-AQ01 |
| Sampling Date | Jul 30 2019 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12654 N, 122.97168 W |
| Altitude | 656 |
| Local Basin Name | Crabapple Creek |
| | River of Golden Dreams |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 0.0% | 0.0% | 0.0% | 16.5% | 82.5% | 1.0% |
| CABIN Assessment of CRB-DS-AQ01 on Jul 30, 2019 | Mildly Divergent | | | | | |

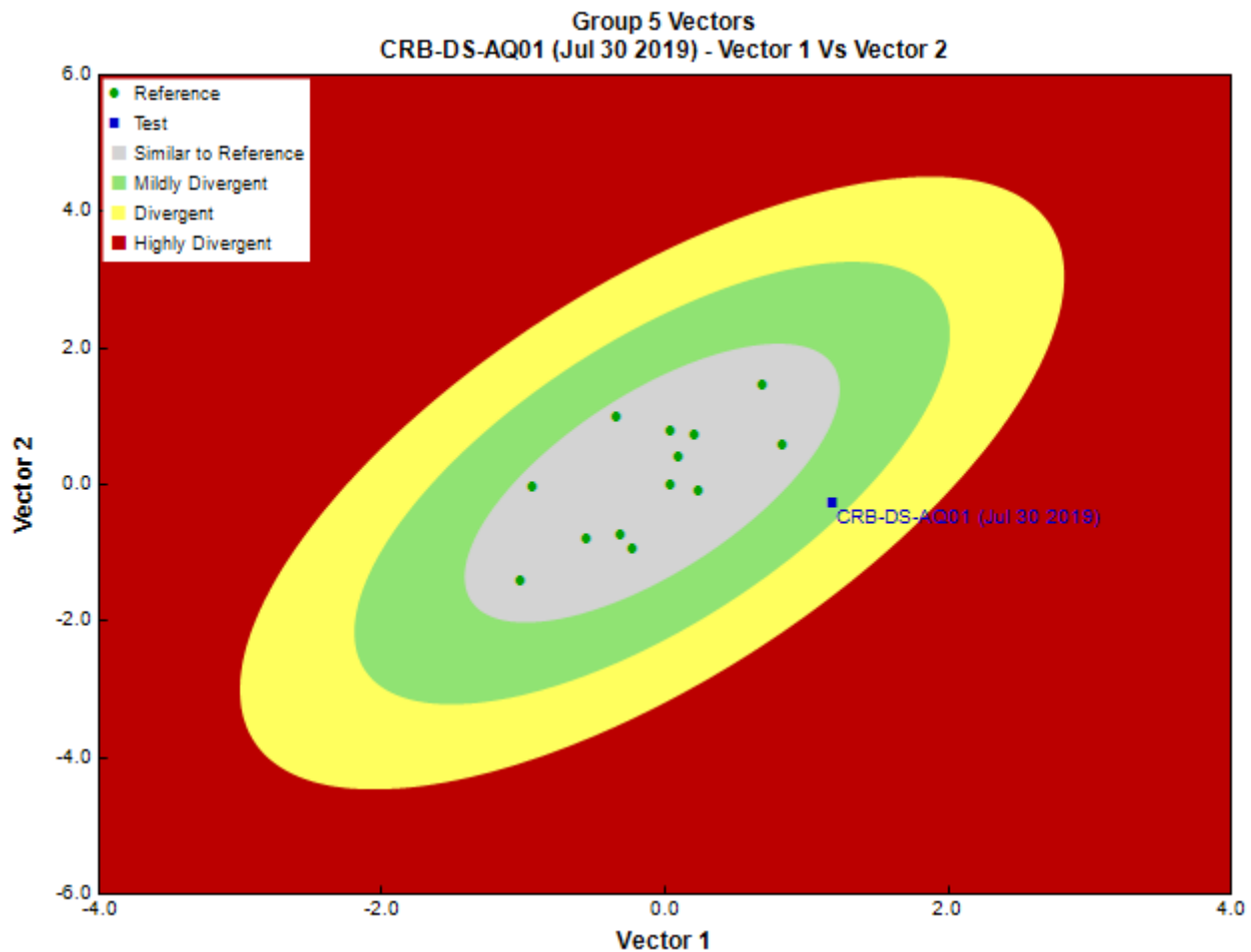


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 10/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|--------------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hygrobatidae | 3 | 30.0 |
| | | | Sperchontidae | 2 | 20.0 |
| | | | Stygothrombidiidae | 1 | 10.0 |
| | Collembola | Collembola | | 1 | 10.0 |
| | Insecta | Diptera | Chironomidae | 63 | 630.0 |
| | | | Empididae | 1 | 10.0 |
| | | | Simuliidae | 13 | 130.0 |
| | | | Tipulidae | 1 | 10.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|----------|---------------|-----------------|-----------|-------------|
| | | Ephemeroptera | Baetidae | 161 | 1,610.0 |
| | | | Ephemerellidae | 1 | 10.0 |
| | | | Heptageniidae | 1 | 10.0 |
| | | | Leptophlebiidae | 14 | 140.0 |
| | | Plecoptera | Chloroperlidae | 5 | 50.0 |
| | | | Nemouridae | 77 | 770.0 |
| | | Trichoptera | | 1 | 10.0 |
| | | | Rhyacophilidae | 4 | 40.0 |
| Mollusca | Bivalvia | Veneroida | Pisidiidae | 1 | 10.0 |
| | | | Total | 350 | 3,500.0 |

Metrics

| Name | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.72 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 48.6 | 67.6 \pm 30.3 |
| % Predatores | 24.9 | 41.1 \pm 20.2 |
| % Scrapers | 50.0 | 34.3 \pm 21.0 |
| % Shredder | 22.3 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 15.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 24.4 | 47.4 \pm 26.3 |
| % EPT Individuals | 75.6 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 94.3 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.8 | 0.6 \pm 0.2 |
| Total Abundance | 3500.0 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.5 \pm 0.5 |
| Diptera taxa | 4.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 4.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 2630.0 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 7.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.6 \pm 0.1 |
| Plecoptera taxa | 2.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 1.5 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.7 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 15.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 1.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.08 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.00 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.08 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.13 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.08 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.01 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.67 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.00 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.19 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.32 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.37 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.49 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.41 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.06 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.00 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.18 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.17 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.31 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.67 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.00 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.00 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.13 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.06 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.62 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.31 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.22 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.01 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.02 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.08 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.52 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.27 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.52 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.00 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.19 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.13 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.05 |
| Limnésiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.40 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.41 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.07 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.00 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.78 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.27 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.13 |
| Pelecoryhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.01 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.12 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.39 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.00 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.00 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.46 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.26 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at CRB-DS-AQ01 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.04 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.14 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.11 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.12 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.28 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.07 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.22 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.00 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.44 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.25 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.03 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.13 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.07 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 4.77 |
| RIVPACS : Observed taxa P>0.50 | 4.00 |
| RIVPACS : O:E (p > 0.5) | 0.84 |
| RIVPACS : Expected taxa P>0.70 | 1.78 |
| RIVPACS : Observed taxa P>0.70 | 1.00 |
| RIVPACS : O:E (p > 0.7) | 0.56 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 0.33000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 12.8 | 40.5 \pm 22.4 |
| Depth-BankfullMinusWetted (cm) | 35.00 | 188.00 |
| Depth-Max (cm) | 17.5 | 55.5 \pm 31.7 |
| Macrophyte (PercentRange) | 0 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 4.00 | 0.23 \pm 0.44 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 1 \pm 0 |
| Slope (m/m) | 3.0000000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 0 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.56 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 0.70 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 3.4 | 75.1 \pm 72.8 |
| Width-Wetted (m) | 2.3 | 50.6 \pm 60.4 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 163.00000 | 171.50745 \pm 107.47690 |
| Temp07_JULmax (Degrees Celsius) | 20.48000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 0.00000 | 3.62533 \pm 10.17162 |

Habitat Description

| Variable | CRB-DS-AQ01 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Natl-Water (%) | 0.00000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 1 | 0 |
| %Cobble (%) | 58 | 58 |
| %Gravel (%) | 2 | 1 |
| %Pebble (%) | 31 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 8 | 0 |
| D50 (cm) | 7.10 | 3.30 |
| Dg (cm) | 5.1 | 6.6 |
| Dominant-1st (Category(0-9)) | 6 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 7 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 3 |
| SurroundingMaterial (Category(0-9)) | 2 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 26.12000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-Conductivity (μ S/cm) | 184.9000000 | 79.0846153 \pm 50.3407694 |
| General-DO (mg/L) | 10.0000000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 7.6 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 234.9000000 | 176.1000000 |
| General-TempAir (Degrees Celsius) | 13.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 13.9000000 | 13.2730769 \pm 4.7663725 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | JOR-DS-AQ31 |
| Sampling Date | Aug 03 2016 |
| Know Your Watershed Basin | Strait of Georgia - East Shore |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.09545 N, 122.99735 W |
| Altitude | 623 |
| Local Basin Name | Jordan Creek |
| | Jordan Creek |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 1 2 3 4 5 6 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 13.7% | 8.3% | 0.2% | 55.1% | 2.4% | 20.3% |
| CABIN Assessment of JOR-DS-AQ31 on Aug 03, 2016 | Mildly Divergent | | | | | |

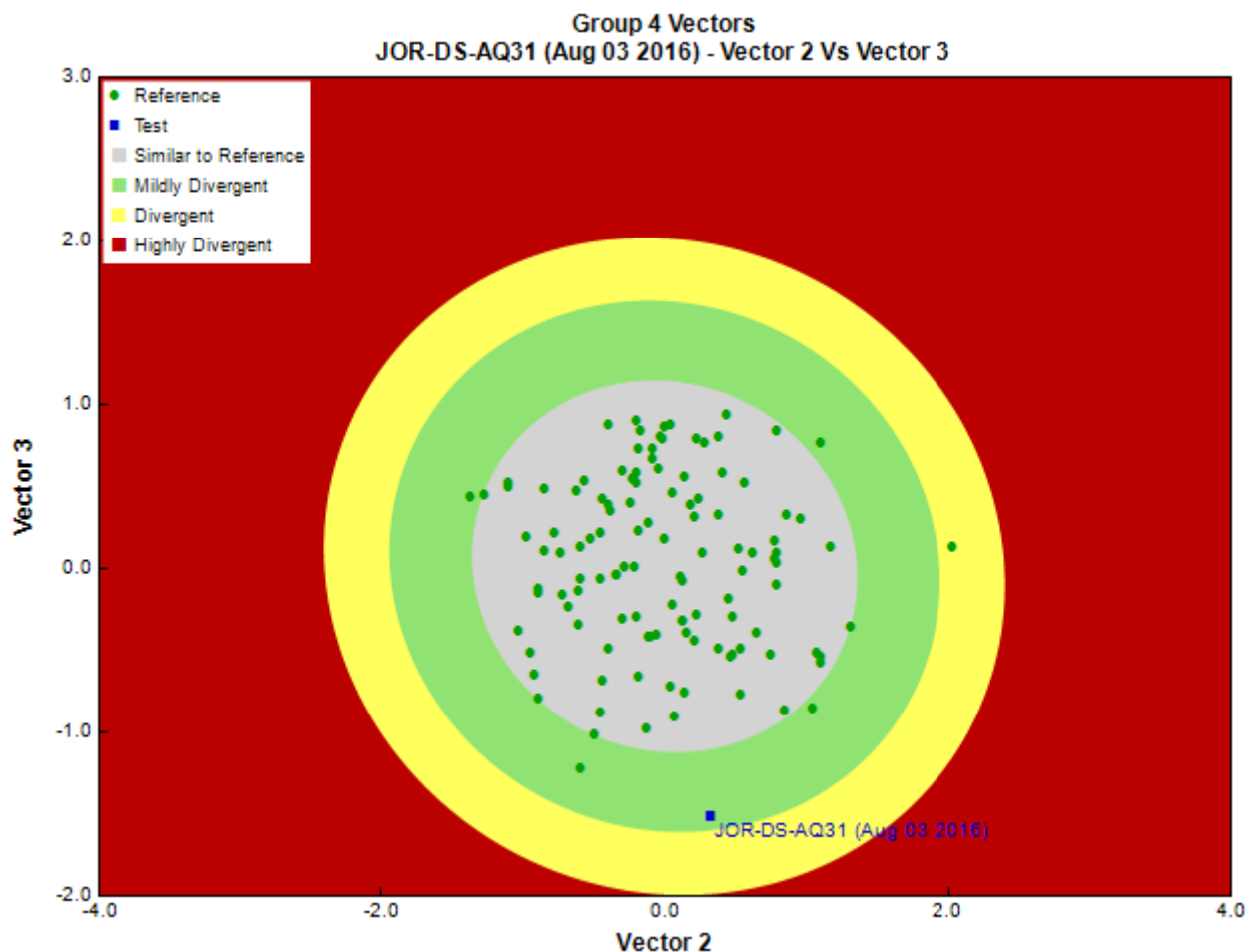


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | October 03, 2016 |
| | Marchant Box |
| Sub-Sample Proportion | 16/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 4 | 25.0 |
| | Insecta | Diptera | Ceratopogonidae | 1 | 6.3 |
| | | | Chironomidae | 43 | 268.8 |
| | | | Empididae | 2 | 12.5 |
| | | | Simuliidae | 116 | 725.0 |
| | | Ephemeroptera | Baetidae | 9 | 56.3 |
| | | | Ephemerellidae | 3 | 18.8 |
| | | Plecoptera | Chloroperlidae | 1 | 6.3 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | | Nemouridae | 145 | 906.3 |
| | | | Perlidae | 5 | 31.3 |
| | | Trichoptera | Hydropsychidae | 5 | 31.3 |
| | | | Rhyacophilidae | 1 | 6.3 |
| | | | Total | 335 | 2,094.2 |

Metrics

| Name | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|---|-------------|--|
| Bray-Curtis Distance | 0.78 | 0.5 \pm 0.1 |
| Functional Measures | | |
| % Filterers | -- | 17.2 \pm 42.4 |
| % Gatherers | 93.1 | 57.6 \pm 27.3 |
| % Predatores | 52.5 | 31.3 \pm 20.3 |
| % Scrapers | 38.8 | 37.4 \pm 22.0 |
| % Shredder | 43.3 | 16.1 \pm 11.0 |
| No. Clinger Taxa | 8.0 | 15.8 \pm 6.3 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 49.6 | 32.8 \pm 26.0 |
| % EPT Individuals | 50.4 | 66.1 \pm 26.2 |
| % of 5 dominant taxa | 94.9 | 82.2 \pm 8.7 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.8 | 0.7 \pm 0.2 |
| Total Abundance | 2093.8 | 2646.7 \pm 2772.7 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.1 |
| Coleoptera taxa | 0.0 | 0.3 \pm 0.4 |
| Diptera taxa | 4.0 | 3.2 \pm 1.3 |
| Ephemeroptera taxa | 2.0 | 3.6 \pm 1.1 |
| EPT Individuals (Sum) | 1056.3 | 1501.0 \pm 1294.6 |
| EPT taxa (no) | 7.0 | 10.8 \pm 3.5 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.1 \pm 1.8 |
| Shannon-Wiener Diversity | 1.4 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.7 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 12.0 | 18.0 \pm 4.5 |
| Trichoptera taxa | 2.0 | 3.1 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|---|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.46 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.02 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.09 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.00 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.06 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.04 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.91 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.02 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.35 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.01 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.67 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.24 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.84 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.01 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.02 |
| Dugesiidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.32 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.57 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.33 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.90 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.02 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.02 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.27 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.95 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.61 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.16 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.10 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.08 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.04 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.43 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.39 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.06 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.38 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.34 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.29 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.14 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.15 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.04 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.42 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.79 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.01 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.05 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.28 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.72 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.02 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.16 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.02 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.19 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.60 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.32 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.43 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.03 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.53 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.03 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.65 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.31 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.02 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.13 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 9.71 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 0.82 |
| RIVPACS : Expected taxa P>0.70 | 6.09 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.82 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 60.51000 | 28.74839 \pm 35.48825 |
| Channel | | |
| Depth-Avg (cm) | 18.5 | 28.0 \pm 13.9 |
| Depth-BankfullMinusWetted (cm) | 74.00 | 47.88 \pm 26.69 |
| Depth-Max (cm) | 24.0 | 41.3 \pm 21.8 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 3.00 | 0.92 \pm 1.11 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 3 \pm 1 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0300000 | 0.0249850 \pm 0.0294369 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 0 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 1 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.40 | 0.45 \pm 0.21 |
| Velocity-Max (m/s) | 0.77 | 0.67 \pm 0.25 |
| Width-Bankfull (m) | 7.1 | 35.9 \pm 41.6 |
| Width-Wetted (m) | 4.2 | 17.8 \pm 20.2 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 \pm 0 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 \pm 1 |
| Climate | | |
| Precip02_FEB (mm) | 163.27000 | 94.95103 \pm 61.64910 |
| Temp07_JULmax (Degrees Celsius) | 18.72000 | 17.48320 \pm 2.57900 |
| Landcover | | |
| Natl-SnowIce (%) | 3.08000 | 4.62982 \pm 9.77010 |
| Natl-Water (%) | 1.45000 | 1.55060 \pm 2.36345 |
| Natl-WetlandHerb (%) | 0.00000 | 0.18446 \pm 0.50703 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 1 |
| %Boulder (%) | 15 | 11 \pm 11 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| %Cobble (%) | 52 | 53 \pm 11 |
| %Gravel (%) | 8 | 5 \pm 4 |
| %Pebble (%) | 25 | 30 \pm 12 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 1 \pm 3 |
| D50 (cm) | 12.00 | 8.04 \pm 4.60 |
| Dg (cm) | 9.2 | 8.2 \pm 3.1 |
| Dominant-1st (Category(0-9)) | 7 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 6 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 40.36000 | 31.09165 \pm 12.51836 |
| Water Chemistry | | |
| General-DO (mg/L) | 9.3200000 | 11.4180702 \pm 1.2821697 |
| General-pH (pH) | 7.1 | 7.7 \pm 0.7 |
| General-SpCond (μ S/cm) | 63.6000000 | 105.8321429 \pm 89.5097928 |
| General-TempAir (Degrees Celsius) | 17.4 | 12.1 \pm 4.3 |
| General-TempWater (Degrees Celsius) | 15.8000000 | 7.6535897 \pm 3.4680513 |
| General-Turbidity (NTU) | 0.6300000 | 0.5500000 \pm 0.6138116 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | JOR-DS-AQ31 |
| Sampling Date | Jul 26 2017 |
| Know Your Watershed Basin | Strait of Georgia - East Shore |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.09528 N, 122.99778 W |
| Altitude | 602 |
| Local Basin Name | Jordan Creek |
| | Jordan Creek |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 18.0% | 9.7% | 0.2% | 51.5% | 2.1% | 18.5% |
| CABIN Assessment of JOR-DS-AQ31 on Jul 26, 2017 | Mildly Divergent | | | | | |

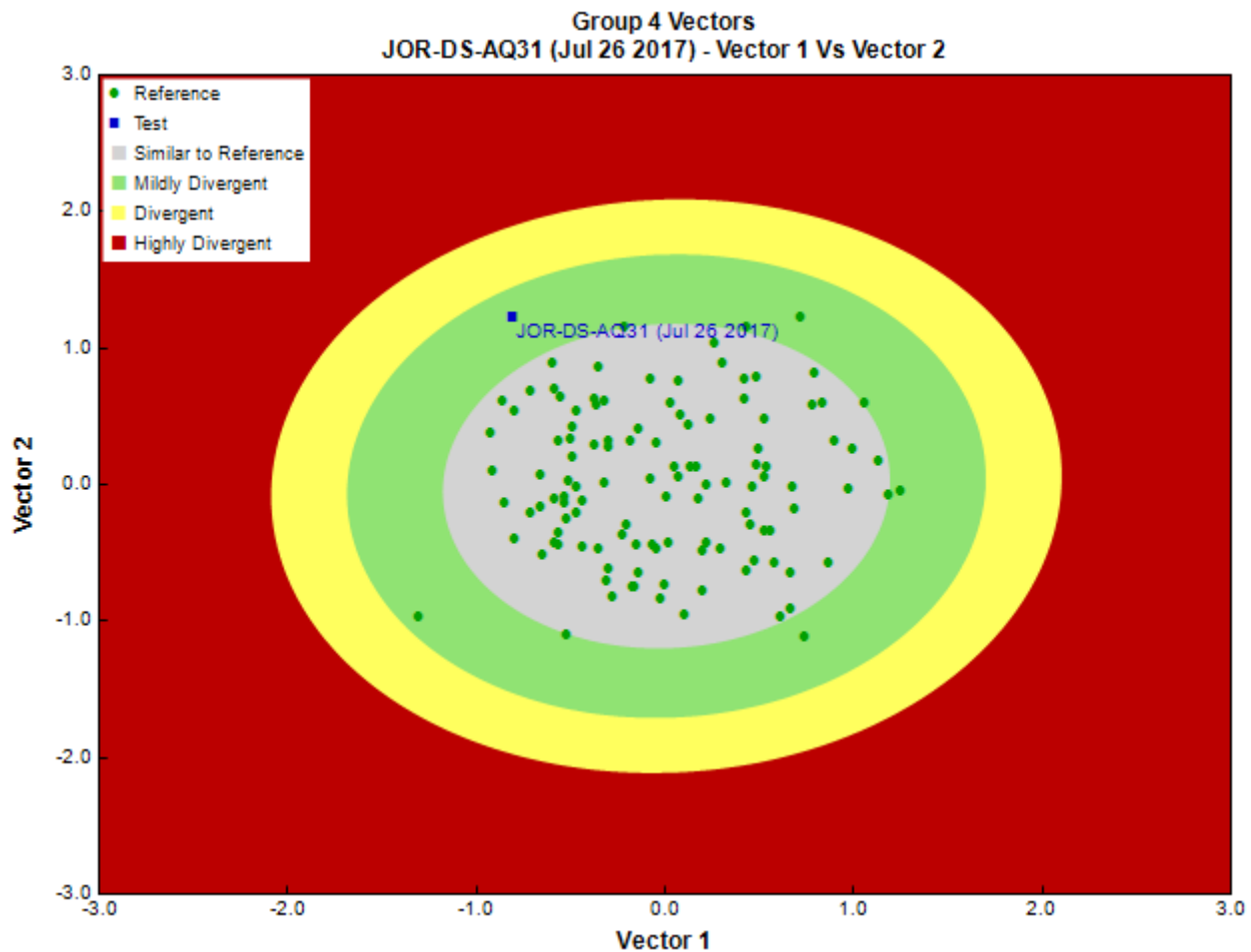


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | November 03, 2017 |
| | Marchant Box |
| Sub-Sample Proportion | 14/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 1 | 7.1 |
| | Collembola | Collembola | | 1 | 7.1 |
| | Insecta | Diptera | Chironomidae | 49 | 350.0 |
| | | | Empididae | 4 | 28.6 |
| | | | Simuliidae | 233 | 1,664.3 |
| | | | Tipulidae | 1 | 7.1 |
| | | Ephemeroptera | Baetidae | 41 | 292.9 |
| | | | Ephemerellidae | 3 | 21.4 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|-----------------|-----------|-------------|
| | | | Leptophlebiidae | 3 | 21.4 |
| | | Plecoptera | Chloroperlidae | 1 | 7.1 |
| | | | Leuctridae | 1 | 7.1 |
| | | | Nemouridae | 20 | 142.9 |
| | | | Perlidae | 1 | 7.1 |
| | | Trichoptera | Rhyacophilidae | 1 | 7.1 |
| | | | Total | 360 | 2,571.2 |

Metrics

| Name | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.76 | 0.5 \pm 0.1 |
| Functional Measures | | |
| % Filterers | -- | 17.2 \pm 42.4 |
| % Gatherers | 85.8 | 57.6 \pm 27.3 |
| % Predatores | 80.3 | 31.3 \pm 20.3 |
| % Scrapers | 76.1 | 37.4 \pm 22.0 |
| % Shredder | 6.1 | 16.1 \pm 11.0 |
| No. Clinger Taxa | 9.0 | 15.8 \pm 6.3 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 80.2 | 32.8 \pm 26.0 |
| % EPT Individuals | 19.8 | 66.1 \pm 26.2 |
| % of 5 dominant taxa | 96.7 | 82.2 \pm 8.7 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.6 | 0.7 \pm 0.2 |
| Total Abundance | 2571.4 | 2646.7 \pm 2772.7 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.1 |
| Coleoptera taxa | 0.0 | 0.3 \pm 0.4 |
| Diptera taxa | 4.0 | 3.2 \pm 1.3 |
| Ephemeroptera taxa | 3.0 | 3.6 \pm 1.1 |
| EPT Individuals (Sum) | 507.1 | 1501.0 \pm 1294.6 |
| EPT taxa (no) | 8.0 | 10.8 \pm 3.5 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.5 | 0.7 \pm 0.1 |
| Plecoptera taxa | 4.0 | 4.1 \pm 1.8 |
| Shannon-Wiener Diversity | 1.2 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.5 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 13.0 | 18.0 \pm 4.5 |
| Trichoptera taxa | 1.0 | 3.1 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.44 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.02 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.09 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.00 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.05 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.04 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.91 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.02 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.36 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.01 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.67 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.24 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.84 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.01 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.02 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.34 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.57 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.32 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.89 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.02 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.02 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.28 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.95 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.61 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.17 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.10 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.08 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.03 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.42 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.40 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.06 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.40 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.34 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.29 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.13 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.15 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.03 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.41 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.79 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.01 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.05 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.29 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.71 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.17 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.02 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.20 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.59 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.33 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.41 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.03 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.52 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.02 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.65 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.31 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.02 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.13 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 9.70 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 0.82 |
| RIVPACS : Expected taxa P>0.70 | 6.09 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.82 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 60.51000 | 28.74839 \pm 35.48825 |
| Channel | | |
| Depth-Avg (cm) | 16.3 | 28.0 \pm 13.9 |
| Depth-BankfullMinusWetted (cm) | 45.00 | 47.88 \pm 26.69 |
| Depth-Max (cm) | 30.0 | 41.3 \pm 21.8 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 4.00 | 0.92 \pm 1.11 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 3 \pm 1 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0200000 | 0.0249850 \pm 0.0294369 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 0 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 1 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.96 | 0.45 \pm 0.21 |
| Velocity-Max (m/s) | 1.69 | 0.67 \pm 0.25 |
| Width-Bankfull (m) | 5.7 | 35.9 \pm 41.6 |
| Width-Wetted (m) | 3.4 | 17.8 \pm 20.2 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 \pm 0 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 \pm 1 |
| Climate | | |
| Precip02_FEB (mm) | 163.27000 | 94.95103 \pm 61.64910 |
| Temp07_JULmax (Degrees Celsius) | 18.72000 | 17.48320 \pm 2.57900 |
| Landcover | | |
| Natl-SnowIce (%) | 3.08000 | 4.62982 \pm 9.77010 |
| Natl-Water (%) | 1.45000 | 1.55060 \pm 2.36345 |
| Natl-WetlandHerb (%) | 0.00000 | 0.18446 \pm 0.50703 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 1 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| %Boulder (%) | 14 | 11 \pm 11 |
| %Cobble (%) | 53 | 53 \pm 11 |
| %Gravel (%) | 6 | 5 \pm 4 |
| %Pebble (%) | 27 | 30 \pm 12 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 1 \pm 3 |
| D50 (cm) | 11.00 | 8.04 \pm 4.60 |
| Dg (cm) | 8.9 | 8.2 \pm 3.1 |
| Dominant-1st (Category(0-9)) | 7 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 6 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 40.36000 | 31.09165 \pm 12.51836 |
| Water Chemistry | | |
| General-DO (mg/L) | 8.9000000 | 11.4180702 \pm 1.2821697 |
| General-pH (pH) | 7.1 | 7.7 \pm 0.7 |
| General-SpCond (μ S/cm) | 105.1000000 | 105.8321429 \pm 89.5097928 |
| General-TempAir (Degrees Celsius) | 16.0 | 12.1 \pm 4.3 |
| General-TempWater (Degrees Celsius) | 14.9000000 | 7.6535897 \pm 3.4680513 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | JOR-DS-AQ31 |
| Sampling Date | Aug 01 2018 |
| Know Your Watershed Basin | Strait of Georgia - East Shore |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.09561 N, 122.99744 W |
| Altitude | 644 |
| Local Basin Name | Jordan Creek |
| | Jordan Creek |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 10.3% | 7.0% | 0.1% | 57.0% | 7.5% | 18.1% |
| CABIN Assessment of JOR-DS-AQ31 on Aug 01, 2018 | Mildly Divergent | | | | | |

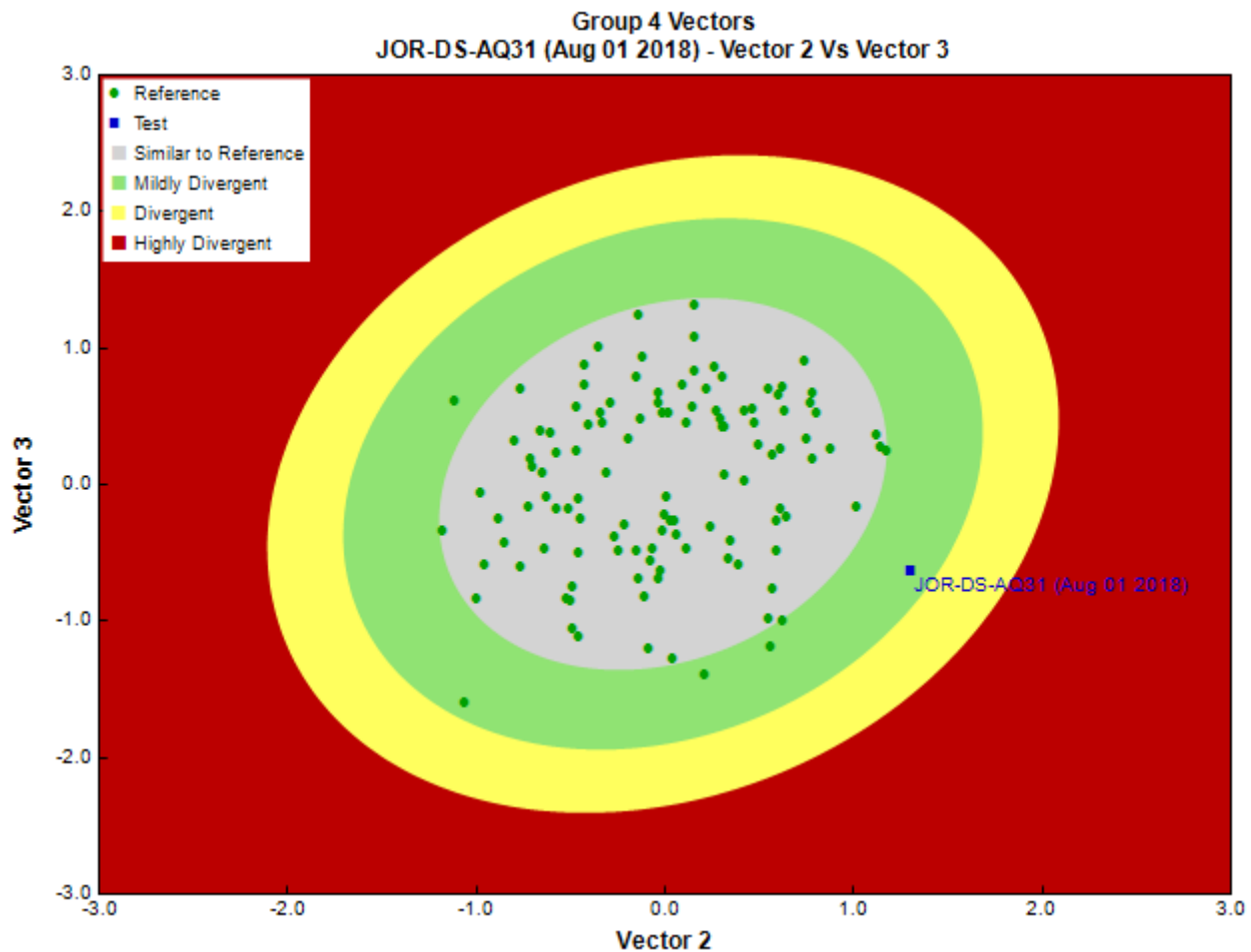


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 17/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Sperchontidae | 1 | 5.9 |
| | Insecta | Diptera | Chironomidae | 46 | 270.5 |
| | | | Simuliidae | 223 | 1,311.8 |
| | | | Tipulidae | 1 | 5.9 |
| | | Ephemeroptera | Ameletidae | 1 | 5.9 |
| | | | Baetidae | 24 | 141.2 |
| | | | Ephemerellidae | 4 | 23.5 |
| | | | Leptophlebiidae | 9 | 52.9 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|----------|-------------|------------------|-----------|-------------|
| | | Plecoptera | Chloroperlidae | 1 | 5.9 |
| | | | Nemouridae | 40 | 235.3 |
| | | | Perlidae | 5 | 29.4 |
| | | Trichoptera | Hydropsychidae | 2 | 11.8 |
| | | | Lepidostomatidae | 2 | 11.8 |
| | | | Rhyacophilidae | 4 | 23.5 |
| Mollusca | Bivalvia | Veneroida | Pisidiidae | 2 | 11.8 |
| | | | Total | 365 | 2,147.1 |

Metrics

| Name | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.73 | 0.5 \pm 0.1 |
| Functional Measures | | |
| % Filterers | -- | 17.2 \pm 42.4 |
| % Gatherers | 89.9 | 57.6 \pm 27.3 |
| % Predatores | 77.0 | 31.3 \pm 20.3 |
| % Scrapers | 68.2 | 37.4 \pm 22.0 |
| % Shredder | 11.8 | 16.1 \pm 11.0 |
| No. Clinger Taxa | 15.0 | 15.8 \pm 6.3 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 74.8 | 32.8 \pm 26.0 |
| % EPT Individuals | 25.2 | 66.1 \pm 26.2 |
| % of 5 dominant taxa | 93.7 | 82.2 \pm 8.7 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.7 | 0.7 \pm 0.2 |
| Total Abundance | 2147.1 | 2646.7 \pm 2772.7 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.1 |
| Coleoptera taxa | 0.0 | 0.3 \pm 0.4 |
| Diptera taxa | 3.0 | 3.2 \pm 1.3 |
| Ephemeroptera taxa | 4.0 | 3.6 \pm 1.1 |
| EPT Individuals (Sum) | 541.2 | 1501.0 \pm 1294.6 |
| EPT taxa (no) | 10.0 | 10.8 \pm 3.5 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.5 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.1 \pm 1.8 |
| Shannon-Wiener Diversity | 1.4 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.6 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 15.0 | 18.0 \pm 4.5 |
| Trichoptera taxa | 3.0 | 3.1 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.44 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.09 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.01 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.06 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.04 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.90 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.34 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.03 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.64 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.26 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 1.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.82 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.01 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.30 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.55 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.33 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.89 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.02 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.26 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.93 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.60 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.16 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.10 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.08 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.04 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.44 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.39 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.09 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.37 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.33 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.27 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.15 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.17 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.04 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.45 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.75 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.02 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.05 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.27 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.70 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.02 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.18 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.03 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.18 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.57 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.31 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.43 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.03 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.51 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.03 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.64 |
| Torrencicolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.31 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.02 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.13 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.01 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 9.48 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 0.84 |
| RIVPACS : Expected taxa P>0.70 | 5.28 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.95 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 60.51000 | 28.74839 \pm 35.48825 |
| Channel | | |
| Depth-Avg (cm) | 30.3 | 28.0 \pm 13.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 4.00 | 0.92 \pm 1.11 |
| Reach-%Logging (PercentRange) | 0 | 0 \pm 1 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 3 \pm 1 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0500000 | 0.0249850 \pm 0.0294369 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 0 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 1 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.55 | 0.45 \pm 0.21 |
| Velocity-Max (m/s) | 0.83 | 0.67 \pm 0.25 |
| Width-Bankfull (m) | 4.4 | 35.9 \pm 41.6 |
| Width-Wetted (m) | 4.2 | 17.8 \pm 20.2 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 \pm 1 |
| Climate | | |
| Precip02_FEB (mm) | 163.27000 | 94.95103 \pm 61.64910 |
| Temp07_JULmax (Degrees Celsius) | 18.72000 | 17.48320 \pm 2.57900 |
| Landcover | | |
| Natl-SnowIce (%) | 3.08000 | 4.62982 \pm 9.77010 |
| Natl-Water (%) | 1.45000 | 1.55060 \pm 2.36345 |
| Natl-WetlandHerb (%) | 0.00000 | 0.18446 \pm 0.50703 |
| Substrate Data | | |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pmSD |
|-------------------------------------|--------------------|--|
| %Bedrock (%) | 0 | 0 \pm 1 |
| %Boulder (%) | 2 | 11 \pm 11 |
| %Cobble (%) | 58 | 53 \pm 11 |
| %Gravel (%) | 5 | 5 \pm 4 |
| %Pebble (%) | 34 | 30 \pm 12 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 1 | 1 \pm 3 |
| D50 (cm) | 8.00 | 8.04 \pm 4.60 |
| Dg (cm) | 6.6 | 8.2 \pm 3.1 |
| Dominant-1st (Category(0-9)) | 6 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 7 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 3 | 4 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 40.36000 | 31.09165 \pm 12.51836 |
| Water Chemistry | | |
| General-DO (mg/L) | 7.7400000 | 11.4180702 \pm 1.2821697 |
| General-pH (pH) | 7.1 | 7.7 \pm 0.7 |
| General-SpCond (μ S/cm) | 65.4000000 | 105.8321429 \pm 89.5097928 |
| General-TempAir (Degrees Celsius) | 23.5 | 12.1 \pm 4.3 |
| General-TempWater (Degrees Celsius) | 18.8000000 | 7.6535897 \pm 3.4680513 |
| General-Turbidity (NTU) | 36.0000000 | 0.5500000 \pm 0.6138116 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | JOR-DS-AQ31 |
| Sampling Date | Jul 30 2019 |
| Know Your Watershed Basin | Strait of Georgia - East Shore |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.09550 N, 122.99729 W |
| Altitude | 0 |
| Local Basin Name | Jordan Creek |
| | Jordan Creek |
| Stream Order | 2 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 7.9% | 6.1% | 0.1% | 61.9% | 2.6% | 21.4% |
| CABIN Assessment of JOR-DS-AQ31 on Jul 30, 2019 | Similar to Reference | | | | | |

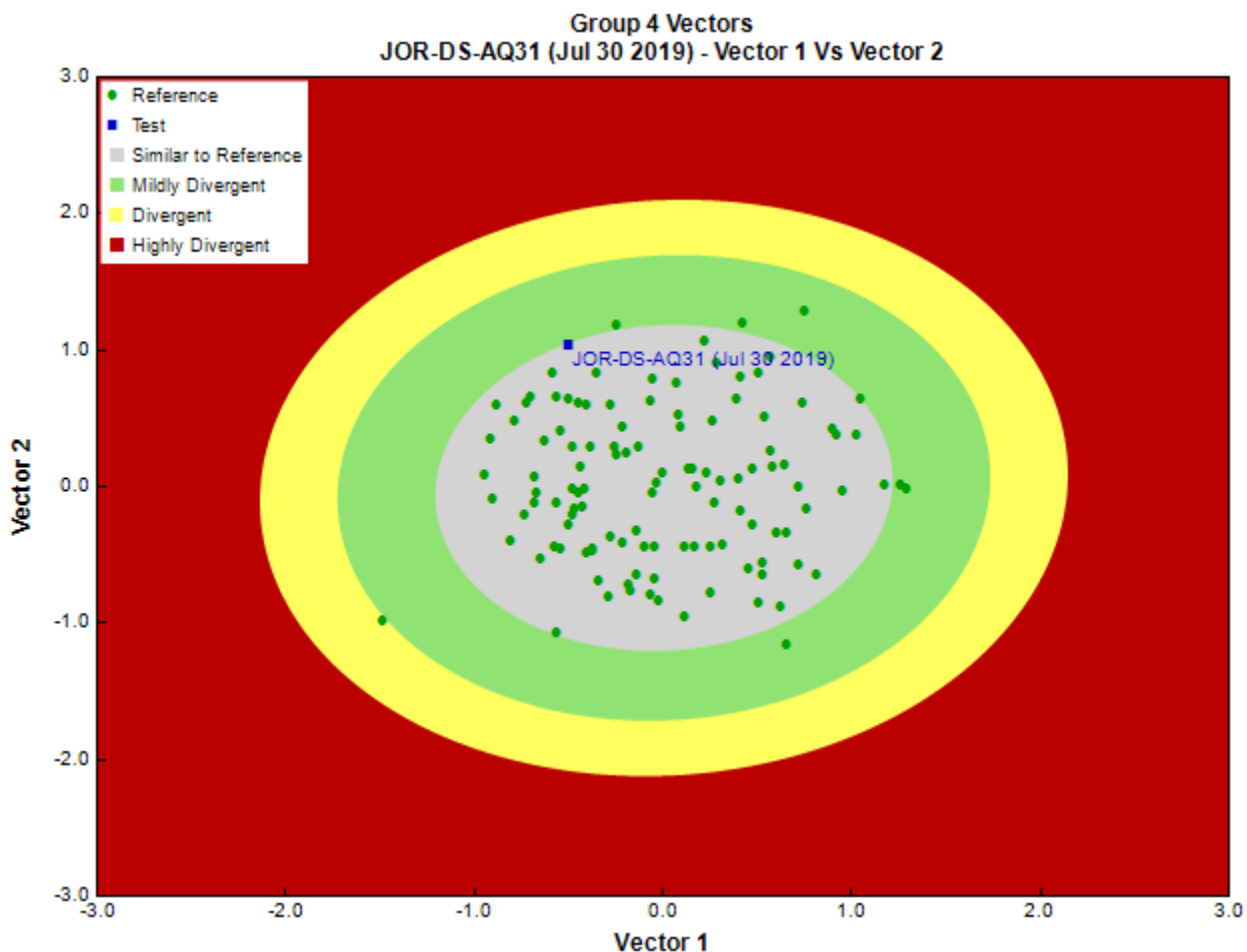


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| Sub-Sample Proportion | 35/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|----------------|-----------|-------------|
| Annelida | Clitellata | Tubificida | Naididae | 2 | 5.7 |
| Arthropoda | Arachnida | Trombidiformes | Hygrobatidae | 1 | 2.9 |
| | Collembola | Collembola | | 1 | 2.9 |
| | Insecta | Diptera | Chironomidae | 160 | 457.3 |
| | | | Simuliidae | 10 | 28.6 |
| | | | Tipulidae | 1 | 2.9 |
| | | Ephemeroptera | Baetidae | 74 | 211.5 |
| | | | Ephemerellidae | 7 | 20.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|------------|----------------|------------------|-----------|-------------|
| | | | Leptophlebiidae | 2 | 5.7 |
| | | Plecoptera | Chloroperlidae | 1 | 2.9 |
| | | | Nemouridae | 48 | 137.2 |
| | | | Perlidae | 3 | 8.6 |
| | | Trichoptera | Glossosomatidae | 1 | 2.9 |
| | | | Hydropsychidae | 1 | 2.9 |
| | | | Lepidostomatidae | 4 | 11.4 |
| | | | Philopotamidae | 1 | 2.9 |
| Cnidaria | Hydrozoa | Anthoathecatae | Hydridae | 1 | 2.9 |
| Mollusca | Bivalvia | Veneroida | Pisidiidae | 8 | 22.9 |
| | Gastropoda | Basommatophora | Physidae | 2 | 5.7 |
| | | | Total | 328 | 937.8 |

Metrics

| Name | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.57 | 0.5 \pm 0.1 |
| Functional Measures | | |
| % Filterers | 0.3 | 17.2 \pm 42.4 |
| % Gatherers | 73.7 | 57.6 \pm 27.3 |
| % Predatores | 53.8 | 31.3 \pm 20.3 |
| % Scrapers | 26.9 | 37.4 \pm 22.0 |
| % Shredder | 16.2 | 16.1 \pm 11.0 |
| No. Clinger Taxa | 18.0 | 15.8 \pm 6.3 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 56.4 | 32.8 \pm 26.0 |
| % EPT Individuals | 43.6 | 66.1 \pm 26.2 |
| % of 5 dominant taxa | 92.0 | 82.2 \pm 8.7 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.5 | 0.7 \pm 0.2 |
| Total Abundance | 934.3 | 2646.7 \pm 2772.7 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.1 |
| Coleoptera taxa | 0.0 | 0.3 \pm 0.4 |
| Diptera taxa | 3.0 | 3.2 \pm 1.3 |
| Ephemeroptera taxa | 3.0 | 3.6 \pm 1.1 |
| EPT Individuals (Sum) | 405.7 | 1501.0 \pm 1294.6 |
| EPT taxa (no) | 10.0 | 10.8 \pm 3.5 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.5 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.1 \pm 1.8 |
| Shannon-Wiener Diversity | 1.5 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.7 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 18.0 | 18.0 \pm 4.5 |
| Trichoptera taxa | 4.0 | 3.1 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.47 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.09 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.00 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.06 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.04 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.91 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.02 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.35 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.02 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.66 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.24 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.99 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.85 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.00 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.01 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.30 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.57 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.34 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.90 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.02 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.27 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.95 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.62 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.15 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.10 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.08 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.04 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.43 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.39 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.06 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.36 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.34 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.29 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.14 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.16 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.04 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.43 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.78 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.01 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.05 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.28 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.73 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.02 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.15 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.02 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at JOR-DS-AQ31 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.01 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.18 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.60 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.31 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.45 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.01 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.03 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.53 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.03 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.01 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.65 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.30 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.03 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.14 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.00 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 9.74 |
| RIVPACS : Observed taxa P>0.50 | 7.00 |
| RIVPACS : O:E (p > 0.5) | 0.72 |
| RIVPACS : Expected taxa P>0.70 | 6.11 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 0.82 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 60.51000 | 28.74839 \pm 35.48825 |
| Channel | | |
| Depth-BankfullMinusWetted (cm) | 14.00 | 47.88 \pm 26.69 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 3.00 | 0.92 \pm 1.11 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 \pm 1 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0500000 | 0.0249850 \pm 0.0294369 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 0 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 1 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.67 | 0.45 \pm 0.21 |
| Velocity-Max (m/s) | 0.89 | 0.67 \pm 0.25 |
| Width-Bankfull (m) | 4.2 | 35.9 \pm 41.6 |
| Width-Wetted (m) | 4.4 | 17.8 \pm 20.2 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 \pm 1 |
| Climate | | |
| Precip02_FEB (mm) | 163.27000 | 94.95103 \pm 61.64910 |
| Temp07_JULmax (Degrees Celsius) | 18.72000 | 17.48320 \pm 2.57900 |
| Landcover | | |
| Natl-SnowIce (%) | 3.08000 | 4.62982 \pm 9.77010 |

Habitat Description

| Variable | JOR-DS-AQ31 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Natl-Water (%) | 1.45000 | 1.55060 \pm 2.36345 |
| Natl-WetlandHerb (%) | 0.00000 | 0.18446 \pm 0.50703 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 1 |
| %Boulder (%) | 15 | 11 \pm 11 |
| %Cobble (%) | 72 | 53 \pm 11 |
| %Gravel (%) | 2 | 5 \pm 4 |
| %Pebble (%) | 11 | 30 \pm 12 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 1 \pm 3 |
| D50 (cm) | 14.50 | 8.04 \pm 4.60 |
| Dg (cm) | 12.6 | 8.2 \pm 3.1 |
| Dominant-1st (Category(0-9)) | 7 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 6 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 4 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 40.36000 | 31.09165 \pm 12.51836 |
| Water Chemistry | | |
| General-Conductivity (μ S/cm) | 67.1000000 | 92.7298969 \pm 75.6979499 |
| General-DO (mg/L) | 9.4400000 | 11.4180702 \pm 1.2821697 |
| General-pH (pH) | 7.7 | 7.7 \pm 0.7 |
| General-SpCond (μ S/cm) | 78.4000000 | 105.8321429 \pm 89.5097928 |
| General-TempAir (Degrees Celsius) | 12.0 | 12.1 \pm 4.3 |
| General-TempWater (Degrees Celsius) | 17.4000000 | 7.6535897 \pm 3.4680513 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-DS-AQ12 |
| Sampling Date | Aug 05 2016 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.14432 N, 122.95758 W |
| Altitude | 631 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 1 2 3 4 5 6 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 17.2% | 8.4% | 16.0% | 26.5% | 22.7% | 9.2% |
| CABIN Assessment of RGD-DS-AQ12 on Aug 05, 2016 | Similar to Reference | | | | | |

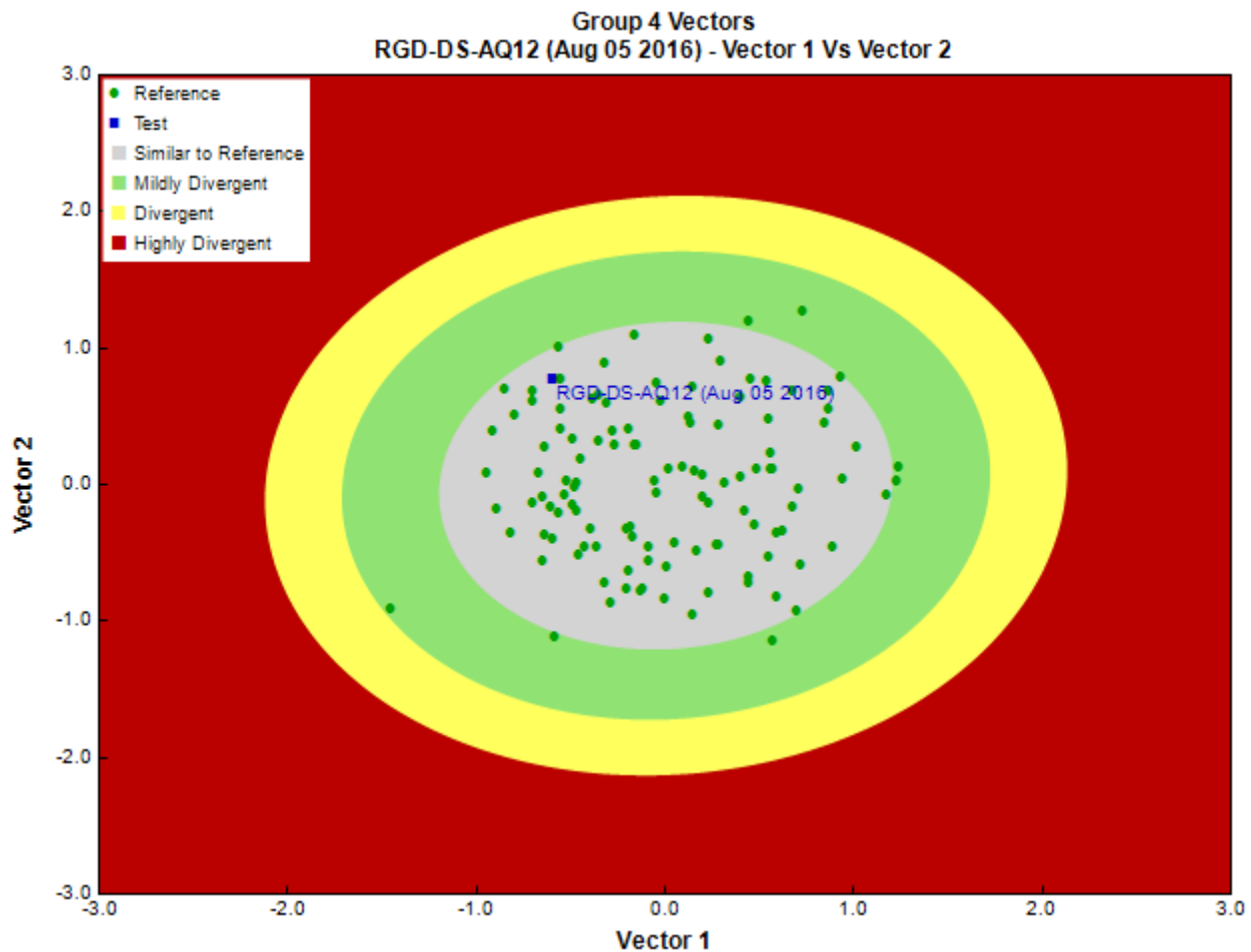


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | October 20, 2016 |
| | Marchant Box |
| Sub-Sample Proportion | 19/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 16 | 84.2 |
| | Insecta | Coleoptera | Dytiscidae | 2 | 10.5 |
| | | Diptera | Ceratopogonidae | 2 | 10.5 |
| | | | Chironomidae | 45 | 236.8 |
| | | | Empididae | 4 | 21.0 |
| | | | Simuliidae | 3 | 15.8 |
| | | Ephemeroptera | Baetidae | 80 | 421.0 |
| | | | Ephemerellidae | 71 | 373.7 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | | Heptageniidae | 2 | 10.5 |
| | | Plecoptera | Chloroperlidae | 1 | 5.3 |
| | | | Leuctridae | 5 | 26.3 |
| | | | Nemouridae | 73 | 384.2 |
| | | | Perlodidae | 1 | 5.3 |
| | | Trichoptera | Hydropsychidae | 5 | 26.3 |
| | | | Limnephilidae | 1 | 5.3 |
| | | | Rhyacophilidae | 1 | 5.3 |
| | | | Total | 312 | 1,642.0 |

Metrics

| Name | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.57 | 0.5 \pm 0.1 |
| Functional Measures | | |
| % Filterers | -- | 17.2 \pm 42.4 |
| % Gatherers | 63.1 | 57.6 \pm 27.3 |
| % Predatores | 24.7 | 31.3 \pm 20.3 |
| % Scrapers | 29.2 | 37.4 \pm 22.0 |
| % Shredder | 25.3 | 16.1 \pm 11.0 |
| No. Clinger Taxa | 11.0 | 15.8 \pm 6.3 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 22.4 | 32.8 \pm 26.0 |
| % EPT Individuals | 76.9 | 66.1 \pm 26.2 |
| % of 5 dominant taxa | 91.3 | 82.2 \pm 8.7 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.8 | 0.7 \pm 0.2 |
| Total Abundance | 1642.1 | 2646.7 \pm 2772.7 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.1 |
| Coleoptera taxa | 1.0 | 0.3 \pm 0.4 |
| Diptera taxa | 4.0 | 3.2 \pm 1.3 |
| Ephemeroptera taxa | 3.0 | 3.6 \pm 1.1 |
| EPT Individuals (Sum) | 1263.1 | 1501.0 \pm 1294.6 |
| EPT taxa (no) | 10.0 | 10.8 \pm 3.5 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.7 \pm 0.1 |
| Plecoptera taxa | 4.0 | 4.1 \pm 1.8 |
| Shannon-Wiener Diversity | 1.9 | 1.9 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.8 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 16.0 | 18.0 \pm 4.5 |
| Trichoptera taxa | 3.0 | 3.1 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.30 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.08 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.04 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.05 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.03 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.83 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.28 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.09 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.55 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.27 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.98 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.74 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.02 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.28 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.44 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.26 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.82 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.02 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.20 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.87 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.45 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.17 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.40 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.31 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.17 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.31 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.24 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.18 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.18 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.03 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.44 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.62 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.04 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.03 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.22 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.63 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.23 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.08 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.14 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.08 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.44 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.26 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.34 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.02 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.51 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.55 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.26 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.02 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.13 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 7.11 |
| RIVPACS : Observed taxa P>0.50 | 7.00 |
| RIVPACS : O:E (p > 0.5) | 0.99 |
| RIVPACS : Expected taxa P>0.70 | 4.24 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.18 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.17000 | 28.74839 \pm 35.48825 |
| Channel | | |
| Depth-Avg (cm) | 11.5 | 28.0 \pm 13.9 |
| Depth-BankfullMinusWetted (cm) | 62.00 | 47.88 \pm 26.69 |
| Depth-Max (cm) | 16.0 | 41.3 \pm 21.8 |
| Macrophyte (PercentRange) | 1 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.92 \pm 1.11 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 3 \pm 1 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0050000 | 0.0249850 \pm 0.0294369 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 0 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 1 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.27 | 0.45 \pm 0.21 |
| Velocity-Max (m/s) | 0.31 | 0.67 \pm 0.25 |
| Width-Bankfull (m) | 14.8 | 35.9 \pm 41.6 |
| Width-Wetted (m) | 13.3 | 17.8 \pm 20.2 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 \pm 0 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 \pm 1 |
| Climate | | |
| Precip02_FEB (mm) | 156.00000 | 94.95103 \pm 61.64910 |
| Temp07_JULmax (Degrees Celsius) | 18.66000 | 17.48320 \pm 2.57900 |
| Landcover | | |
| Natl-SnowIce (%) | 22.06000 | 4.62982 \pm 9.77010 |
| Natl-Water (%) | 2.36000 | 1.55060 \pm 2.36345 |
| Natl-WetlandHerb (%) | 0.00000 | 0.18446 \pm 0.50703 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 1 |
| %Boulder (%) | 0 | 11 \pm 11 |
| %Cobble (%) | 2 | 53 \pm 11 |
| %Gravel (%) | 23 | 5 \pm 4 |
| %Pebble (%) | 75 | 30 \pm 12 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 1 \pm 3 |
| D50 (cm) | 3.00 | 8.04 \pm 4.60 |
| Dg (cm) | 2.5 | 8.2 \pm 3.1 |
| Dominant-1st (Category(0-9)) | 5 | 6 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 6 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 3 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 36.72000 | 31.09165 \pm 12.51836 |
| Water Chemistry | | |
| General-DO (mg/L) | 9.8900000 | 11.4180702 \pm 1.2821697 |
| General-pH (pH) | 7.8 | 7.7 \pm 0.7 |
| General-SpCond (μ S/cm) | 69.0000000 | 105.8321429 \pm 89.5097928 |
| General-TempAir (Degrees Celsius) | 22.0 | 12.1 \pm 4.3 |
| General-TempWater (Degrees Celsius) | 15.2000000 | 7.6535897 \pm 3.4680513 |
| General-Turbidity (NTU) | 0.3000000 | 0.5500000 \pm 0.6138116 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-DS-AQ12 |
| Sampling Date | Jul 25 2017 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.14417 N, 122.95750 W |
| Altitude | 194 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |

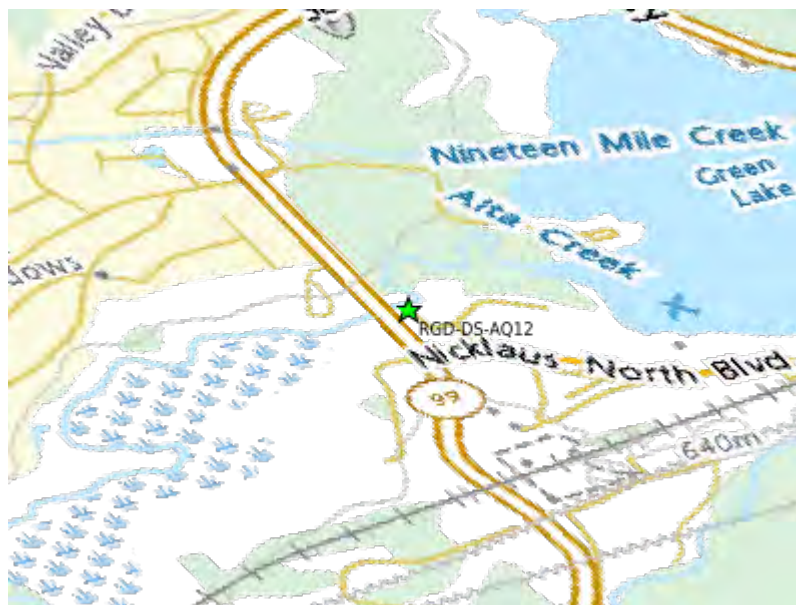


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 15.6% | 6.7% | 9.6% | 16.9% | 45.8% | 5.4% |
| CABIN Assessment of RGD-DS-AQ12 on Jul 25, 2017 | Similar to Reference | | | | | |

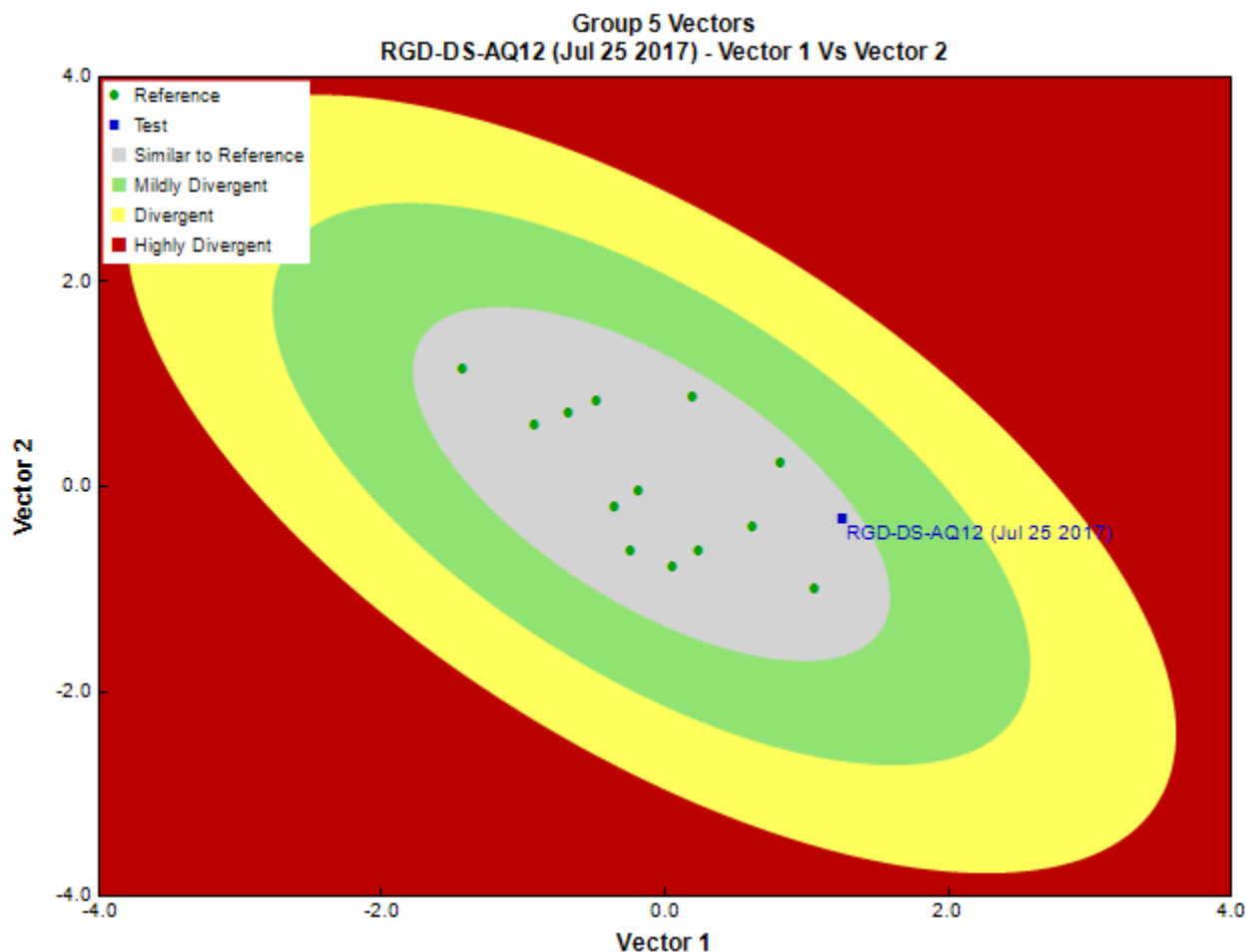


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | November 02, 2017 |
| | Marchant Box |
| Sub-Sample Proportion | 35/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 9 | 25.7 |
| | Insecta | Coleoptera | Dytiscidae | 2 | 5.7 |
| | | Diptera | Chironomidae | 91 | 260.0 |
| | | | Empididae | 2 | 5.7 |
| | | | Simuliidae | 24 | 68.6 |
| | | Ephemeroptera | Ameletidae | 2 | 5.7 |
| | | | Baetidae | 86 | 245.7 |
| | | | Ephemerellidae | 28 | 80.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | | Heptageniidae | 93 | 265.7 |
| | | Plecoptera | Chloroperlidae | 5 | 14.3 |
| | | Trichoptera | Limnephilidae | 1 | 2.9 |
| | | | Rhyacophilidae | 7 | 20.0 |
| | | | Total | 350 | 1,000.0 |

Metrics

| Name | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.72 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 41.4 | 67.6 \pm 30.3 |
| % Predatores | 38.6 | 41.1 \pm 20.2 |
| % Scrapers | 58.3 | 34.3 \pm 21.0 |
| % Shredder | 0.3 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 7.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 36.0 | 47.4 \pm 26.3 |
| % EPT Individuals | 63.4 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 92.0 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.7 | 0.6 \pm 0.2 |
| Total Abundance | 1000.0 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 1.0 | 0.5 \pm 0.5 |
| Diptera taxa | 3.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 4.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 634.3 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 7.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.6 \pm 0.1 |
| Plecoptera taxa | 1.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 1.8 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.4 | 0.3 \pm 0.1 |
| Total No. of Taxa | 12.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 2.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.20 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.08 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.07 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.06 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.77 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.25 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.18 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.49 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.36 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.99 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.61 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.04 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesiidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.26 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.33 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.28 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.76 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.02 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.07 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.14 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.77 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.39 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.20 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.05 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.44 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.29 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.30 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.28 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.20 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.16 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.27 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.26 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.04 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.56 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.49 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.07 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.02 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.18 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.52 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.33 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.15 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.11 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.10 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.30 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.22 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.30 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.04 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.39 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.50 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.26 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.09 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.07 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.05 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 5.48 |
| RIVPACS : Observed taxa P>0.50 | 5.00 |
| RIVPACS : O:E (p > 0.5) | 0.91 |
| RIVPACS : Expected taxa P>0.70 | 3.29 |
| RIVPACS : Observed taxa P>0.70 | 4.00 |
| RIVPACS : O:E (p > 0.7) | 1.22 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.17000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 32.2 | 40.5 \pm 22.4 |
| Depth-BankfullMinusWetted (cm) | 50.00 | 188.00 |
| Depth-Max (cm) | 39.0 | 55.5 \pm 31.7 |
| Macrophyte (PercentRange) | 1 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.23 \pm 0.44 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 2 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0050000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 1 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.33 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 0.47 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 15.4 | 75.1 \pm 72.8 |
| Width-Wetted (m) | 15.4 | 50.6 \pm 60.4 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 156.00000 | 171.50745 \pm 107.47690 |
| Temp07_JULmax (Degrees Celsius) | 18.66000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 22.06000 | 3.62533 \pm 10.17162 |
| Natl-Water (%) | 2.36000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 0 | 0 |
| %Cobble (%) | 0 | 58 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pmSD |
|-------------------------------------|--------------------|--|
| %Gravel (%) | 38 | 1 |
| %Pebble (%) | 62 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 0 | 0 |
| D50 (cm) | 2.00 | 3.30 |
| Dg (cm) | 1.7 | 6.6 |
| Dominant-1st (Category(0-9)) | 4 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 3 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 3 |
| SurroundingMaterial (Category(0-9)) | 2 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 36.72000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-DO (mg/L) | 9.7700000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 7.0 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 73.3000000 | 176.1000000 |
| General-TempAir (Degrees Celsius) | 26.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 13.0000000 | 13.2730769 \pm 4.7663725 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-DS-AQ12 |
| Sampling Date | Aug 01 2018 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.14440 N, 122.95752 W |
| Altitude | 641 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |

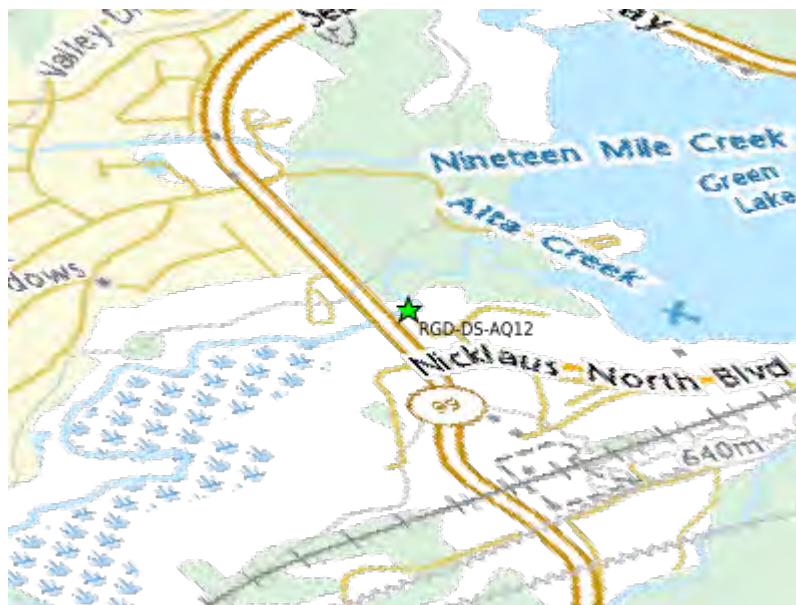


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 12.2% | 4.4% | 4.7% | 7.9% | 68.5% | 2.4% |
| CABIN Assessment of RGD-DS-AQ12 on Aug 01, 2018 | Similar to Reference | | | | | |

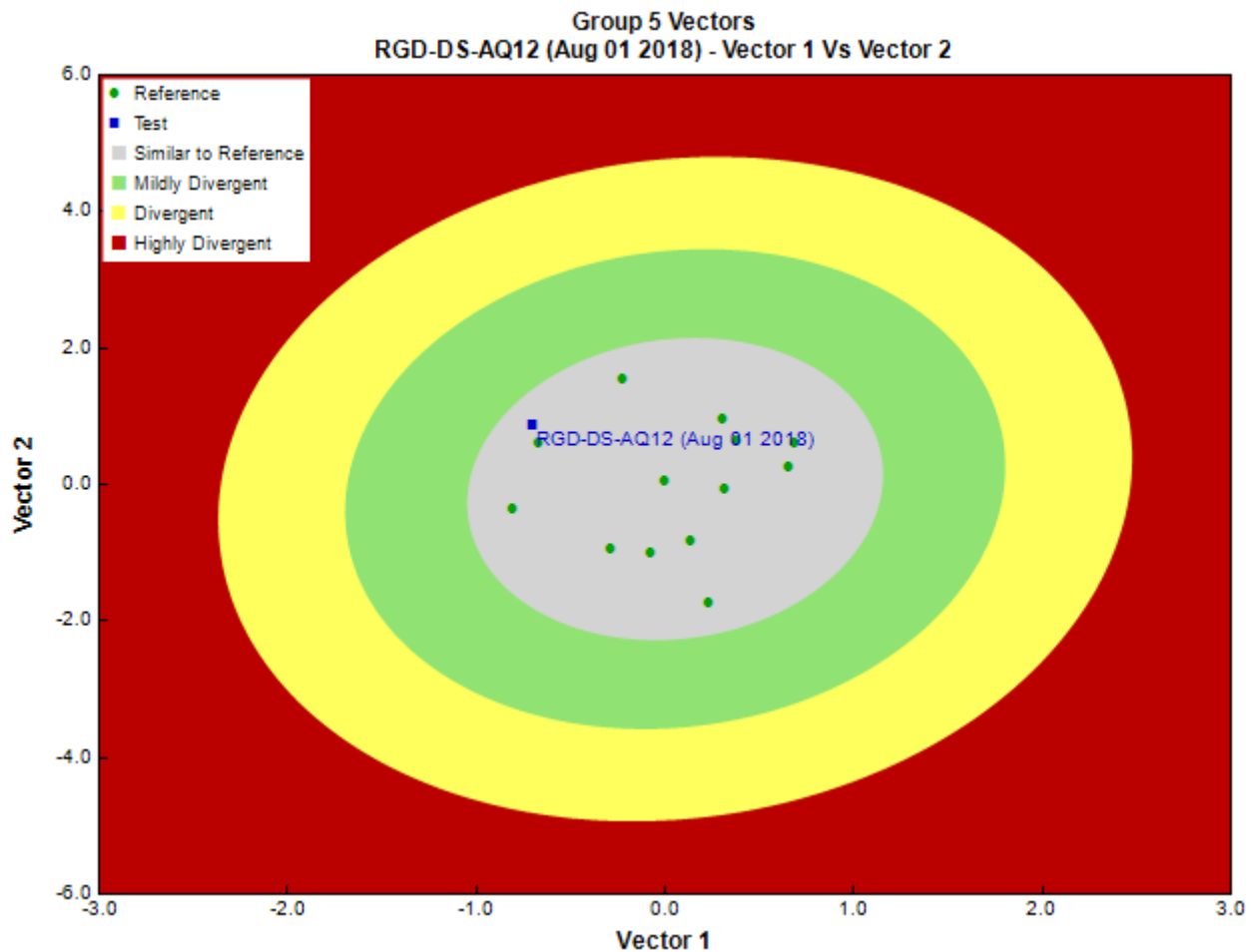


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 100/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|-----------------|-----------|-------------|
| Annelida | Clitellata | Lumbriculida | Lumbriculidae | 9 | 9.1 |
| | | Tubificida | Naididae | 161 | 161.4 |
| Arthropoda | Arachnida | Trombidiformes | | 2 | 2.3 |
| | | | Hygrobatidae | 22 | 22.8 |
| | | | Lebertiidae | 9 | 9.1 |
| | | | Sperchontidae | 5 | 4.5 |
| | Insecta | Coleoptera | Dytiscidae | 78 | 77.2 |
| | | Diptera | Ceratopogonidae | 2 | 2.3 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|--------------|---------------|------------------|-----------|-------------|
| | | | Chironomidae | 188 | 188.7 |
| | | | Empididae | 2 | 2.3 |
| | | | Simuliidae | 15 | 14.6 |
| | | | Tipulidae | 4 | 4.6 |
| | | Ephemeroptera | Ameletidae | 14 | 13.6 |
| | | | Baetidae | 85 | 84.0 |
| | | | Ephemerellidae | 96 | 95.4 |
| | | | Leptophlebiidae | 34 | 34.1 |
| | | Megaloptera | Sialidae | 2 | 2.3 |
| | | Plecoptera | Chloroperlidae | 15 | 16.0 |
| | | | Nemouridae | 40 | 38.5 |
| | | Trichoptera | Lepidostomatidae | 2 | 2.3 |
| | | | Limnephilidae | 4 | 4.6 |
| | | | Rhyacophilidae | 5 | 4.5 |
| | Malacostraca | Amphipoda | | 2 | 2.3 |
| | | | Crangonyctidae | 2 | 2.3 |
| Mollusca | Bivalvia | Veneroida | Pisidiidae | 7 | 6.8 |
| | | | Total | 805 | 805.6 |

Metrics

| Name | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.59 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 90.7 | 67.6 \pm 30.3 |
| % Predatores | 40.5 | 41.1 \pm 20.2 |
| % Scrapers | 12.8 | 34.3 \pm 21.0 |
| % Shredder | 6.2 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 20.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 53.5 | 47.4 \pm 26.3 |
| % EPT Individuals | 36.6 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 75.8 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.6 | 0.6 \pm 0.2 |
| Total Abundance | 805.5 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 1.0 | 0.5 \pm 0.5 |
| Diptera taxa | 5.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 4.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 293.2 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 9.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.6 \pm 0.1 |
| Plecoptera taxa | 2.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 2.3 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.9 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 23.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 3.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.11 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.08 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.11 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.06 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.01 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.71 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.00 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.22 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.26 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.42 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.43 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.99 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.48 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.06 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.22 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.23 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.29 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.70 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.00 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.11 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.09 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.67 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.32 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.21 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.01 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.03 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.00 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.48 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.27 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.43 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.00 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.24 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.15 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.09 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.35 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.34 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.06 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.00 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.68 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.36 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.11 |
| Pelecchynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.01 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.14 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.43 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.00 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.00 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.42 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.21 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.07 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.13 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.18 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.17 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.26 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.05 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.29 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.00 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.46 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.25 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.05 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.11 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.06 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 3.75 |
| RIVPACS : Observed taxa P>0.50 | 4.00 |
| RIVPACS : O:E (p > 0.5) | 1.07 |
| RIVPACS : Expected taxa P>0.70 | 1.71 |
| RIVPACS : Observed taxa P>0.70 | 2.00 |
| RIVPACS : O:E (p > 0.7) | 1.17 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.17000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 28.9 | 40.5 \pm 22.4 |
| Macrophyte (PercentRange) | 1 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.23 \pm 0.44 |
| Reach-%Logging (PercentRange) | 0 | 0 \pm 0 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 1 \pm 0 |
| Slope (m/m) | 0.000000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 0 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.26 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 0.31 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 16.6 | 75.1 \pm 72.8 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Width-Wetted (m) | 16.4 | 50.6 \pm 60.4 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 156.00000 | 171.50745 \pm 107.47690 |
| Temp07_JULmax (Degrees Celsius) | 18.66000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 22.06000 | 3.62533 \pm 10.17162 |
| Natl-Water (%) | 2.36000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 0 | 0 |
| %Cobble (%) | 1 | 58 |
| %Gravel (%) | 36 | 1 |
| %Pebble (%) | 47 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 16 | 0 |
| D50 (cm) | 1.30 | 3.30 |
| Dg (cm) | 0.9 | 6.6 |
| Dominant-1st (Category(0-9)) | 3 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 4 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 3 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 3 |
| SurroundingMaterial (Category(0-9)) | 2 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 36.72000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-DO (mg/L) | 8.1600000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 6.7 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 48.3000000 | 176.1000000 |
| General-TempAir (Degrees Celsius) | 29.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 17.8000000 | 13.2730769 \pm 4.7663725 |
| General-Turbidity (NTU) | 31.0000000 | 0.0000000 \pm 0.0000000 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-DS-AQ12 |
| Sampling Date | Jul 31 2019 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.14430 N, 122.95764 W |
| Altitude | 614 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |

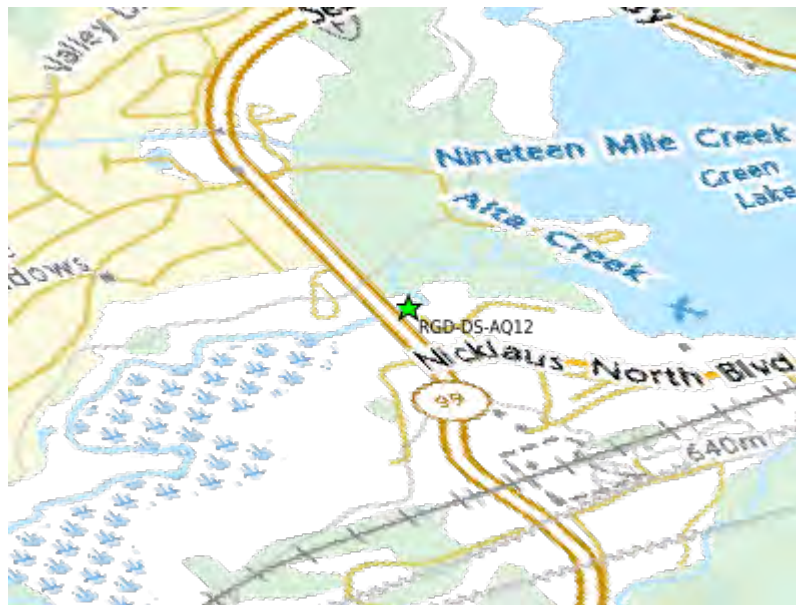


Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|--|----------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 17.6% | 7.2% | 10.0% | 16.2% | 43.8% | 5.2% |
| CABIN Assessment of RGD-DS-AQ12 on Jul 31, 2019 | Similar to Reference | | | | | |

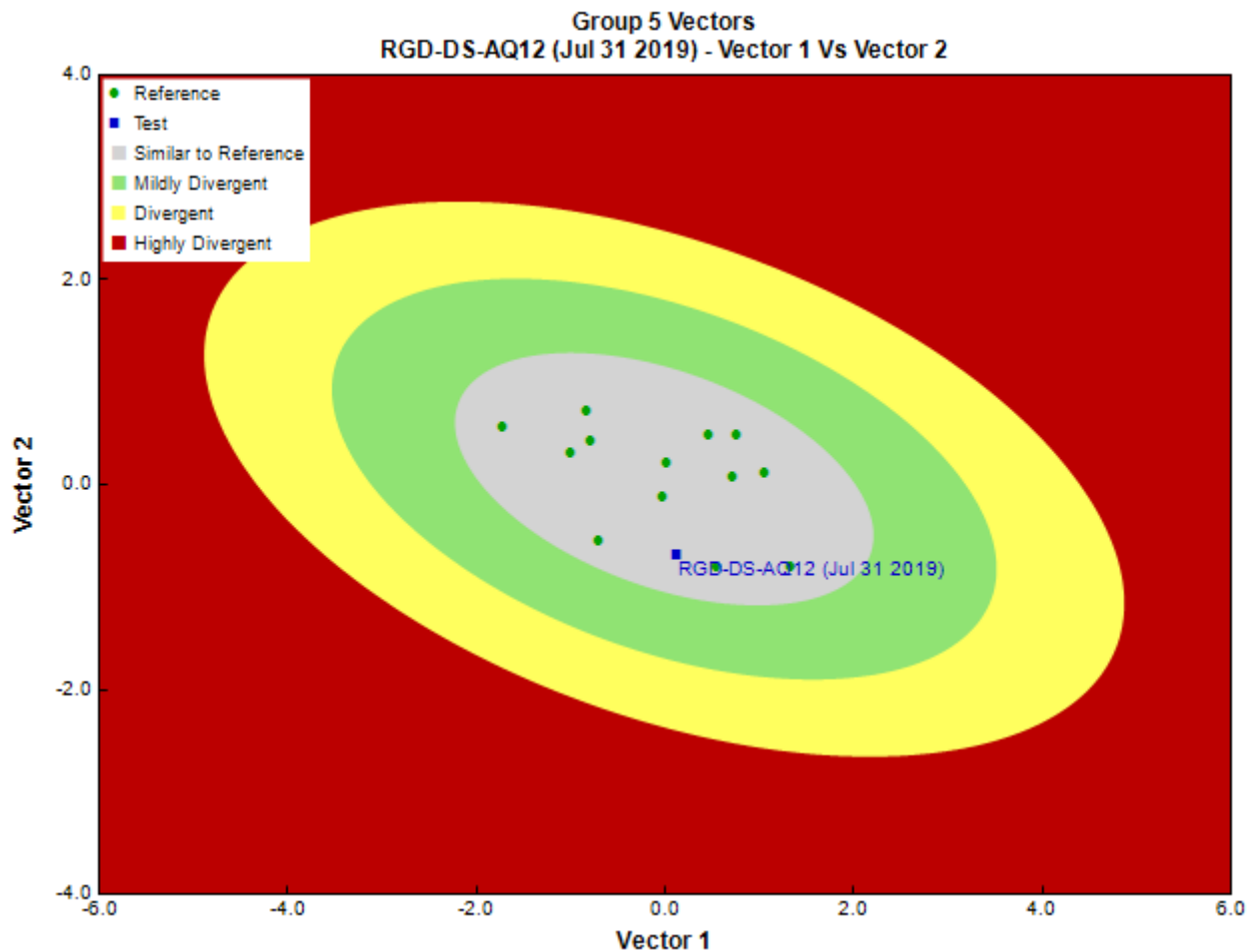


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| Sub-Sample Proportion | 23/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|---------------|-----------|-------------|
| Annelida | Clitellata | Lumbriculida | Lumbriculidae | 3 | 13.0 |
| | | Tubificida | Naididae | 80 | 347.8 |
| Arthropoda | Arachnida | Trombidiformes | Hygrobatidae | 8 | 34.7 |
| | | | Lebertiidae | 1 | 4.3 |
| | | | Sperchontidae | 2 | 8.7 |
| | Insecta | Coleoptera | Dytiscidae | 37 | 160.9 |
| | | Diptera | Chironomidae | 80 | 347.5 |
| | | | Simuliidae | 1 | 4.3 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|--------------|---------------|-----------------|-----------|-------------|
| | | | Tipulidae | 1 | 4.3 |
| | | Ephemeroptera | Ameletidae | 1 | 4.3 |
| | | | Baetidae | 25 | 108.7 |
| | | | Ephemerellidae | 55 | 239.1 |
| | | | Heptageniidae | 1 | 4.3 |
| | | | Leptophlebiidae | 4 | 17.4 |
| | | Plecoptera | Chloroperlidae | 3 | 13.0 |
| | | | Nemouridae | 4 | 17.3 |
| | | Trichoptera | Limnephilidae | 2 | 8.6 |
| | | | Rhyacophilidae | 1 | 4.3 |
| | Malacostraca | Amphipoda | | 2 | 8.7 |
| | | | Crangonyctidae | 1 | 4.3 |
| Mollusca | Bivalvia | Veneroida | Pisidiidae | 1 | 4.3 |
| | | | Total | 313 | 1,359.8 |

Metrics

| Name | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|---|-------------|---|
| Bray-Curtis Distance | 0.39 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 99.4 | 67.6 \pm 30.3 |
| % Predatores | 41.5 | 41.1 \pm 20.2 |
| % Scrapers | 9.3 | 34.3 \pm 21.0 |
| % Shredder | 2.2 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 18.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 57.2 | 47.4 \pm 26.3 |
| % EPT Individuals | 30.9 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 89.1 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.5 | 0.6 \pm 0.2 |
| Total Abundance | 1360.9 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 1.0 | 0.5 \pm 0.5 |
| Diptera taxa | 3.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 5.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 417.4 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 9.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.6 \pm 0.1 |
| Plecoptera taxa | 2.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 2.0 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 20.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 2.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.21 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.07 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.05 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.78 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.26 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.17 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.50 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.35 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.99 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.62 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.04 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.26 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.34 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.27 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.76 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.02 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.07 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.15 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halipidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.78 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.39 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.20 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.05 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.44 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.30 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.29 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.29 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.20 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.16 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.26 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.25 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.04 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.55 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.50 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.07 |
| Pelecchynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.01 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.02 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.18 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.53 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-DS-AQ12 |
|----------------------|--|---------|---------|---------|---------|---------|--|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.33 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.14 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.11 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.10 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.31 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.23 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.30 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.04 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.40 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.51 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.26 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.09 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.07 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.05 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 6.01 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 1.33 |
| RIVPACS : Expected taxa P>0.70 | 3.31 |
| RIVPACS : Observed taxa P>0.70 | 4.00 |
| RIVPACS : O:E (p > 0.7) | 1.21 |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|--|-------------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.17000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 33.9 | 40.5 \pm 22.4 |
| Depth-BankfullMinusWetted (cm) | 7.00 | 188.00 |
| Depth-Max (cm) | 54.5 | 55.5 \pm 31.7 |
| Macrophyte (PercentRange) | 1 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.23 \pm 0.44 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 2 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 1 \pm 0 |
| Slope (m/m) | 0.0000000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 0 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.22 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 0.54 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 15.6 | 75.1 \pm 72.8 |
| Width-Wetted (m) | 15.5 | 50.6 \pm 60.4 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |

Habitat Description

| Variable | RGD-DS-AQ12 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|-------------|--|
| Precip02_FEB (mm) | 156.00000 | 171.50745 \pm 107.47690 |
| Precip03_MAR (mm) | 156.00000 | 152.05098 \pm 91.49370 |
| Temp07_JULmax (Degrees Celsius) | 18.66000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 22.06000 | 3.62533 \pm 10.17162 |
| Natl-Water (%) | 2.36000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 0 | 0 |
| %Cobble (%) | 0 | 58 |
| %Gravel (%) | 29 | 1 |
| %Pebble (%) | 71 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 0 | 0 |
| D50 (cm) | 2.35 | 3.30 |
| Dg (cm) | 2.1 | 6.6 |
| Dominant-1st (Category(0-9)) | 4 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 5 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 2 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 4 | 3 |
| SurroundingMaterial (Category(0-9)) | 2 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 36.72000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-Conductivity (μ S/cm) | 60.6000000 | 79.0846153 \pm 50.3407694 |
| General-DO (mg/L) | 9.9300000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 7.6 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 78.4000000 | 176.1000000 |
| General-TempAir (Degrees Celsius) | 17.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 13.1000000 | 13.2730769 \pm 4.7663725 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-AQ11 |
| Sampling Date | Aug 03 2016 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12703 N, 122.97202 W |
| Altitude | 642 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|---|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 9.1% | 4.7% | 37.9% | 21.9% | 16.7% | 9.7% |
| CABIN Assessment of RGD-AQ11 on Aug 03, 2016 | Mildly Divergent | | | | | |

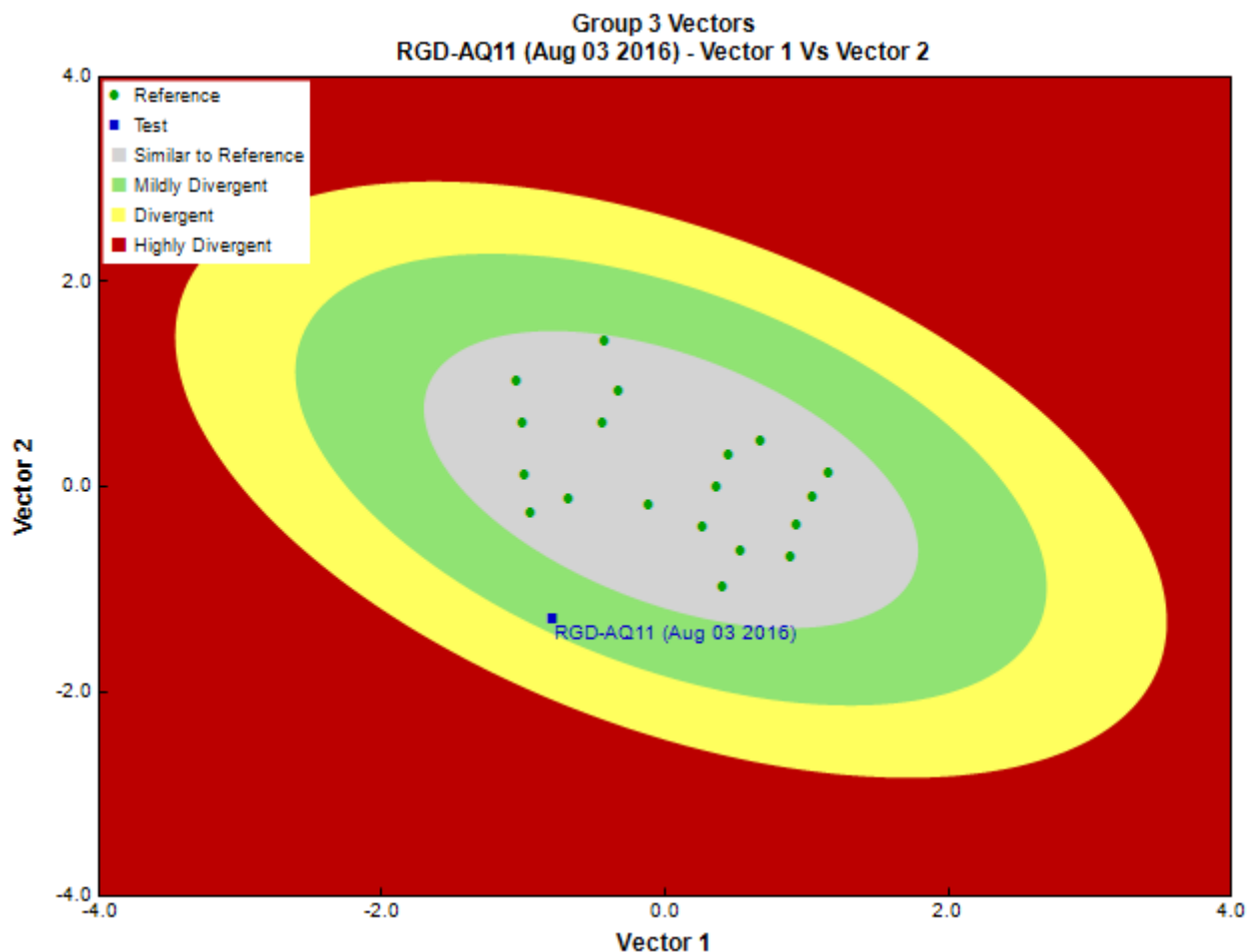


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | September 29, 2016 |
| | Marchant Box |
| Sub-Sample Proportion | 26/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 5 | 19.2 |
| | Insecta | Diptera | Ceratopogonidae | 8 | 30.8 |
| | | | Chironomidae | 8 | 30.8 |
| | | | Empididae | 5 | 19.2 |
| | | | Simuliidae | 30 | 115.4 |
| | | Ephemeroptera | Baetidae | 102 | 392.3 |
| | | | Ephemerellidae | 7 | 26.9 |
| | | | Heptageniidae | 71 | 273.1 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | Plecoptera | Chloroperlidae | 34 | 130.8 |
| | | | Nemouridae | 26 | 100.0 |
| | | | Perlodidae | 5 | 19.2 |
| | | Trichoptera | Rhyacophilidae | 1 | 3.8 |
| | | | Total | 302 | 1,161.5 |

Metrics

| Name | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|---|----------|---|
| Bray-Curtis Distance | 0.7 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 23.5 | 55.3 \pm 17.5 |
| % Predatores | 17.9 | 22.0 \pm 15.6 |
| % Scrapers | 67.2 | 53.5 \pm 23.0 |
| % Shredder | 8.6 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 9.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 18.5 | 20.6 \pm 17.1 |
| % EPT Individuals | 81.5 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 87.1 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 1.0 | 0.8 \pm 0.2 |
| Total Abundance | 1161.5 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 0.0 | 0.2 \pm 0.4 |
| Diptera taxa | 4.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 3.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 946.2 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 7.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.8 | 0.7 \pm 0.1 |
| Plecoptera taxa | 3.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 1.9 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.4 | 0.3 \pm 0.1 |
| Total No. of Taxa | 12.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 1.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.31 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.04 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.21 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.07 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.50 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.21 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.96 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.79 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.20 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.45 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.22 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.18 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halipidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.90 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.39 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.13 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.34 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.24 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.12 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.23 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.21 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.16 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.14 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.34 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.59 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.18 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.68 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.17 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.06 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.09 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.06 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.23 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.36 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.65 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.52 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.18 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.20 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 7.24 |
| RIVPACS : Observed taxa P>0.50 | 7.00 |
| RIVPACS : O:E (p > 0.5) | 0.97 |
| RIVPACS : Expected taxa P>0.70 | 4.30 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.16 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|--|-----------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-Avg (cm) | 18.7 | 28.5 \pm 10.6 |
| Depth-BankfullMinusWetted (cm) | 37.00 | 163.00 |
| Depth-Max (cm) | 28.0 | 44.5 \pm 18.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.55 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 0.89 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 16.5 | 85.0 \pm 66.5 |
| Width-Wetted (m) | 6.8 | 23.1 \pm 31.8 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.25000 | 16.49843 \pm 2.42987 |
| Landcover | | |
| Natl-SnowIce (%) | 26.42000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |
| %Cobble (%) | 8 | 63 \pm 4 |
| %Gravel (%) | 6 | 3 \pm 4 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|------------|--|
| %Pebble (%) | 86 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 3.50 | 6.67 \pm 3.25 |
| Dg (cm) | 3.4 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 2 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.43000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-DO (mg/L) | 8.2700000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 7.4 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 64.0000000 | 74.4000000 \pm 44.3472660 |
| General-TempAir (Degrees Celsius) | 14.8 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 11.7000000 | 5.7731579 \pm 1.9704316 |
| General-Turbidity (NTU) | 1.3400000 | 1.3000000 \pm 0.9899495 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-AQ11 |
| Sampling Date | Jul 25 2017 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12722 N, 122.97194 W |
| Altitude | 190 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st NatI-SnowIce NatI-Water NatI-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|---|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 8.4% | 4.4% | 40.7% | 20.5% | 16.4% | 9.7% |
| CABIN Assessment of RGD-AQ11 on Jul 25, 2017 | Mildly Divergent | | | | | |

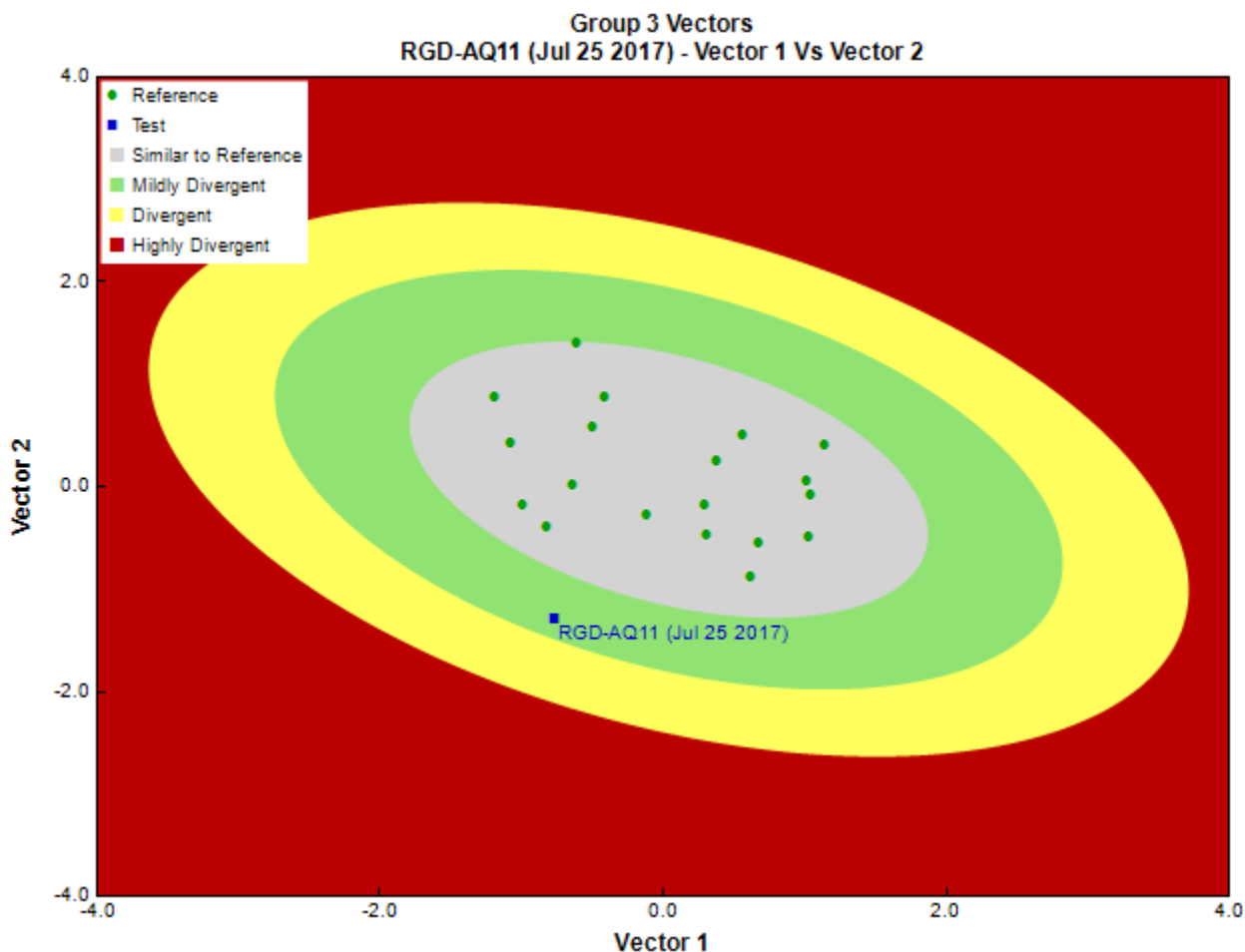


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|---|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | Karen Needham, Spencer Entomological Collecti |
| Date Taxonomy Completed | November 01, 2017 |
| | Marchant Box |
| Sub-Sample Proportion | 31/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|-----------|----------------|-----------------|-----------|-------------|
| Arthropoda | Arachnida | Trombidiformes | Hydrachnidae | 3 | 9.7 |
| | Insecta | Diptera | Ceratopogonidae | 3 | 9.7 |
| | | | Chironomidae | 18 | 58.1 |
| | | | Simuliidae | 46 | 148.4 |
| | | | Tipulidae | 1 | 3.2 |
| | | Ephemeroptera | Baetidae | 35 | 112.9 |
| | | | Ephemerellidae | 2 | 6.4 |
| | | | Heptageniidae | 205 | 661.3 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|-------|-------------|----------------|-----------|-------------|
| | | Plecoptera | Chloroperlidae | 23 | 74.2 |
| | | | Nemouridae | 2 | 6.4 |
| | | | Perlidae | 2 | 6.4 |
| | | | Perlodidae | 1 | 3.2 |
| | | Trichoptera | Rhyacophilidae | 4 | 12.9 |
| | | | Total | 345 | 1,112.8 |

Metrics

| Name | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|---|----------|---|
| Bray-Curtis Distance | 0.7 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 20.0 | 55.3 \pm 17.5 |
| % Predatores | 21.4 | 22.0 \pm 15.6 |
| % Scrapers | 82.9 | 53.5 \pm 23.0 |
| % Shredder | 0.9 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 9.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 20.6 | 20.6 \pm 17.1 |
| % EPT Individuals | 79.4 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 94.8 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.8 \pm 0.2 |
| Total Abundance | 1112.9 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 0.0 | 0.2 \pm 0.4 |
| Diptera taxa | 4.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 3.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 883.9 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 8.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.5 | 0.7 \pm 0.1 |
| Plecoptera taxa | 4.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 1.4 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.6 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 13.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 1.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.30 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.03 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.20 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.06 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.49 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.21 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.96 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.80 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.19 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.45 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.21 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.18 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.90 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.38 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.13 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.06 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.33 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.23 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.12 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.00 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.22 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.20 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.15 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.14 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.33 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.59 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecophrynidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.18 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.68 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.16 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.05 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.09 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.06 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.22 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.36 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.66 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.51 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.17 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.21 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.02 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 6.74 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 1.19 |
| RIVPACS : Expected taxa P>0.70 | 4.30 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.16 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|--|-----------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-Avg (cm) | 0.4 | 28.5 \pm 10.6 |
| Depth-BankfullMinusWetted (cm) | 40.00 | 163.00 |
| Depth-Max (cm) | 0.5 | 44.5 \pm 18.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 3 | 3 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 0 \pm 1 |
| Slope (m/m) | 0.0100000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 1 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.84 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 1.06 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 19.4 | 85.0 \pm 66.5 |
| Width-Wetted (m) | 9.2 | 23.1 \pm 31.8 |
| XSEC-VelInstrumentDirect (Category(1-3)) | 1 | 2 |
| XSEC-VelMethod (Category(1-3)) | 3 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.25000 | 16.49843 \pm 2.42987 |
| Landcover | | |
| Natl-SnowIce (%) | 26.42000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|------------|--|
| %Cobble (%) | 8 | 63 \pm 4 |
| %Gravel (%) | 12 | 3 \pm 4 |
| %Pebble (%) | 78 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 3.50 | 6.67 \pm 3.25 |
| Dg (cm) | 3.0 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 2 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.43000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-DO (mg/L) | 11.0200000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 7.1 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 50.5000000 | 74.4000000 \pm 44.3472660 |
| General-TempAir (Degrees Celsius) | 23.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 10.5000000 | 5.7731579 \pm 1.9704316 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-AQ11 |
| Sampling Date | Jul 31 2018 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12714 N, 122.97202 W |
| Altitude | 0 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|---|-----------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 9.0% | 4.1% | 26.5% | 15.5% | 38.2% | 6.5% |
| CABIN Assessment of RGD-AQ11 on Jul 31, 2018 | Divergent | | | | | |

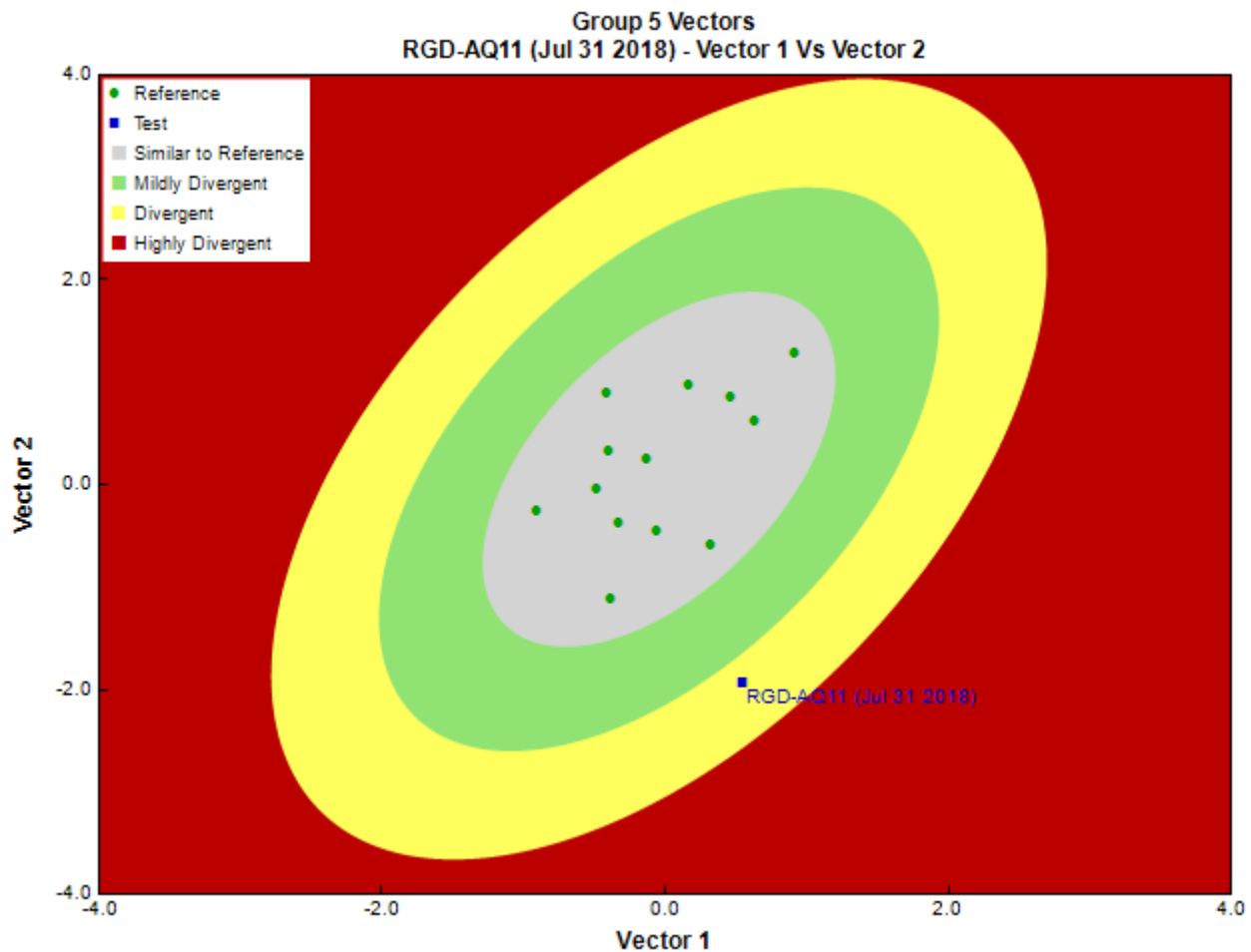


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| | - |
| Sub-Sample Proportion | 100/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|-----------------|-----------|-------------|
| Annelida | Clitellata | Lumbriculida | Lumbriculidae | 6 | 6.0 |
| Arthropoda | Arachnida | Sarcoptiformes | | 2 | 2.0 |
| | | Trombidiformes | | 2 | 2.0 |
| | | | Hydryphantidae | 1 | 1.0 |
| | | | Hygrobatidae | 8 | 8.0 |
| | | | Lebertiidae | 2 | 2.0 |
| | | | Sperchontidae | 8 | 8.0 |
| | | | Torrenticolidae | 1 | 1.0 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|----------|--------------|----------------|-----------------|-----------|-------------|
| | Collembola | Collembola | Sminthuridae | 2 | 2.0 |
| | Insecta | Diptera | Chironomidae | 27 | 27.0 |
| | | | Simuliidae | 282 | 282.0 |
| | | | Tipulidae | 1 | 1.0 |
| | | Ephemeroptera | Ameletidae | 31 | 31.0 |
| | | | Baetidae | 164 | 164.0 |
| | | | Ephemerellidae | 14 | 14.0 |
| | | | Heptageniidae | 119 | 119.0 |
| | | | Leptophlebiidae | 12 | 12.0 |
| | | Lepidoptera | | 1 | 1.0 |
| | | Plecoptera | Capniidae | 4 | 4.0 |
| | | | Chloroperlidae | 80 | 80.0 |
| | | | Leuctridae | 1 | 1.0 |
| | | | Nemouridae | 7 | 7.0 |
| | | | Perlidae | 51 | 51.0 |
| | | | Perlodidae | 11 | 11.0 |
| | | Trichoptera | Brachycentridae | 1 | 1.0 |
| | | | Rhyacophilidae | 6 | 6.0 |
| | Malacostraca | Amphipoda | Crangonyctidae | 1 | 1.0 |
| Mollusca | Gastropoda | Basommatophora | Planorbidae | 1 | 1.0 |
| | | | Total | 846 | 846.0 |

Metrics

| Name | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|---|----------|---|
| Bray-Curtis Distance | 0.94 | 0.6 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 11.5 \pm 10.5 |
| % Gatherers | 45.2 | 67.6 \pm 30.3 |
| % Predators | 47.0 | 41.1 \pm 20.2 |
| % Scrapers | 66.9 | 34.3 \pm 21.0 |
| % Shredder | 1.7 | 13.7 \pm 9.2 |
| No. Clinger Taxa | 31.0 | 13.0 \pm 5.7 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 40.2 | 47.4 \pm 26.3 |
| % EPT Individuals | 59.6 | 49.6 \pm 26.3 |
| % of 5 dominant taxa | 82.8 | 86.1 \pm 8.4 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.6 \pm 0.2 |
| Total Abundance | 846.0 | 13706.8 \pm 8626.5 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.0 |
| Coleoptera taxa | 0.0 | 0.5 \pm 0.5 |
| Diptera taxa | 3.0 | 2.6 \pm 1.2 |
| Ephemeroptera taxa | 5.0 | 3.4 \pm 1.2 |
| EPT Individuals (Sum) | 501.0 | 7446.2 \pm 6472.9 |
| EPT taxa (no) | 13.0 | 9.3 \pm 3.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.6 | 0.6 \pm 0.1 |
| Plecoptera taxa | 6.0 | 3.4 \pm 1.8 |
| Shannon-Wiener Diversity | 2.1 | 1.7 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.2 | 0.3 \pm 0.1 |
| Total No. of Taxa | 25.0 | 16.0 \pm 4.0 |
| Trichoptera taxa | 2.0 | 2.5 \pm 1.6 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.22 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.06 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.05 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.77 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.21 |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.15 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.46 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.30 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.97 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.67 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.03 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.00 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.20 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.36 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.24 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.77 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.06 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.14 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.81 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.36 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.16 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.02 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.05 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.01 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.39 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.24 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.25 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.00 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.22 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.18 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.16 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.23 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.22 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.04 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.00 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.01 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.47 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.49 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.06 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.03 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.16 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.58 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.00 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.27 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.12 |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.08 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.09 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.33 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.20 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.32 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.03 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.00 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.51 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.49 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.20 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.14 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.06 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.04 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 5.07 |
| RIVPACS : Observed taxa P>0.50 | 6.00 |
| RIVPACS : O:E (p > 0.5) | 1.18 |
| RIVPACS : Expected taxa P>0.70 | 3.32 |
| RIVPACS : Observed taxa P>0.70 | 4.00 |
| RIVPACS : O:E (p > 0.7) | 1.20 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|--|-----------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 15.90266 \pm 33.91726 |
| Channel | | |
| Depth-Avg (cm) | 20.7 | 40.5 \pm 22.4 |
| Macrophyte (PercentRange) | 0 | 1 \pm 2 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.23 \pm 0.44 |
| Reach-%Logging (PercentRange) | 0 | 0 \pm 0 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 2 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 0 \pm 0 |
| Reach-StraightRun (Binary) | 1 | 1 \pm 0 |
| Slope (m/m) | 0.0100000 | 0.0047331 \pm 0.0082050 |
| Veg-Coniferous (Binary) | 0 | 0 \pm 1 |
| Veg-Deciduous (Binary) | 1 | 0 \pm 1 |
| Veg-GrassesFerns (Binary) | 0 | 1 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|------------|--|
| Velocity-Avg (m/s) | 0.62 | 0.23 \pm 0.24 |
| Velocity-Max (m/s) | 1.13 | 0.31 \pm 0.35 |
| Width-Bankfull (m) | 18.1 | 75.1 \pm 72.8 |
| Width-Wetted (m) | 6.7 | 50.6 \pm 60.4 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 171.50745 \pm 107.47690 |
| Temp07_JULmax (Degrees Celsius) | 18.25000 | 20.34230 \pm 2.49485 |
| Landcover | | |
| Natl-SnowIce (%) | 26.42000 | 3.62533 \pm 10.17162 |
| Natl-Water (%) | 2.82000 | 1.80201 \pm 1.29922 |
| Natl-WetlandHerb (%) | 0.00000 | 0.68488 \pm 0.92347 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 |
| %Boulder (%) | 0 | 0 |
| %Cobble (%) | 2 | 58 |
| %Gravel (%) | 12 | 1 |
| %Pebble (%) | 86 | 41 |
| %Sand (%) | 0 | 0 |
| %Silt+Clay (%) | 0 | 0 |
| D50 (cm) | 2.75 | 3.30 |
| Dg (cm) | 2.6 | 6.6 |
| Dominant-1st (Category(0-9)) | 4 | 4 \pm 2 |
| Dominant-2nd (Category(0-9)) | 5 | 4 \pm 2 |
| Embeddedness (Category(1-5)) | 5 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 3 |
| SurroundingMaterial (Category(0-9)) | 3 | 2 \pm 1 |
| Topography | | |
| SlopeAvg (%) | 39.43000 | 30.12236 \pm 18.75100 |
| Water Chemistry | | |
| General-DO (mg/L) | 7.5000000 | 9.3400000 \pm 2.0171679 |
| General-pH (pH) | 7.2 | 6.8 \pm 1.0 |
| General-SpCond (μ S/cm) | 35.6000000 | 176.1000000 |
| General-TempAir (Degrees Celsius) | 34.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 15.5000000 | 13.2730769 \pm 4.7663725 |

Site Description

| | |
|--|---|
| Study Name | BC-Resort Municipality of Whistler-Ecosystem Monitoring |
| Site | RGD-AQ11 |
| Sampling Date | Jul 30 2019 |
| Know Your Watershed Basin | Harrison |
| Province / Territory | British Columbia |
| Terrestrial Ecological Classification | Pacific Maritime EcoZone Pacific Ranges EcoRegion |
| Coordinates (decimal degrees) | 50.12711 N, 122.97198 W |
| Altitude | 647 |
| Local Basin Name | River of Golden Dreams |
| | River of Golden Dreams |
| Stream Order | 3 |



Figure 1. Location Map

Across Reach (No image found)
 Aerial (No image found)
 Down Stream (No image found)
 Field Sheet (No image found)
 Miscellaneous (No image found)
 Substrate (No image found)
 Up Stream (No image found)

Cabin Assessment Results

| Reference Model Summary | |
|-----------------------------------|--|
| Model | Fraser River 2014 |
| Analysis Date | January 28, 2020 |
| Taxonomic Level | Family |
| Predictive Model Variables | Dominant-1st Natl-SnowIce Natl-Water Natl-WetlandHerb Precip02_FEB Reach-Riffles Sedimentary Slope SlopeAvg stream order Temp07_Julmax Width-Bankfull |
| Reference Groups | 123456 |

| | | | | | | |
|---|------------------|-------|-------|-------|-------|-------|
| Number of Reference Sites | 64 | 57 | 19 | 103 | 13 | 46 |
| Group Error Rate | 47.8% | 41.9% | 26.3% | 45.3% | 38.5% | 54.9% |
| Overall Model Error Rate | 45.3% | | | | | |
| Probability of Group Membership | 8.9% | 4.6% | 38.6% | 21.6% | 16.6% | 9.7% |
| CABIN Assessment of RGD-AQ11 on Jul 30, 2019 | Mildly Divergent | | | | | |

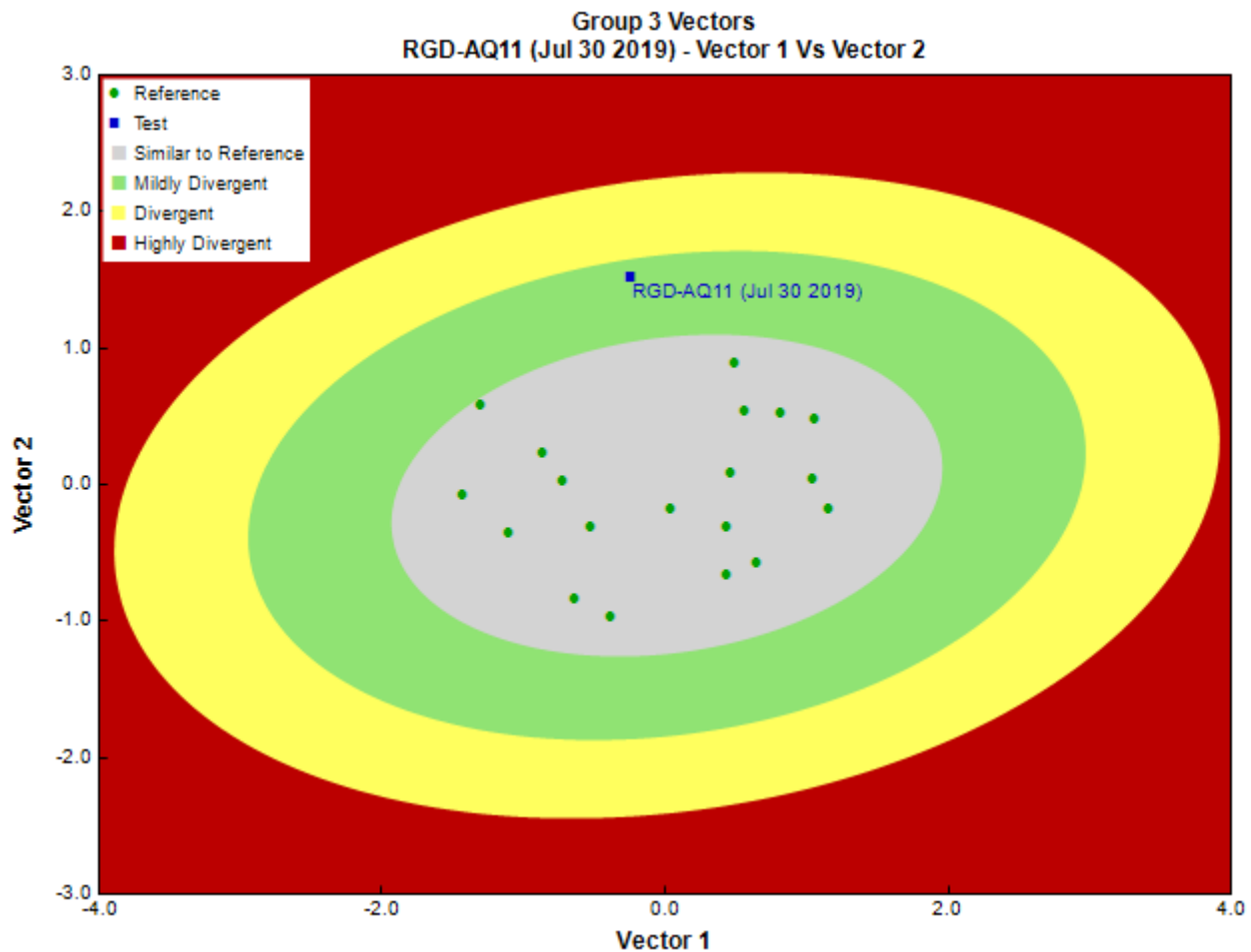


Figure 3. CABIN ordination assessment of the test site with the predicted group of reference sites. Each axis represents the relative abundance of the entire benthic invertebrate community with different organisms weighted differently on each axis.

Sample Information

| | |
|--------------------------------|----------|
| Sampling Device | Kick Net |
| Mesh Size | 400 |
| Sampling Time | 3 |
| Taxonomist | - |
| Date Taxonomy Completed | - |
| Sub-Sample Proportion | 27/100 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|------------|------------|----------------|---------------|-----------|-------------|
| Annelida | Clitellata | Tubificida | Naididae | 2 | 7.4 |
| Arthropoda | Arachnida | Trombidiformes | Hygrobatidae | 3 | 11.1 |
| | | | Sperchontidae | 6 | 22.2 |
| | Insecta | Coleoptera | Dytiscidae | 2 | 7.4 |
| | | Diptera | Chironomidae | 17 | 62.9 |
| | | | Simuliidae | 54 | 200.0 |
| | | | Tipulidae | 1 | 3.7 |
| | | Ephemeroptera | Ameletidae | 14 | 51.9 |

Community Structure

| Phylum | Class | Order | Family | Raw Count | Total Count |
|--------|--------------|-------------|-----------------|-----------|-------------|
| | | | Baetidae | 97 | 359.3 |
| | | | Ephemerellidae | 2 | 7.4 |
| | | | Heptageniidae | 69 | 255.5 |
| | | | Leptophlebiidae | 2 | 7.4 |
| | | Plecoptera | Capniidae | 5 | 18.5 |
| | | | Chloroperlidae | 28 | 103.7 |
| | | | Nemouridae | 5 | 18.5 |
| | | | Perlidae | 5 | 18.5 |
| | | | Perlodidae | 3 | 11.1 |
| | | Trichoptera | Rhyacophilidae | 1 | 3.7 |
| | Malacostraca | Amphipoda | Crangonyctidae | 4 | 14.8 |
| | | | Total | 320 | 1,185.0 |

Metrics

| Name | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|---|----------|---|
| Bray-Curtis Distance | 0.71 | 0.4 \pm 0.2 |
| Functional Measures | | |
| % Filterers | -- | 1.2 \pm 1.0 |
| % Gatherers | 32.2 | 55.3 \pm 17.5 |
| % Predatores | 28.4 | 22.0 \pm 15.6 |
| % Scrapers | 68.8 | 53.5 \pm 23.0 |
| % Shredder | 3.4 | 30.3 \pm 22.3 |
| No. Clinger Taxa | 23.0 | 13.6 \pm 4.4 |
| Number Of Individuals | | |
| % Diptera + Non-insects | 27.2 | 20.6 \pm 17.1 |
| % EPT Individuals | 72.2 | 78.2 \pm 17.8 |
| % of 5 dominant taxa | 82.8 | 86.1 \pm 8.2 |
| No. EPT individuals/Chironomids+EPT Individuals | 0.9 | 0.8 \pm 0.2 |
| Total Abundance | 1185.2 | 3776.0 \pm 2948.0 |
| Richness | | |
| Chironomidae taxa (genus level only) | 1.0 | 1.0 \pm 0.2 |
| Coleoptera taxa | 1.0 | 0.2 \pm 0.4 |
| Diptera taxa | 3.0 | 2.9 \pm 1.4 |
| Ephemeroptera taxa | 5.0 | 3.3 \pm 0.9 |
| EPT Individuals (Sum) | 855.6 | 2962.4 \pm 2556.9 |
| EPT taxa (no) | 11.0 | 9.8 \pm 2.6 |
| Odonata taxa | -- | 0.0 \pm 0.0 |
| Pielou's Evenness | 0.7 | 0.7 \pm 0.1 |
| Plecoptera taxa | 5.0 | 4.0 \pm 1.2 |
| Shannon-Wiener Diversity | 2.1 | 1.8 \pm 0.4 |
| Simpson's Diversity | 0.8 | 0.7 \pm 0.1 |
| Simpson's Evenness | 0.3 | 0.3 \pm 0.1 |
| Total No. of Taxa | 19.0 | 14.8 \pm 4.3 |
| Trichoptera taxa | 1.0 | 2.5 \pm 1.5 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Ameletidae | 29% | 26% | 26% | 46% | 0% | 69% | 0.31 |
| Ametropodidae | 0% | 0% | 0% | 0% | 0% | 6% | 0.01 |
| Anisitsiellidae | 0% | 10% | 0% | 0% | 0% | 4% | 0.01 |
| Apataniidae | 7% | 2% | 5% | 11% | 8% | 8% | 0.07 |
| Arrenuridae | 0% | 2% | 0% | 0% | 15% | 0% | 0.03 |
| Asellidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Athericidae | 0% | 8% | 0% | 9% | 8% | 0% | 0.03 |
| Aturidae | 4% | 6% | 0% | 5% | 0% | 2% | 0.02 |
| Baetidae | 97% | 92% | 79% | 93% | 62% | 84% | 0.82 |
| Blephariceridae | 1% | 0% | 0% | 2% | 0% | 2% | 0.01 |
| Brachycentridae | 42% | 53% | 5% | 35% | 15% | 27% | 0.21 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Caenidae | 0% | 0% | 0% | 1% | 38% | 0% | 0.07 |
| Capniidae | 81% | 60% | 37% | 65% | 31% | 69% | 0.50 |
| Ceratopogonidae | 23% | 23% | 5% | 27% | 54% | 14% | 0.21 |
| Chironomidae | 100% | 100% | 89% | 99% | 100% | 100% | 0.96 |
| Chloroperlidae | 87% | 61% | 95% | 92% | 31% | 76% | 0.80 |
| Crangonyctidae | 1% | 2% | 0% | 0% | 8% | 0% | 0.01 |
| Curculionidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Deuterophlebiidae | 0% | 0% | 0% | 1% | 0% | 2% | 0.00 |
| Dixidae | 1% | 6% | 0% | 1% | 0% | 2% | 0.01 |
| Dugesidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Dytiscidae | 0% | 2% | 0% | 1% | 0% | 4% | 0.01 |
| Elmidae | 48% | 73% | 5% | 28% | 15% | 18% | 0.20 |
| Empididae | 51% | 65% | 47% | 62% | 8% | 51% | 0.45 |
| Enchytraeidae | 22% | 26% | 5% | 34% | 31% | 39% | 0.21 |
| Ephemerellidae | 88% | 79% | 84% | 95% | 62% | 84% | 0.83 |
| Ephemeridae | 1% | 0% | 0% | 3% | 0% | 0% | 0.01 |
| Feltriidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Gammaridae | 7% | 6% | 0% | 0% | 0% | 0% | 0.01 |
| Gerridae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Glossiphoniidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Glossosomatidae | 32% | 21% | 16% | 34% | 0% | 12% | 0.18 |
| Gomphidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Halacaridae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Haliplidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Heptageniidae | 99% | 84% | 100% | 99% | 54% | 88% | 0.90 |
| Hyalellidae | 0% | 5% | 0% | 0% | 0% | 0% | 0.00 |
| Hydraenidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydrophilidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Hydropsychidae | 52% | 56% | 21% | 70% | 23% | 49% | 0.39 |
| Hydroptilidae | 19% | 39% | 5% | 15% | 23% | 6% | 0.13 |
| Hydrozetidae | 1% | 3% | 5% | 4% | 0% | 2% | 0.03 |
| Hydryphantidae | 10% | 10% | 5% | 12% | 0% | 6% | 0.07 |
| Hygrobatidae | 6% | 5% | 5% | 8% | 8% | 12% | 0.07 |
| Hypogastruridae | 0% | 0% | 0% | 3% | 0% | 12% | 0.02 |
| Isotomidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Lebertiidae | 38% | 39% | 16% | 45% | 54% | 39% | 0.34 |
| Lepidostomatidae | 39% | 52% | 5% | 46% | 23% | 18% | 0.23 |
| Leptoceridae | 1% | 15% | 0% | 5% | 62% | 2% | 0.12 |
| Leptohyphidae | 0% | 3% | 0% | 2% | 0% | 0% | 0.01 |
| Leptophlebiidae | 54% | 71% | 5% | 39% | 15% | 14% | 0.22 |
| Leuctridae | 35% | 26% | 11% | 39% | 8% | 25% | 0.21 |
| Limnephilidae | 36% | 24% | 21% | 26% | 0% | 39% | 0.22 |
| Limnesiidae | 7% | 13% | 5% | 9% | 46% | 25% | 0.15 |
| Lumbriculidae | 4% | 10% | 0% | 15% | 46% | 22% | 0.14 |
| Lymnaeidae | 1% | 3% | 0% | 2% | 8% | 2% | 0.02 |
| Margaritiferidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Mideopsidae | 1% | 5% | 0% | 0% | 0% | 2% | 0.01 |
| Muscidae | 0% | 2% | 0% | 4% | 0% | 6% | 0.02 |
| Naididae | 26% | 47% | 5% | 46% | 85% | 33% | 0.34 |
| Nemouridae | 93% | 73% | 53% | 81% | 15% | 73% | 0.59 |
| Oxidae | 0% | 2% | 0% | 1% | 15% | 0% | 0.03 |
| Pelecorhynchidae | 3% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Peltoperlidae | 1% | 0% | 5% | 7% | 0% | 4% | 0.04 |
| Perlidae | 32% | 29% | 11% | 32% | 8% | 20% | 0.18 |
| Perlodidae | 64% | 60% | 79% | 75% | 31% | 76% | 0.68 |
| Philopotamidae | 1% | 3% | 0% | 1% | 0% | 2% | 0.01 |
| Physidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pionidae | 0% | 2% | 0% | 3% | 0% | 0% | 0.01 |
| Piscicolidae | 0% | 0% | 0% | 1% | 0% | 0% | 0.00 |
| Pisidiidae | 16% | 53% | 0% | 11% | 54% | 12% | 0.16 |
| Planorbidae | 1% | 3% | 0% | 1% | 31% | 0% | 0.06 |

Frequency and Probability of Taxa Occurrence

| Reference Model Taxa | Frequency of Occurrence in Reference Sites | | | | | | Probability Of Occurrence at RGD-AQ11 |
|----------------------|--|---------|---------|---------|---------|---------|---------------------------------------|
| | Group 1 | Group 2 | Group 3 | Group 4 | Group 5 | Group 6 | |
| Poduridae | 0% | 3% | 0% | 0% | 0% | 2% | 0.00 |
| Polycentropodidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Psychodidae | 33% | 21% | 0% | 22% | 0% | 4% | 0.09 |
| Pteronarcyidae | 7% | 8% | 0% | 9% | 15% | 6% | 0.06 |
| Rhyacophilidae | 59% | 32% | 47% | 66% | 0% | 61% | 0.45 |
| Sialidae | 0% | 2% | 0% | 0% | 0% | 0% | 0.00 |
| Simuliidae | 52% | 29% | 16% | 32% | 8% | 24% | 0.23 |
| Sperchontidae | 17% | 34% | 37% | 50% | 23% | 45% | 0.36 |
| Staphylinidae | 0% | 0% | 0% | 0% | 0% | 2% | 0.00 |
| Stratiomyidae | 0% | 0% | 0% | 1% | 8% | 2% | 0.02 |
| Stygothrombidiidae | 3% | 0% | 0% | 4% | 0% | 2% | 0.01 |
| Tabanidae | 0% | 0% | 0% | 1% | 0% | 4% | 0.01 |
| Taeniopterygidae | 59% | 16% | 100% | 51% | 15% | 71% | 0.65 |
| Tanyderidae | 0% | 0% | 0% | 3% | 0% | 6% | 0.01 |
| Thaumaleidae | 0% | 0% | 0% | 2% | 0% | 0% | 0.00 |
| Tipulidae | 67% | 50% | 42% | 68% | 38% | 63% | 0.52 |
| Torrenticolidae | 32% | 58% | 0% | 36% | 23% | 6% | 0.18 |
| Trhypochthoniidae | 1% | 5% | 0% | 3% | 0% | 0% | 0.01 |
| Uenoidae | 9% | 8% | 37% | 17% | 0% | 10% | 0.20 |
| Unionicolidae | 0% | 0% | 0% | 0% | 15% | 0% | 0.03 |
| Valvatidae | 3% | 2% | 0% | 3% | 8% | 2% | 0.03 |

RIVPACS Ratios

| | |
|--------------------------------|------|
| RIVPACS : Expected taxa P>0.50 | 6.74 |
| RIVPACS : Observed taxa P>0.50 | 8.00 |
| RIVPACS : O:E (p > 0.5) | 1.19 |
| RIVPACS : Expected taxa P>0.70 | 4.30 |
| RIVPACS : Observed taxa P>0.70 | 5.00 |
| RIVPACS : O:E (p > 0.7) | 1.16 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|--|-----------|---|
| Bedrock Geology | | |
| Sedimentary (%) | 2.56000 | 18.33344 \pm 33.50703 |
| Channel | | |
| Depth-Avg (cm) | 23.4 | 28.5 \pm 10.6 |
| Depth-BankfullMinusWetted (cm) | 33.00 | 163.00 |
| Depth-Max (cm) | 31.0 | 44.5 \pm 18.9 |
| Macrophyte (PercentRange) | 0 | 0 \pm 0 |
| Reach-%CanopyCoverage (PercentRange) | 1.00 | 0.16 \pm 0.37 |
| Reach-DomStreamsideVeg (Category(1-4)) | 2 | 3 |
| Reach-Pools (Binary) | 1 | 0 \pm 0 |
| Reach-Rapids (Binary) | 0 | 0 \pm 0 |
| Reach-Riffles (Binary) | 1 | 1 \pm 0 |
| Reach-StraightRun (Binary) | 0 | 0 \pm 1 |
| Slope (m/m) | 0.0100000 | 0.0259896 \pm 0.0313728 |
| Veg-Coniferous (Binary) | 1 | 1 \pm 0 |
| Veg-Deciduous (Binary) | 1 | 1 \pm 1 |
| Veg-GrassesFerns (Binary) | 0 | 0 \pm 0 |
| Veg-Shrubs (Binary) | 1 | 1 \pm 0 |
| Velocity-Avg (m/s) | 0.80 | 0.49 \pm 0.15 |
| Velocity-Max (m/s) | 1.17 | 0.67 \pm 0.21 |
| Width-Bankfull (m) | 17.3 | 85.0 \pm 66.5 |
| Width-Wetted (m) | 6.6 | 23.1 \pm 31.8 |
| XSEC-VelMethod (Category(1-3)) | 1 | 3 |
| Climate | | |
| Precip02_FEB (mm) | 155.11000 | 127.54903 \pm 58.24882 |
| Temp07_JULmax (Degrees Celsius) | 18.25000 | 16.49843 \pm 2.42987 |

Habitat Description

| Variable | RGD-AQ11 | Predicted Group Reference Mean \pm SD |
|-------------------------------------|------------|--|
| Landcover | | |
| Natl-SnowIce (%) | 26.42000 | 30.72486 \pm 23.89539 |
| Natl-Water (%) | 2.82000 | 0.99760 \pm 0.86372 |
| Natl-WetlandHerb (%) | 0.00000 | 0.02638 \pm 0.03974 |
| Substrate Data | | |
| %Bedrock (%) | 0 | 0 \pm 0 |
| %Boulder (%) | 0 | 9 \pm 8 |
| %Cobble (%) | 4 | 63 \pm 4 |
| %Gravel (%) | 19 | 3 \pm 4 |
| %Pebble (%) | 77 | 25 \pm 7 |
| %Sand (%) | 0 | 0 \pm 0 |
| %Silt+Clay (%) | 0 | 0 \pm 0 |
| D50 (cm) | 3.00 | 6.67 \pm 3.25 |
| Dg (cm) | 2.7 | 8.6 \pm 1.6 |
| Dominant-1st (Category(0-9)) | 5 | 7 \pm 1 |
| Dominant-2nd (Category(0-9)) | 4 | 7 \pm 1 |
| Embeddedness (Category(1-5)) | 4 | 4 \pm 1 |
| PeriphytonCoverage (Category(1-5)) | 1 | 2 \pm 1 |
| SurroundingMaterial (Category(0-9)) | 3 | 4 \pm 2 |
| Topography | | |
| SlopeAvg (%) | 39.43000 | 41.69956 \pm 6.13915 |
| Water Chemistry | | |
| General-Conductivity (μ S/cm) | 33.3000000 | 62.9529406 \pm 33.2341330 |
| General-DO (mg/L) | 9.8100000 | 12.6052631 \pm 1.2122173 |
| General-pH (pH) | 6.8 | 7.4 \pm 0.4 |
| General-SpCond (μ S/cm) | 44.3000000 | 74.4000000 \pm 44.3472660 |
| General-TempAir (Degrees Celsius) | 12.0 | 0.0 \pm 0.0 |
| General-TempWater (Degrees Celsius) | 12.8000000 | 5.7731579 \pm 1.9704316 |



Project: 16025 Whistler 2018

Palmer Environmental Group, Alyssa Murdoch, May Mason Irene Mencke,

Taxonomist: Scott Finlayson

scottfinlayson@cordilleraconsulting.ca

250-494-7553

| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
|-------------------------------|-------------|-------------|-------------|-----------|-------------|
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| Phylum: Arthropoda | 0 | 0 | 0 | 0 | 0 |
| Order: Collembola | 0 | 0 | 0 | 0 | 0 |
| Family: Sminthuridae | 0 | 0 | 0 | 2 | 0 |
| | | | | | |
| Subphylum: Hexapoda | 0 | 0 | 0 | 0 | 0 |
| Class: Insecta | 0 | 0 | 0 | 0 | 0 |
| Order: Ephemeroptera | 0 | 0 | 0 | 0 | 0 |
| Family: Ameletidae | 0 | 0 | 0 | 0 | 0 |
| <u>Ameletus</u> | 28 | 6 | 0 | 31 | 14 |
| Family: Baetidae | 217 | 12 | 380 | 23 | 11 |
| <u>Baetis</u> | 128 | 35 | 680 | 58 | 32 |
| <u>Baetis rhodani group</u> | 17 | 94 | 100 | 81 | 34 |
| <u>Baetis bicaudatus</u> | 0 | 0 | 0 | 2 | 5 |
| <u>Centroptilum</u> | 0 | 0 | 0 | 0 | 0 |
| <u>Anafroptilum</u> | 0 | 0 | 0 | 0 | 14 |
| <u>Dipheter hageni</u> | 0 | 0 | 0 | 0 | 0 |
| Family: Ephemerellidae | 6 | 6 | 0 | 3 | 18 |
| <u>Caudatella</u> | 0 | 0 | 0 | 0 | 0 |
| <u>Drunella</u> | 0 | 0 | 0 | 0 | 0 |
| <u>Drunella grandis group</u> | 0 | 0 | 0 | 0 | 16 |
| <u>Drunella doddsii</u> | 33 | 0 | 0 | 9 | 5 |

| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
|---------------------------------|-------------|-------------|-------------|-----------|-------------|
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| <i>Drunella spinifera</i> | 11 | 0 | 10 | 1 | 57 |
| <i>Serratella</i> | 0 | 18 | 0 | 1 | 0 |
| Family: Heptageniidae | 28 | 0 | 40 | 20 | 0 |
| <i>Cinygmula</i> | 117 | 0 | 0 | 30 | 0 |
| <i>Epeorus</i> | 239 | 0 | 0 | 46 | 0 |
| <i>Rhithrogena</i> | 28 | 0 | 0 | 23 | 0 |
| Family: Leptophlebiidae | 6 | 53 | 240 | 12 | 34 |
| | | | | | |
| Order: Plecoptera | 0 | 0 | 0 | 0 | 0 |
| Family: Capniidae | 6 | 0 | 0 | 4 | 0 |
| Family: Chloroperlidae | 0 | 0 | 10 | 7 | 2 |
| <i>Neaviperla</i> | 0 | 0 | 0 | 1 | 0 |
| <i>Paraperla</i> | 0 | 0 | 0 | 5 | 0 |
| <i>Suwallia</i> | 0 | 0 | 0 | 5 | 2 |
| <i>Sweltsa</i> | 128 | 6 | 90 | 62 | 11 |
| Family: Leuctridae | 0 | 0 | 0 | 1 | 0 |
| <i>Paraleuctra</i> | 6 | 0 | 0 | 0 | 0 |
| Family: Nemouridae | 0 | 0 | 0 | 0 | 0 |
| <i>Malenka</i> | 0 | 0 | 10 | 1 | 5 |
| <i>Zapada</i> | 6 | 153 | 630 | 4 | 30 |
| <i>Zapada oregonensis group</i> | 0 | 0 | 10 | 0 | 0 |
| <i>Zapada cinctipes</i> | 6 | 82 | 130 | 0 | 5 |
| <i>Zapada columbiana</i> | 11 | 0 | 0 | 2 | 0 |
| Family: Perlidae | 78 | 12 | 0 | 47 | 0 |
| <i>Doroneuria</i> | 17 | 0 | 0 | 3 | 0 |
| <i>Hesperoperla</i> | 0 | 18 | 0 | 1 | 0 |
| Family: Perlodidae | 28 | 0 | 10 | 3 | 0 |
| <i>Megarcys</i> | 11 | 0 | 0 | 8 | 0 |
| | | | | | |
| Order: Trichoptera | 0 | 0 | 0 | 0 | 0 |
| Family: Brachycentridae | 0 | 0 | 0 | 0 | 0 |

| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
|--|-------------|-------------|-------------|-----------|-------------|
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| <i>Micrasema</i> | 0 | 0 | 0 | 1 | 0 |
| Family: Hydropsychidae | 0 | 12 | 0 | 0 | 0 |
| <i>Arctopsyche</i> | 0 | 0 | 0 | 0 | 0 |
| Family: Hydroptilidae | 0 | 0 | 0 | 0 | 0 |
| <i>Oxyethira</i> | 6 | 0 | 0 | 0 | 0 |
| Family: Lepidostomatidae | 0 | 0 | 0 | 0 | 0 |
| <i>Lepidostoma</i> | 0 | 12 | 0 | 0 | 2 |
| Family: Limnephilidae | 0 | 0 | 0 | 0 | 0 |
| <i>Dicosmoecus</i> | 0 | 0 | 0 | 0 | 2 |
| <i>Onocosmoecus</i> | 0 | 0 | 50 | 0 | 2 |
| Family: Rhyacophilidae | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila</i> | 22 | 18 | 10 | 3 | 5 |
| <i>Rhyacophila angelita group</i> | 0 | 6 | 10 | 0 | 0 |
| <i>Rhyacophila betteni group</i> | 0 | 0 | 0 | 1 | 0 |
| <i>Rhyacophila brunnea/vemna group</i> | 17 | 0 | 0 | 1 | 0 |
| <i>Rhyacophila hyalinata group</i> | 0 | 0 | 0 | 0 | 0 |
| <i>Rhyacophila vagrita group</i> | 6 | 0 | 0 | 1 | 0 |
| <i>Rhyacophila arnaudi</i> | 0 | 0 | 30 | 0 | 0 |
| | | | | | |
| Order: Coleoptera | 0 | 0 | 0 | 0 | 0 |
| Family: Dytiscidae | 0 | 0 | 0 | 0 | 0 |
| <i>Oreodytes</i> | 0 | 0 | 0 | 0 | 39 |
| Subfamily: Hydroporinae | 0 | 0 | 0 | 0 | 39 |
| | | | | | |
| Order: Diptera | 0 | 0 | 0 | 0 | 0 |
| Family: Ceratopogonidae | 0 | 0 | 0 | 0 | 0 |
| <i>Bezzia/ Palpomyia</i> | 39 | 0 | 0 | 0 | 2 |
| <i>Mallochohelea</i> | 6 | 0 | 0 | 0 | 0 |
| Family: Chironomidae | 17 | 24 | 30 | 6 | 43 |
| Subfamily: Chironominae | 0 | 0 | 0 | 0 | 0 |
| Tribe: Chironomini | 0 | 0 | 0 | 0 | 0 |

| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
|----------------------------------|-------------|-------------|-------------|-----------|-------------|
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| <i>Polypedilum</i> | 17 | 0 | 0 | 1 | 2 |
| <i>Saetheria</i> | 0 | 0 | 0 | 0 | 20 |
| Tribe: Tanytarsini | 17 | 0 | 0 | 0 | 0 |
| <i>Micropsectra</i> | 22 | 59 | 30 | 6 | 14 |
| <i>Rheotanytarsus</i> | 0 | 0 | 0 | 0 | 0 |
| <i>Stempellinella</i> | 0 | 0 | 0 | 0 | 2 |
| Subfamily: Orthocladiinae | 6 | 6 | 20 | 0 | 0 |
| <i>Brillia</i> | 0 | 29 | 20 | 0 | 0 |
| <i>Eukiefferiella</i> | 22 | 53 | 150 | 3 | 0 |
| <i>Heterotrissocladius</i> | 0 | 0 | 0 | 0 | 2 |
| <i>Hydrobaenus</i> | 6 | 0 | 0 | 0 | 0 |
| <i>Metriocnemus</i> | 0 | 0 | 0 | 1 | 0 |
| <i>Orthocladius complex</i> | 0 | 0 | 30 | 2 | 5 |
| <i>Parakiefferiella</i> | 0 | 0 | 0 | 0 | 9 |
| <i>Parametriocnemus</i> | 11 | 0 | 0 | 2 | 2 |
| <i>Psectrocladius</i> | 0 | 0 | 0 | 0 | 2 |
| <i>Rheocricotopus</i> | 0 | 0 | 0 | 0 | 9 |
| <i>Thienemanniella</i> | 0 | 0 | 30 | 0 | 7 |
| <i>Tvetenia</i> | 0 | 100 | 130 | 1 | 30 |
| Subfamily: Tanypodinae | 0 | 0 | 0 | 0 | 0 |
| <i>Zavrelimyia</i> | 0 | 0 | 0 | 2 | 0 |
| Tribe: Pentaneurini | 0 | 0 | 0 | 0 | 0 |
| <i>Thienemannimyia group</i> | 28 | 0 | 0 | 3 | 41 |
| Family: Deuterophlebiidae | 0 | 0 | 0 | 0 | 0 |
| <i>Deuterophlebia</i> | 6 | 0 | 0 | 0 | 0 |
| Family: Empididae | 0 | 0 | 10 | 0 | 0 |
| <i>Chelifera/ Metachela</i> | 0 | 0 | 30 | 0 | 2 |
| <i>Oreogeton</i> | 0 | 0 | 0 | 0 | 0 |
| Family: Simuliidae | 11 | 41 | 20 | 5 | 0 |
| <i>Prosimulium</i> | 0 | 0 | 0 | 1 | 0 |
| <i>Prosimulium/Helodon</i> | 0 | 0 | 0 | 1 | 0 |

| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
|-------------------------|-------------|-------------|-------------|-----------|-------------|
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| <u>Simulium</u> | 478 | 1271 | 120 | 275 | 14 |
| Family: Tipulidae | 0 | 0 | 0 | 0 | 2 |
| <u>Dicranota</u> | 0 | 0 | 0 | 1 | 2 |
| <u>Erioptera</u> | 0 | 0 | 20 | 0 | 0 |
| <u>Hexatoma</u> | 0 | 6 | 20 | 0 | 0 |
| | | | | | |
| Order: Lepidoptera | 0 | 0 | 0 | 1 | 0 |
| Order: Megaloptera | 0 | 0 | 0 | 0 | 0 |
| Family: Sialidae | 0 | 0 | 0 | 0 | 0 |
| <u>Sialis</u> | 0 | 0 | 0 | 0 | 2 |
| | | | | | |
| Subphylum: Chelicerata | 0 | 0 | 0 | 0 | 0 |
| Class: Arachnida | 0 | 0 | 0 | 0 | 0 |
| Order: Trombidiformes | 0 | 0 | 0 | 2 | 2 |
| Family: Aturidae | 0 | 0 | 0 | 0 | 0 |
| <u>Aturus</u> | 0 | 0 | 10 | 0 | 0 |
| Family: Hydryphantidae | 0 | 0 | 0 | 0 | 0 |
| <u>Protzia</u> | 0 | 0 | 10 | 1 | 0 |
| Family: Hygrobatidae | 0 | 0 | 0 | 0 | 0 |
| <u>Atractides</u> | 33 | 0 | 10 | 8 | 2 |
| <u>Hygrobates</u> | 0 | 0 | 20 | 0 | 20 |
| Family: Lebertiidae | 0 | 0 | 0 | 0 | 0 |
| <u>Lebertia</u> | 11 | 0 | 10 | 2 | 9 |
| Family: Sperchontidae | 0 | 0 | 0 | 0 | 0 |
| <u>Sperchon</u> | 22 | 0 | 10 | 8 | 5 |
| <u>Sperchonopsis</u> | 0 | 6 | 0 | 0 | 0 |
| Family: Torrenticolidae | 0 | 0 | 0 | 0 | 0 |
| <u>Testudacarus</u> | 6 | 0 | 10 | 1 | 0 |
| <u>Torrenticola</u> | 0 | 0 | 0 | 0 | 0 |
| | | | | | |
| Order: Sarcoptiformes | 0 | 0 | 0 | 0 | 0 |

| | | | | | |
|--|-------------|-------------|-------------|-----------|-------------|
| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| Order: Oribatida | 0 | 0 | 0 | 2 | 0 |
| | | | | | |
| Class: Malacostraca | 0 | 0 | 0 | 0 | 0 |
| Order: Amphipoda | 0 | 0 | 0 | 0 | 2 |
| Family: Crangonyctidae | 0 | 0 | 0 | 0 | 0 |
| <i>Crangonyx</i> | 0 | 0 | 0 | 1 | 2 |
| | | | | | |
| Phylum: Mollusca | 0 | 0 | 0 | 0 | 0 |
| Class: Bivalvia | 0 | 0 | 0 | 0 | 0 |
| Order: Veneroida | 0 | 0 | 0 | 0 | 0 |
| Family: Pisidiidae | 0 | 6 | 0 | 0 | 2 |
| <i>Pisidium</i> | 0 | 6 | 0 | 0 | 5 |
| | | | | | |
| Class: Gastropoda | 0 | 0 | 0 | 0 | 0 |
| Order: Basommatophora | 0 | 0 | 0 | 0 | 0 |
| Family: Planorbidae | 0 | 0 | 0 | 1 | 0 |
| | | | | | |
| Phylum: Annelida | 0 | 0 | 0 | 0 | 0 |
| Subphylum: Clitellata | 0 | 0 | 0 | 0 | 0 |
| Class: Oligochaeta | 0 | 0 | 0 | 0 | 0 |
| Order: Lumbriculida | 0 | 0 | 0 | 0 | 0 |
| Family: Lumbriculidae | 22 | 0 | 0 | 0 | 0 |
| <i>Lumbriculus</i> | 6 | 0 | 0 | 6 | 9 |
| | | | | | |
| Order: Tubificida | 0 | 0 | 0 | 0 | 0 |
| Family: Naididae | 0 | 0 | 0 | 0 | 0 |
| <i>Nais</i> | 0 | 0 | 10 | 0 | 0 |
| Subfamily: Tubificinae with hair chaetae | 0 | 0 | 0 | 0 | 161 |
| Totals: | 1992 | 2150 | 3190 | 846 | 815 |
| | | | | | |
| Taxa present but not included: | | | | | |

| | | | | | |
|--------------------------------|--------------------|--------------------|--------------------|------------------|--------------------|
| Year: | 2018 | 2018 | 2018 | 2018 | 2018 |
| Sample: | 21M-DS-AQ21 | JOR-DS-AQ31 | CRB-DS-AQ01 | RGD-AQ11 | RDG-DS-AQ12 |
| Sample Collection Date: | 31-Jul-18 | 01-Aug-18 | 01-Aug-18 | 31-Jul-18 | 01-Aug-18 |
| CC#: | CC191659 | CC191661 | CC191663 | CC191664 | CC191665 |
| | | | | | |
| <i>Terrestrials</i> | 0 | 0 | 0 | 0 | 2 |
| | | | | | |
| Phylum: Arthropoda | 0 | 0 | 0 | 0 | 0 |
| Subphylum: Hexapoda | 0 | 0 | 0 | 0 | 0 |
| Class: Insecta | 0 | 0 | 0 | 0 | 0 |
| Order: Psocodea | 0 | 0 | 0 | 1 | 0 |
| | | | | | |
| Subphylum: Crustacea | 0 | 0 | 0 | 0 | 0 |
| Class: Ostracoda | 0 | 0 | 0 | 1 | 0 |
| Class: Maxillipoda | 0 | 0 | 0 | 0 | 0 |
| Class: Copepoda | 0 | 0 | 10 | 0 | 0 |
| | | | | | |
| Phylum: Nemata | 0 | 6 | 0 | 1 | 2 |
| Phylum: Platyhelminthes | 0 | 0 | 0 | 0 | 0 |
| Class: Turbellaria | 0 | 6 | 0 | 1 | 0 |
| Totals: | 0 | 12 | 10 | 4 | 4 |



Project: Whistler 160255 2019

Palmer Environmental Group, Alyssa Murdoch, May Mason Irene Mencke,

Taxonomist: Scott Finlayson

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250-494-7553

| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
|---------------------------------|-----------|-------------|-------------|-------------|-------------|
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| Phylum: Arthropoda | 0 | 0 | 0 | 0 | 0 |
| Order: Collembola | 0 | 10 | 0 | 3 | 6 |
| | | | | | |
| Subphylum: Hexapoda | 0 | 0 | 0 | 0 | 0 |
| Class: Insecta | 0 | 0 | 0 | 0 | 0 |
| Order: Ephemeroptera | 0 | 0 | 0 | 0 | 0 |
| Family: Ameletidae | 0 | 0 | 0 | 0 | 0 |
| <u>Ameletus</u> | 52 | 0 | 4 | 0 | 89 |
| Family: Baetidae | 63 | 530 | 26 | 34 | 111 |
| <u>Baetis</u> | 78 | 1040 | 35 | 154 | 139 |
| <u>Baetis rhodani group</u> | 219 | 40 | 9 | 20 | 178 |
| <u>Baetis bicaudatus</u> | 0 | 0 | 13 | 0 | 28 |
| <u>Centroptilum</u> | 0 | 0 | 0 | 0 | 0 |
| <u>Anafroptilum</u> | 0 | 0 | 26 | 0 | 0 |
| <u>Dipheter hageni</u> | 0 | 0 | 0 | 3 | 0 |
| Family: Ephemerellidae | 0 | 0 | 100 | 11 | 17 |
| <u>Drunella</u> | 0 | 0 | 0 | 0 | 0 |
| <u>Drunella grandis group</u> | 4 | 10 | 109 | 6 | 0 |
| <u>Drunella coloradensis</u> | 0 | 0 | 0 | 0 | 6 |
| <u>Drunella doddsii</u> | 0 | 0 | 0 | 0 | 22 |
| <u>Drunella spinifera</u> | 4 | 0 | 4 | 0 | 17 |

| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
|---------------------------------|-----------|-------------|-------------|-------------|-------------|
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| <u>Ephemerella</u> | 0 | 0 | 17 | 0 | 0 |
| <u>Serratella</u> | 0 | 0 | 9 | 3 | 0 |
| Family: Heptageniidae | 48 | 10 | 4 | 0 | 67 |
| <u>Cinygmula</u> | 74 | 0 | 0 | 0 | 139 |
| <u>Epeorus</u> | 107 | 0 | 0 | 0 | 283 |
| <u>Rhithrogena</u> | 26 | 0 | 0 | 0 | 6 |
| Family: Leptophlebiidae | 7 | 140 | 17 | 6 | 11 |
| | | | | | |
| Order: Plecoptera | 0 | 0 | 0 | 0 | 0 |
| Family: Capniidae | 19 | 0 | 0 | 0 | 11 |
| Family: Chloroperlidae | 22 | 20 | 0 | 0 | 22 |
| <u>Sweltsa</u> | 81 | 30 | 13 | 3 | 56 |
| Family: Nemouridae | 0 | 90 | 0 | 3 | 0 |
| <u>Malenka</u> | 4 | 10 | 4 | 0 | 6 |
| <u>Zapada</u> | 11 | 440 | 4 | 80 | 33 |
| <u>Zapada oregonensis group</u> | 0 | 50 | 0 | 0 | 0 |
| <u>Zapada cinctipes</u> | 0 | 180 | 9 | 54 | 6 |
| <u>Zapada columbiana</u> | 4 | 0 | 0 | 0 | 0 |
| Family: Perlidae | 11 | 0 | 0 | 6 | 17 |
| <u>Calineuria californica</u> | 4 | 0 | 0 | 0 | 0 |
| <u>Doroneuria</u> | 4 | 0 | 0 | 0 | 11 |
| <u>Hesperoperla</u> | 0 | 0 | 0 | 3 | 0 |
| Family: Perlodidae | 7 | 0 | 0 | 0 | 17 |
| <u>Megarcys</u> | 4 | 0 | 0 | 0 | 17 |
| | | | | | |
| Order: Trichoptera | 0 | 10 | 0 | 0 | 0 |
| Family: Glossosomatidae | 0 | 0 | 0 | 0 | 0 |
| <u>Glossosoma</u> | 0 | 0 | 0 | 3 | 0 |
| Family: Hydropsychidae | 0 | 0 | 0 | 3 | 6 |
| Family: Lepidostomatidae | 0 | 0 | 0 | 0 | 0 |
| <u>Lepidostoma</u> | 0 | 0 | 0 | 11 | 0 |

| | | | | | |
|--|-----------|-------------|-------------|-------------|-------------|
| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| Family: Limnephilidae | 0 | 0 | 0 | 0 | 11 |
| <u>Onocosmoecus</u> | 0 | 0 | 4 | 0 | 0 |
| <u>Psychoglypha</u> | 0 | 0 | 4 | 0 | 0 |
| Family: Philopotamidae | 0 | 0 | 0 | 0 | 0 |
| <u>Wormaldia</u> | 0 | 0 | 0 | 3 | 0 |
| Family: Rhyacophilidae | 0 | 0 | 0 | 0 | 0 |
| <u>Rhyacophila</u> | 0 | 20 | 4 | 0 | 6 |
| <u>Rhyacophila angelita group</u> | 4 | 0 | 0 | 0 | 0 |
| <u>Rhyacophila brunnea/vemna group</u> | 0 | 0 | 0 | 0 | 11 |
| <u>Rhyacophila arnaudi</u> | 0 | 20 | 0 | 0 | 0 |
| | | | | | |
| Order: Coleoptera | 0 | 0 | 0 | 0 | 0 |
| Family: Dytiscidae | 0 | 0 | 0 | 0 | 0 |
| <u>Oreodytes</u> | 7 | 0 | 83 | 0 | 0 |
| Subfamily: Hydroporinae | 0 | 0 | 78 | 0 | 0 |
| | | | | | |
| Order: Diptera | 0 | 0 | 0 | 0 | 0 |
| Family: Ceratopogonidae | 0 | 0 | 0 | 0 | 0 |
| <u>Bezzia/ Palpomyia</u> | 0 | 0 | 0 | 0 | 11 |
| Family: Chironomidae | 4 | 30 | 135 | 11 | 28 |
| Subfamily: Chironominae | 0 | 0 | 0 | 0 | 0 |
| Tribe: Chironomini | 0 | 0 | 0 | 0 | 0 |
| <u>Microtendipes</u> | 4 | 10 | 0 | 0 | 6 |
| <u>Paracladopelma</u> | 0 | 0 | 4 | 0 | 0 |
| <u>Polypedilum</u> | 4 | 0 | 9 | 0 | 33 |
| Tribe: Tanytarsini | 0 | 10 | 0 | 0 | 0 |
| <u>Micropsectra</u> | 22 | 80 | 22 | 40 | 6 |
| <u>Stempellinella</u> | 0 | 0 | 0 | 3 | 6 |
| <u>Tanytarsus</u> | 4 | 0 | 30 | 3 | 0 |
| Subfamily: Diamesinae | 0 | 0 | 0 | 0 | 0 |
| Tribe: Diamesini | 0 | 0 | 0 | 0 | 0 |

| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
|------------------------------|-----------|-------------|-------------|-------------|-------------|
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| <i>Diamesa</i> | 0 | 0 | 0 | 0 | 6 |
| Subfamily: Orthoclaadiinae | 0 | 40 | 4 | 80 | 6 |
| <i>Brillia</i> | 0 | 70 | 4 | 80 | 6 |
| <i>Corynoneura</i> | 0 | 20 | 0 | 0 | 0 |
| <i>Eukiefferiella</i> | 15 | 150 | 4 | 23 | 11 |
| <i>Heterotanytarsus</i> | 0 | 0 | 4 | 0 | 0 |
| <i>Heterotrissocladius</i> | 0 | 0 | 17 | 0 | 0 |
| <i>Orthocladus complex</i> | 0 | 30 | 30 | 3 | 6 |
| <i>Parakiefferiella</i> | 0 | 0 | 17 | 3 | 0 |
| <i>Parametriocnemus</i> | 0 | 20 | 0 | 0 | 0 |
| <i>Rheocricotopus</i> | 0 | 20 | 4 | 0 | 0 |
| <i>Synorthocladus</i> | 0 | 0 | 0 | 6 | 0 |
| <i>Thienemanniella</i> | 0 | 0 | 13 | 0 | 0 |
| <i>Tvetenia</i> | 4 | 150 | 0 | 200 | 0 |
| Subfamily: Tanypodinae | 4 | 0 | 0 | 0 | 0 |
| <i>Nilotanypus</i> | 0 | 0 | 0 | 3 | 0 |
| Tribe: Pentaneurini | 0 | 0 | 0 | 0 | 0 |
| <i>Thienemannimyia group</i> | 4 | 0 | 48 | 3 | 6 |
| Family: Empididae | 0 | 0 | 0 | 0 | 6 |
| <i>Neoplasta</i> | 0 | 10 | 0 | 0 | 0 |
| <i>Oreogeton</i> | 0 | 0 | 0 | 0 | 6 |
| Family: Simuliidae | 4 | 10 | 0 | 0 | 6 |
| <i>Helodon</i> | 4 | 0 | 0 | 0 | 6 |
| <i>Simulium</i> | 193 | 120 | 4 | 29 | 217 |
| Family: Tipulidae | 0 | 0 | 0 | 0 | 0 |
| <i>Dicranota</i> | 4 | 10 | 4 | 3 | 0 |
| <i>Hexatoma</i> | 0 | 0 | 0 | 0 | 6 |
| | | | | | |
| Subphylum: Chelicerata | 0 | 0 | 0 | 0 | 0 |
| Class: Arachnida | 0 | 0 | 0 | 0 | 0 |
| Order: Trombidiformes | 0 | 0 | 0 | 0 | 0 |

| | | | | | |
|-----------------------------------|-----------|-------------|-------------|-------------|-------------|
| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| Family: Hydryphantidae | 0 | 0 | 0 | 0 | 0 |
| <u>Protzia</u> | 0 | 0 | 0 | 0 | 6 |
| Family: Hygrobatidae | 0 | 0 | 0 | 0 | 0 |
| <u>Atractides</u> | 11 | 0 | 4 | 3 | 11 |
| <u>Hygrobates</u> | 0 | 30 | 30 | 0 | 0 |
| Family: Lebertiidae | 0 | 0 | 0 | 0 | 0 |
| <u>Lebertia</u> | 0 | 0 | 4 | 0 | 11 |
| Family: Sperchontidae | 0 | 0 | 0 | 0 | 0 |
| <u>Sperchon</u> | 22 | 20 | 9 | 0 | 6 |
| Family: Torrenticolidae | 0 | 0 | 0 | 0 | 0 |
| <u>Testudacarus</u> | 0 | 0 | 0 | 0 | 6 |
| | | | | | |
| Suborder: Prostigmata | 0 | 0 | 0 | 0 | 0 |
| Family: Stygothrombidiidae | 0 | 0 | 0 | 0 | 0 |
| <u>Stygothrombium</u> | 0 | 10 | 0 | 0 | 0 |
| | | | | | |
| Order: Sarcoptiformes | 0 | 0 | 0 | 0 | 0 |
| Order: Oribatida | 0 | 0 | 0 | 0 | 6 |
| | | | | | |
| Class: Malacostraca | 0 | 0 | 0 | 0 | 0 |
| Order: Amphipoda | 0 | 0 | 9 | 0 | 0 |
| Family: Crangonyctidae | 0 | 0 | 0 | 0 | 0 |
| <u>Crangonyx</u> | 15 | 0 | 4 | 0 | 0 |
| | | | | | |
| Phylum: Mollusca | 0 | 0 | 0 | 0 | 0 |
| Class: Bivalvia | 0 | 0 | 0 | 0 | 0 |
| Order: Veneroida | 0 | 0 | 0 | 0 | 0 |
| Family: Pisidiidae | 0 | 10 | 0 | 14 | 0 |
| <u>Pisidium</u> | 0 | 0 | 4 | 9 | 0 |
| | | | | | |
| Class: Gastropoda | 0 | 0 | 0 | 0 | 0 |

| | | | | | |
|---|-------------|-------------|-------------|-------------|-------------|
| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| Order: Basommatophora | 0 | 0 | 0 | 0 | 0 |
| Family: Physidae | 0 | 0 | 0 | 6 | 0 |
| | | | | | |
| Phylum: Annelida | 0 | 0 | 0 | 0 | 0 |
| Subphylum: Clitellata | 0 | 0 | 0 | 0 | 0 |
| Class: Oligochaeta | 0 | 0 | 0 | 0 | 0 |
| Order: Lumbriculida | 0 | 0 | 0 | 0 | 0 |
| Family: Lumbriculidae | 0 | 0 | 0 | 0 | 0 |
| <u>Lumbriculus</u> | 0 | 0 | 13 | 0 | 0 |
| | | | | | |
| Order: Tubificida | 0 | 0 | 0 | 0 | 0 |
| Family: Naididae | 0 | 0 | 0 | 0 | 0 |
| Subfamily: Tubificinae with hair chaetae | 4 | 0 | 335 | 3 | 0 |
| Subfamily: Tubificinae without hair chaetae | 4 | 0 | 13 | 3 | 33 |
| | | | | | |
| Phylum: Cnidaria | 0 | 0 | 0 | 0 | 0 |
| Class: Hydrozoa | 0 | 0 | 0 | 0 | 0 |
| Order: Anthoathecatae | 0 | 0 | 0 | 0 | 0 |
| Family: Hydridae | 0 | 0 | 0 | 0 | 0 |
| <u>Hydra</u> | 0 | 0 | 0 | 3 | 0 |
| Totals: | 1190 | 3500 | 1354 | 940 | 1806 |
| | | | | | |
| Taxa present but not included: | | | | | |
| | | | | | |
| <u>Terrestrials</u> | 0 | 0 | 0 | 6 | 0 |
| | | | | | |
| Phylum: Arthropoda | 0 | 0 | 0 | 0 | 0 |
| Subphylum: Hexapoda | 0 | 0 | 0 | 0 | 0 |
| Class: Insecta | 0 | 0 | 0 | 0 | 0 |
| Order: Diptera | 0 | 0 | 0 | 0 | 0 |
| Family: Cecidomyiidae | 0 | 10 | 0 | 0 | 0 |

| | | | | | |
|--------------------------------|-----------|-------------|-------------|-------------|-------------|
| Year: | 2019 | 2019 | 2019 | 2019 | 2019 |
| Sample: | RGD-AQ11 | CRB-DS-AQ01 | RGD-DS-AQ12 | JOR-DS-AQ31 | 21M-DS-AQ21 |
| Sample Collection Date: | 30-Jul-19 | 30-Jul-19 | 31-Jul-19 | 31-Jul-19 | 30-Jul-19 |
| CC#: | CC200370 | CC200372 | CC200373 | CC200374 | CC200375 |
| | | | | | |
| Order: Psocodea | 4 | 0 | 0 | 0 | 0 |
| | | | | | |
| Class: Maxillipoda | 0 | 0 | 0 | 0 | 0 |
| Class: Copepoda | 0 | 0 | 4 | 3 | 0 |
| Totals: | 4 | 10 | 4 | 9 | 0 |

Appendix C

Benthic Invertebrate (CABIN) Sampling Datasheets

Site Code: 21M-DS-1002

☐ Occupational Health & Safety: Site Inspection Sheet completed

CABIN Study Name: _____ Local Basin Name: Twenty One Mile

River/Stream Name: Twenty One Mile Stream Order: (map scale 1:50,000) _____

Select one: ☒ Test Site ☐ Potential Reference Site

Cross bridge @ Labrador Parking lot - follow gravel path to 1st fire clearing - go S-SW to stream (follow + line)

Information Source:

- ☒ Forest ☐ Field/Pasture ☐ Agriculture ☒ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☒ Other *Things*

Information Source:

- ☒ Forest ☐ Field/Pasture ☐ Agriculture ☐ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☐ Other Time cleaning

100 0501935 5552826

Latitude: _____ N Longitude: - _____ W (DMS or DD)

Elevation: 643 (fast or masl) GPS Datum: ☐ GRS80 (NAD83/WGS84) ☐ Other:

Location Map Drawing

indicate north

Note: Indicate north

Field Crew: MS + JSite Code: 21M-DS-AQZ1Sampling Date: (DD/MM/YYYY) 30/07/2019**Photos**

- ☒ Field Sheet ☒ Upstream ☒ Downstream ☒ Across Site ☐ Aerial View
☐ Substrate (exposed) ☒ Substrate (aquatic) ☐ Other + 2 of log w/ stream?

REACH DATA (represents 6 times bankfull width)

1. Habitat Types: (check those present)

- ☒ Riffle ☐ Rapids ☐ Straight run ☒ Pool/Back Eddy

2. Canopy Coverage: (stand in middle of stream and look up, check one)

- ☐ 0 % ☒ 1-25 % ☐ 26-50 % ☐ 51-75 % ☐ 76-100 %

3. Macrophyte Coverage: (not algae or moss, check one)

- ☒ 0 % ☐ 1-25 % ☐ 26-50 % ☐ 51-75 % ☐ 76-100 %

4. Streamside Vegetation: (check those present)

- ☒ ferns/grasses ☒ shrubs ☐ deciduous trees ☐ coniferous trees

5. Dominant Streamside Vegetation: (check one)

- ☐ ferns/grasses ☒ shrubs ☐ deciduous trees ☐ coniferous trees

6. Periphyton Coverage on Substrate: (benthic algae, not moss, check one)

- ☐ 1 - Rocks are not slippery, no obvious colour (thin layer < 0.5 mm thick)
☒ 2 - Rocks are slightly slippery, yellow-brown to light green colour (0.5-1 mm thick)
☐ 3 - Rocks have a noticeable slippery feel (footing is slippery), with patches of thicker green to brown algae (1-5 mm thick)
☐ 4 - Rocks are very slippery (algae can be removed with thumbnail), numerous large clumps of green to dark brown algae (5 mm -20 mm thick)
☐ 5 - Rocks are mostly obscured by algal mat, extensive green, brown to black algal mass may have long strands (> 20 mm thick)

- sporadic patches of green algae on rocks.

Note: 1 through 5 represent categories entered into the CABIN database.

BENTHIC MACROINVERTEBRATE DATAHabitat sampled: (check one) ☒ riffle ☐ rapids ☐ straight run

| | |
|---------------------------------|--------------|
| 400 µm mesh Kick Net | |
| Person sampling | <u>M.S.</u> |
| Sampling time (i.e. 3 min.) | <u>3 min</u> |
| No. of sample jars | <u>1</u> |
| Typical depth in kick area (cm) | <u>15cm.</u> |

Preservative used: Ethanol.

Sampled sieved on site using "Bucket Swirling Method":

☒ YES ☐ NOIf YES, debris collected for QAQC ☐+ 2 sculpin

Note: Indicate if a sampling method other than the recommended 400 µm mesh kick net is used.

Field Crew: MS + JSite Code: 21N-DS-AQ21Sampling Date: (DD/MM/YYYY) 30/07/2019**WATER CHEMISTRY DATA** Time: 1400 (24 hr clock) Time zone: _____Air Temp: _____ (°C) Water Temp: 13.3 (°C) pH: 6.97Specific Conductance: 51.8 (µs/cm) DO: 9.78 (mg/L) Turbidity: - (NTU)Conductivity 36.3 µs/cmDO 93.6%

32.5 TDS mg/L

706.1 mmHd

Check if water samples were collected for the following analyses:

- ☐ TSS (Total Suspended Solids)
- ☐ Nitrogen (i.e. Total, Nitrate, Nitrite, Dissolved, and/or Ammonia)
- ☐ Phosphorus (Total, Ortho, and/or Dissolved)
- ☐ Major Ions (i.e. Alkalinity, Hardness, Chloride, and/or Sulphate) ☐ Other _____

Note: Determining alkalinity is recommended, as are other analyses, but not required for CABIN assessments.

CHANNEL DATA**Slope** - Indicate how slope was measured: (check one)

- ☐
- Calculated from map**

Scale: _____ (Note: small scale map recommended if field measurement is not possible - i.e. 1:20,000).

contour interval (vertical distance) _____ (m),

distance between contour intervals (horizontal distance) _____ (m)

slope = vertical distance/horizontal distance = _____

OR

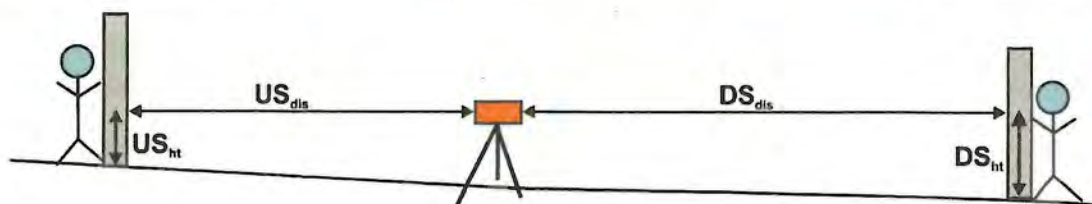
- ☒
- Measured in field**

Circle device used and fill out table according to device:

a. Survey Equipment b. Hand Level & Measuring Tape

Clinometer = 20.5m; < 1%

| Measurements | Upstream (U/S) | Downstream(D/S) | Calculation |
|-----------------------------------|--------------------------------------|--------------------------------------|---|
| ^a Top Hairline (T) | | | |
| ^a Mid Hairline (ht) OR | | | |
| ^b Height of rod | | | |
| ^a Bottom Hairline (B) | | | |
| ^b Distance (dis) OR | | | US _{dis} + DS _{dis} = |
| ^a T-B x 100 | ^a US _{dis} = T-B | ^a DS _{dis} = T-B | |
| Change in height (Δht) | | | DS _{ht} - US _{ht} = |
| Slope (Δht/total dis) | | | |



Field Crew: _____ Site Code: _____

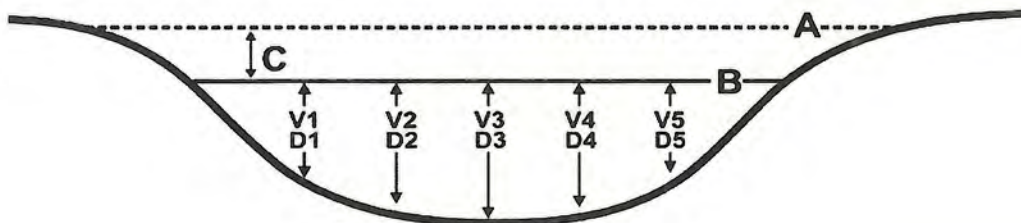
Sampling Date: (DD/MM/YYYY) _____

Widths and Depth

Location at site: 5m ups ad 'base' (Indicate where in sample reach, ex. d/s of kick area)

A - Bankfull Width: 11.4 (m) B - Wetted Stream Width: 11.3 (m)

C - Bankfull-Wetted Depth (height from water surface to Bankfull): 20cm (cm)



Note:
Wetted widths > 5 m, measure a minimum of 5-6 equidistant locations;
Wetted widths < 5 m, measure 3-4 equidistant locations.

Velocity and Depth

Check appropriate velocity measuring device and fill out the appropriate section in chart below. Distance from shore and depth are required regardless of method:

☒ **Velocity Head Rod (or ruler):** Velocity Equation (m/s) = $\sqrt{2(\Delta D/100) * 9.81}$

☐ **Rotary meters:** Gurley/Price/Mini-Price/Propeller (Refer to specific meter conversion chart for calculation)

☐ **Direct velocity measurements:** ☐ Marsh-McBirney ☐ Sontek or ☐ Other _____

@ Base →

| | 1 | 2 | 3 | 4 | 5 | 6 | AVG |
|---|----|------|----|------|------|-----|-----|
| Distance from Shore (m) | 1m | 2.5m | 4m | 5.5m | 7m | 8.5 | |
| Depth (D) (cm) | | | | | | | |
| Velocity Head Rod (ruler) | | | | | | | |
| Flowing water Depth (D ₁) (cm) | 31 | 28.5 | 21 | 17 | 9 | 6 | |
| Depth of Stagnation (D ₂) (cm) | 35 | 32 | 25 | 18 | 10.5 | 8 | |
| Change in depth (ΔD=D ₂ -D ₁) (cm) | | | | | | | |
| Rotary meter | | | | | | | |
| Revolutions | | | | | | | |
| Time (minimum 40 seconds) | | | | | | | |
| Direct Measurement or calculation | | | | | | | |
| Velocity (V) (m/s) | | | | | | | |

Field Crew: _____

Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SUBSTRATE DATA**Surrounding/Interstitial Material**

Circle the substrate size category for the surrounding material.

| Substrate Size Class | Category |
|------------------------------------|----------|
| Organic Cover | 0 |
| < 0.1 cm (fine sand, silt or clay) | 1 |
| 0.1-0.2 cm (coarse sand) | 2 |
| 0.2-1.6 cm (gravel) | (3) |
| 1.6-3.2 cm (small pebble) | 4 |
| 3.2-6.4 cm (large pebble) | 5 |
| 6.4-12.8 cm (small cobble) | 6 |
| 12.8-25.6 cm (cobble) | 7 |
| > 25.6 cm (boulder) | 8 |
| Bedrock | 9 |

100 Pebble Count & Substrate Embeddedness

- Measure the intermediate axis (100 rocks) and embeddedness (10 rocks) of substrate in the stream bed.
- Indicate B for bedrock, S for sand/silt/clay (particles < 0.2 cm) and O for organic material.
- Embeddedness categories (E): Completely embedded = 1, 3/4 embedded, 1/2 embedded, 1/4 embedded, unembedded = 0

| Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E |
|---------------|--------------------|---------------|---|---------------|-----|---------------|---|
| 1 | 2.4 | | | 26 | 7.1 | | |
| 2 | 2.0 | | | 27 | 4.5 | | |
| 3 | 1.3 | | | 28 | 4.1 | | |
| 4 | 3.0 | | | 29 | 4.9 | | |
| 5 | 4.2 | | | 30 | 5.0 | 0 | |
| 6 | 2.2 | | | 31 | 3.0 | | |
| 7 | 4.0 | | | 32 | 5.0 | | |
| 8 | 6.3 | | | 33 | 5.0 | | |
| 9 | .06 | | | 34 | 5.0 | | |
| 10 | 2.1 5.6 | 1/4 | | 35 | 2.0 | | |
| 11 | 5.2 | | | 36 | 4.1 | | |
| 12 | 4.0 | | | 37 | 1.5 | | |
| 13 | 3.3 | | | 38 | .4 | | |
| 14 | 3.9 | | | 39 | 3.8 | | |
| 15 | 5.0 | | | 40 | 5.8 | 1/4 | |
| 16 | 6.2 | | | 41 | 4.3 | | |
| 17 | 4.5 | | | 42 | 4.0 | | |
| 18 | 2.7 | | | 43 | 2.0 | | |
| 19 | 1.5 | | | 44 | 2.5 | | |
| 20 | 6.0 | 1/4 | | 45 | 2.0 | | |
| 21 | 3.0 | | | 46 | 1.5 | | |
| 22 | 2.2 | | | 47 | 4.3 | | |
| 23 | 4.1 | | | 48 | 4.0 | | |
| 24 | 4.2 | | | 49 | 6.2 | | |
| 25 | 6.0 | | | 50 | 5.0 | 0 | |
| | | | | 51 | 3.4 | | |
| | | | | 52 | 7.7 | | |
| | | | | 53 | 3.8 | | |
| | | | | 54 | 2.7 | | |
| | | | | 55 | 1.0 | | |
| | | | | 56 | 2.0 | | |
| | | | | 57 | .7 | | |
| | | | | 58 | 4.6 | | |
| | | | | 59 | 4.4 | | |
| | | | | 60 | 1.4 | 0 | |
| | | | | 61 | 3.0 | | |
| | | | | 62 | .6 | | |
| | | | | 63 | 2.5 | | |
| | | | | 64 | .7 | | |
| | | | | 65 | 2.4 | | |
| | | | | 66 | 5.0 | | |
| | | | | 67 | 3.7 | | |
| | | | | 68 | 2.8 | | |
| | | | | 69 | .9 | | |
| | | | | 70 | 5.3 | 1/4 | |
| | | | | 71 | .2 | | |
| | | | | 72 | 2.9 | | |
| | | | | 73 | 5.7 | | |
| | | | | 74 | 4.3 | | |
| | | | | 75 | 6.2 | | |
| | | | | 76 | 6.5 | | |
| | | | | 77 | 2.4 | | |
| | | | | 78 | 4.3 | | |
| | | | | 79 | 5.0 | | |
| | | | | 80 | 3.8 | 1/2 | |
| | | | | 81 | 4.4 | | |
| | | | | 82 | 3.5 | | |
| | | | | 83 | 4.6 | | |
| | | | | 84 | 1.2 | | |
| | | | | 85 | 2.5 | | |
| | | | | 86 | 2.8 | | |
| | | | | 87 | 3.5 | | |
| | | | | 88 | 2.2 | | |
| | | | | 89 | 5.2 | | |
| | | | | 90 | 3.0 | 0 | |
| | | | | 91 | 4.2 | | |
| | | | | 92 | 3.2 | | |
| | | | | 93 | 2.2 | | |
| | | | | 94 | .4 | | |
| | | | | 95 | 3.4 | | |
| | | | | 96 | 2.2 | | |
| | | | | 97 | 3.4 | | |
| | | | | 98 | 6.0 | | |
| | | | | 99 | 5.0 | | |
| | | | | 100 | 3.9 | 0 | |

Note: The Wolman D50 (i.e. median diameter), Wolman Dg (i.e. geometric mean diameter) and the % composition of the substrate classes will be calculated automatically in the CABIN database using the 100 pebble data. All 100 pebbles must be measured in order for the CABIN database tool to perform substrate calculations.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SITE INSPECTION

Site Inspected by: _____

Communication Information

☐ Itinerary left with contact person (include contact numbers)

Contact Person: _____ Time checked-in: _____

Form of communication: ☐ radio ☐ cell ☐ satellite ☐ hotel/pay phone ☐ SPOT

Phone number: () _____

Vehicle Safety

☐ Safety equipment (first aid, fire extinguisher, blanket, emergency kit in vehicle)

☐ Equipment and chemicals safely secured for transport

☐ Vehicle parked in safe location; pylons, hazard light, reflective vests if necessary

Notes:

Shore & Wading Safety

☐ Wading Task Hazard Analysis read by all field staff

☐ Wading Safe Work Procedures read by all field staff

☐ Instream hazards identified (i.e. log jams, deep pools, slippery rocks)

☐ PFD worn

☐ Appropriate footwear, waders, wading belt

☐ Belay used

Notes:

Field Crew: MS + J. Site Code: CRB-DS-AQ01
Sampling Date: (DD/MM/YYYY) 30/07/2019 (1500 - 1600)

☐ Occupational Health & Safety: Site Inspection Sheet completed

PRIMARY SITE DATA

CABIN Study Name: _____ Local Basin Name: Crabapple Creek

River/Stream Name: Crabapple Stream Order: (map scale 1:50,000) _____

Select one: ☒ Test Site ☐ Potential Reference Site

Geographical Description/Notes:

Lawrence Rd parking lot - follow small/short path to
S-cut into forest ~ 10m in

Surrounding Land Use: (check those present)

☒ Forest ☐ Field/Pasture ☐ Agriculture ☒ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☐ Other _____

Information Source: _____

Dominant Surrounding Land Use: (check one)

☐ Forest ☐ Field/Pasture ☐ Agriculture ☒ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☐ Other _____

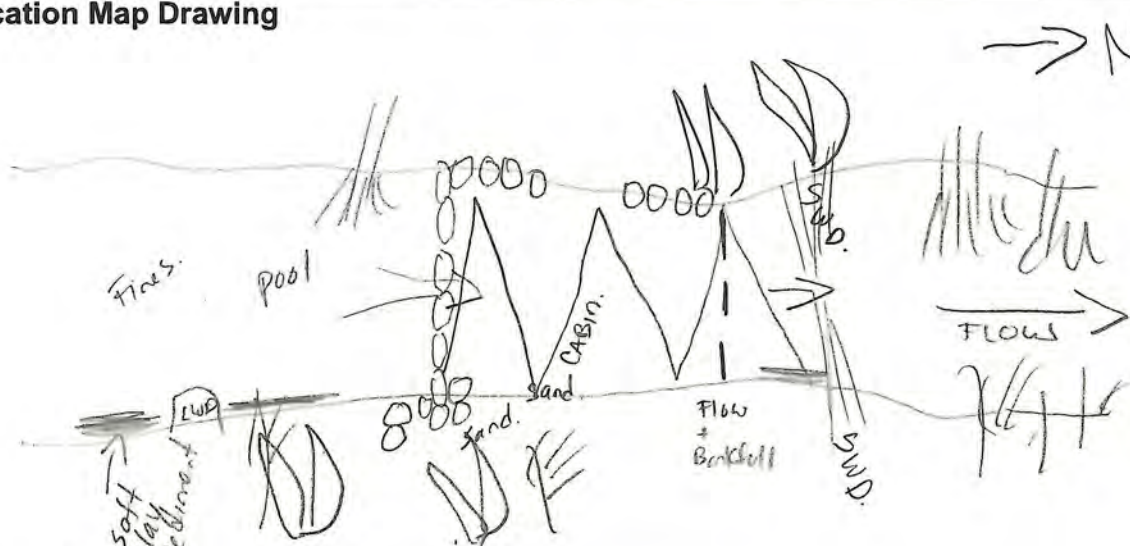
Information Source: _____

Location Data 10U 0502024 5552701

Latitude: _____ N Longitude: - _____ W (DMS or DD)

Elevation: 656 (fast or masl) GPS Datum: ☐ GRS80 (NAD83/WGS84) ☐ Other: _____

Site Location Map Drawing



Note: Indicate north

Field Crew: MS + SSite Code: CEB-DS-AQ01Sampling Date: (DD/MM/YYYY) 30/07/2019**Photos**

- ☒ Field Sheet
 ☒ Upstream
 ☒ Downstream
 ☒ Across Site
 ☐ Aerial View
☐ Substrate (exposed)
 ☒ Substrate (aquatic)
 ☐ Other _____

REACH DATA (represents 6 times bankfull width)

+ 2 moss mats

1. Habitat Types: (check those present)

- ☒ Riffle
 ☐ Rapids
 ☐ Straight run
 ☒ Pool/Back Eddy

2. Canopy Coverage: (stand in middle of stream and look up, check one)

- ☐ 0 %
 ☐ 1-25 %
 ☐ 26-50 %
 ☐ 51-75 %
 ☒ 76-100 %

3. Macrophyte Coverage: (not algae or moss, check one)

- ☒ 0 %
 ☐ 1-25 %
 ☐ 26-50 %
 ☐ 51-75 %
 ☐ 76-100 %

4. Streamside Vegetation: (check those present)

- ☒ ferns/grasses
 ☒ shrubs
 ☒ deciduous trees
 ☐ coniferous trees

5. Dominant Streamside Vegetation: (check one)

- ☐ ferns/grasses
 ☒ shrubs
 ☐ deciduous trees
 ☐ coniferous trees

6. Periphyton Coverage on Substrate: (benthic algae, not moss, check one)

Lot of moss mats on rocks

- ☐ 1 - Rocks are not slippery, no obvious colour (thin layer < 0.5 mm thick)
☒ 2 - Rocks are slightly slippery, yellow-brown to light green colour (0.5-1 mm thick)
☐ 3 - Rocks have a noticeable slippery feel (footing is slippery), with patches of thicker green to brown algae (1-5 mm thick)
☐ 4 - Rocks are very slippery (algae can be removed with thumbnail), numerous large clumps of green to dark brown algae (5 mm -20 mm thick)
☐ 5 - Rocks are mostly obscured by algal mat, extensive green, brown to black algal mass may have long strands (> 20 mm thick)

Note: 1 through 5 represent categories entered into the CABIN database.

BENTHIC MACROINVERTEBRATE DATAHabitat sampled: (check one) ☒ riffle ☐ rapids ☐ straight run

| | |
|---------------------------------|-------|
| 400 µm mesh Kick Net | |
| Person sampling | MS. |
| Sampling time (i.e. 3 min.) | 3 min |
| No. of sample jars | 2 |
| Typical depth in kick area (cm) | 12cm. |

Preservative used: Ethanol

Sampled sieved on site using "Bucket Swirling Method":

☐ YES ☒ NOIf YES, debris collected for QAQC ☐

Note: Indicate if a sampling method other than the recommended 400 µm mesh kick net is used.

Field Crew: M.S.J. Site Code: CRB-DS-AQ01
 Sampling Date: (DD/MM/YYYY) 30/07/2019

WATER CHEMISTRY DATA Time: 1540 (24 hr clock) Time zone: _____
 Air Temp: ~13 (°C) Water Temp: 13.9 (°C) pH: 7.6
 Specific Conductance: 234.9 (µs/cm) DO: 10.0 (mg/L) Turbidity: — (NTU)
Conductivity 184.9 µs/cm 96.7%
 Check if water samples were collected for the following analyses:
☐ TSS (Total Suspended Solids) 706.2 mmHg
☐ Nitrogen (i.e. Total, Nitrate, Nitrite, Dissolved, and/or Ammonia) 152.75 TDS
☐ Phosphorus (Total, Ortho, and/or Dissolved)
☐ Major Ions (i.e. Alkalinity, Hardness, Chloride, and/or Sulphate) ☐ Other _____

Note: Determining alkalinity is recommended, as are other analyses, but not required for CABIN assessments.

CHANNEL DATA

Slope - Indicate how slope was measured: (check one)

☐ **Calculated from map**

Scale: _____ (Note: small scale map recommended if field measurement is not possible - i.e. 1:20,000).
 contour interval (vertical distance) _____ (m),
 distance between contour intervals (horizontal distance) _____ (m)
 slope = vertical distance/horizontal distance = _____

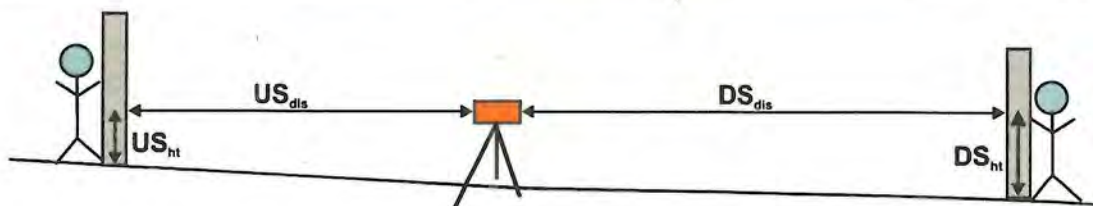
OR

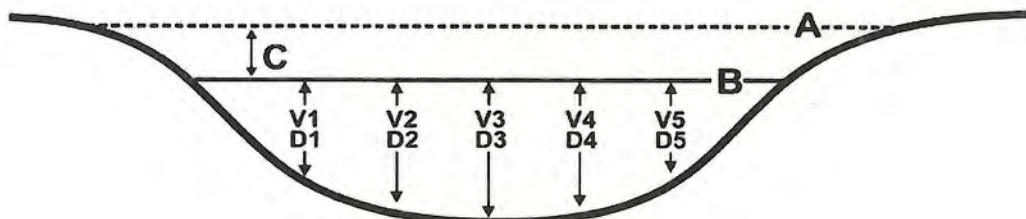
☐ **Measured in field**

Circle device used and fill out table according to device:
 a. Survey Equipment b. Hand Level & Measuring Tape

Clinometer 3% @ 5.2m

| Measurements | Upstream (U/S) | Downstream(D/S) | Calculation |
|-----------------------------------|-------------------------------------|-------------------------------------|--|
| ^a Top Hairline (T) | | | |
| ^a Mid Hairline (ht) OR | | | |
| ^b Height of rod | | | |
| ^a Bottom Hairline (B) | | | |
| ^b Distance (dis) OR | | | US _{dis} +DS _{dis} = |
| ^a T-B x 100 | ^a US _{dis} =T-B | ^a DS _{dis} =T-B | |
| Change in height (Δht) | | | DS _{ht} -US _{ht} = |
| Slope (Δht/total dis) | | | |



Field Crew: MS, JSite Code: CEB-D5-A001Sampling Date: (DD/MM/YYYY) 30/07/2019**Widths and Depth**Location at site: middle of kick area (Indicate where in sample reach, ex. d/s of kick area)A - Bankfull Width: 3.4 (m)B - Wetted Stream Width: 2.3 (m)C - Bankfull-Wetted Depth (height from water surface to Bankfull): 35cm (cm)

Note:

Wetted widths > 5 m, measure a minimum of 5-6 equidistant locations;

Wetted widths < 5 m, measure 3-4 equidistant locations.

Velocity and Depth

Check appropriate velocity measuring device and fill out the appropriate section in chart below. Distance from shore and depth are required regardless of method:

☒ **Velocity Head Rod (or ruler):** Velocity Equation (m/s) = $\sqrt{2(\Delta D/100) * 9.81}$ ☐ **Rotary meters:** Gurley/Price/Mini-Price/Propeller (Refer to specific meter conversion chart for calculation)☐ **Direct velocity measurements:** ☐ Marsh-McBirney ☐ Sontek or ☐ Other _____Base →

| | 1 | 2 | 3 | 4 | 5 | 6 | AVG |
|---|------|------|------|-----|---|---|-----|
| Distance from Shore (m) | 0.5 | 1.0 | 1.5 | 2.0 | | | |
| Depth (D) (cm) | | | | | | | |
| Velocity Head Rod (ruler) | | | | | | | |
| Flowing water Depth (D ₁) (cm) | 10.5 | 17 | 17.5 | 6.0 | | | |
| Depth of Stagnation (D ₂) (cm) | 13 | 18.5 | 19.0 | 7.0 | | | |
| Change in depth ($\Delta D = D_2 - D_1$) (cm) | | | | | | | |
| Rotary meter | | | | | | | |
| Revolutions | | | | | | | |
| Time (minimum 40 seconds) | | | | | | | |
| Direct Measurement or calculation | | | | | | | |
| Velocity (V) (m/s) | | | | | | | |

Field Crew: MS JSite Code: CRB-DS-A901Sampling Date: (DD/MM/YYYY) 30/07/2019**SUBSTRATE DATA****Surrounding/Interstitial Material**

Circle the substrate size category for the surrounding material.

| Substrate Size Class | Category |
|------------------------------------|----------|
| Organic Cover | 0 |
| < 0.1 cm (fine sand, silt or clay) | 1 |
| 0.1-0.2 cm (coarse sand) | (2) |
| 0.2-1.6 cm (gravel) | 3 |
| 1.6-3.2 cm (small pebble) | 4 |
| 3.2-6.4 cm (large pebble) | 5 |
| 6.4-12.8 cm (small cobble) | 6 |
| 12.8-25.6 cm (cobble) | 7 |
| > 25.6 cm (boulder) | 8 |
| Bedrock | 9 |

100 Pebble Count & Substrate Embeddedness

- Measure the intermediate axis (100 rocks) and embeddedness (10 rocks) of substrate in the stream bed.
- Indicate B for bedrock, S for sand/silt/clay (particles < 0.2 cm) and O for organic material.
- Embeddedness categories (E): Completely embedded = 1, 3/4 embedded, 1/2 embedded, 1/4 embedded, unembedded = 0

| Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E |
|---------------|-----|---------------|-----|---------------|-----|---------------|-----|
| 1 10.5 | | 26 11.0 | | 51 4.5 | | 76 9.5 | |
| 2 7.2 | | 27 25.0 | | 52 5.0 | | 77 10 | |
| 3 3.5 | | 28 10.5 | | 53 4.0 | | 78 11 | |
| 4 8.0 | | 29 6.5 | | 54 16.0 | | 79 sand | |
| 5 2.0 | | 30 10.0 | 1/4 | 55 10.0 | | 80 9.0 | |
| 6 3.8 | | 31 3.0 | | 56 11.0 | | 81 13.0 | |
| 7 2.5 | | 32 5.0 | | 57 9.0 | | 82 25 | 1/2 |
| 8 3.0 | | 33 sand. | | 58 6.0 | | 83 17 | |
| 9 5.0 | | 34 3.5 | | 59 11.0 | | 84 21 | |
| 10 9.0 | | 35 5.0 | | 60 13.0 | | 85 10 | |
| 11 6.9 | | 36 13.0 | | 61 13.5 | | 86 7 | |
| 12 3.5 | | 37 6.0 | | 62 10 | | 87 sand | |
| 13 5.0 | 1/4 | 38 14.0 | | 63 14.0 | 3/4 | 88 14 | |
| 14 sand. | | 39 4.0 | | 64 8 | | 89 fines | |
| 15 3.0 | | 40 25 | | 65 16 | | 90 19 | |
| 16 3.5 | | 41 11 | | 66 6.5 | | 91 1.5 | |
| 17 7.5 | | 42 6.0 | 1/2 | 67 5.0 | | 92 16.0 | 1/4 |
| 18 7.0 | | 43 13.0 | | 68 7.0 | | 93 1.0 | |
| 19 3.5 | | 44 26.0 | | 69 12.0 | | 94 11 | |
| 20 3.0 | | 45 13.0 | | 70 16.5 | | 95 sand | |
| 21 7.0 | 1/4 | 46 5.0 | | 71 sand | | 96 sand | |
| 22 sand. | | 47 14.0 | | 72 18 | 1/2 | 97 16 | |
| 23 6.8 | | 48 3.0 | | 73 16 | | 98 1.8 | |
| 24 7.0 | | 49 5.5 | 0 | 74 8 | | 99 2.1 | |
| 25 8.0 | | 50 11.5 | 0 | 75 5.5 | | 100 6.0 | 0 |

Note: The Wolman D50 (i.e. median diameter), Wolman Dg (i.e. geometric mean diameter) and the % composition of the substrate classes will be calculated automatically in the CABIN database using the 100 pebble data. All 100 pebbles must be measured in order for the CABIN database tool to perform substrate calculations.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SITE INSPECTION

Site Inspected by: _____

Communication Information

☐ Itinerary left with contact person (include contact numbers)

Contact Person: _____ Time checked-in: _____

Form of communication: ☐ radio ☐ cell ☐ satellite ☐ hotel/pay phone ☐ SPOT

Phone number: () _____

Vehicle Safety

☐ Safety equipment (first aid, fire extinguisher, blanket, emergency kit in vehicle)

☐ Equipment and chemicals safely secured for transport

☐ Vehicle parked in safe location; pylons, hazard light, reflective vests if necessary

Notes:

Shore & Wading Safety

☐ Wading Task Hazard Analysis read by all field staff

☐ Wading Safe Work Procedures read by all field staff

☐ Instream hazards identified (i.e. log jams, deep pools, slippery rocks)

☐ PFD worn

☐ Appropriate footwear, waders, wading belt

☐ Belay used

Notes:

Site Code: J012-DS-AQ31

Sampling Date: (DD/MM/YYYY) 31/07/2019

☐ Occupational Health & Safety: Site Inspection Sheet completed

PRIMARY SITE DATA

CABIN Study Name: _____ Local Basin Name: Jordan Creek

River/Stream Name: Jordan Creek Stream Order: (map scale 1:50,000) _____

Select one: ☐ Test Site ☐ Potential Reference Site

Geographical Description/Notes: Alpha Lake Parking lot - take path Northward, just past treatment facility - go into bush @ 1q corner.

Surrounding Land Use: (check those present)

☒ Forest☐ Field/Pasture☐ Agriculture

Information Source:

☒ Residential/Urban☐ Logging☐ Mining☐ Commercial/Industrial☐ Other

Dominant Surrounding Land Use: (check one)

Information Source:

☐ Forest☐ Field/Pasture☐ Agriculture☒ Residential/Urban☐ Logging☐ Mining☐ Commercial/Industrial☐ Other

Location Data

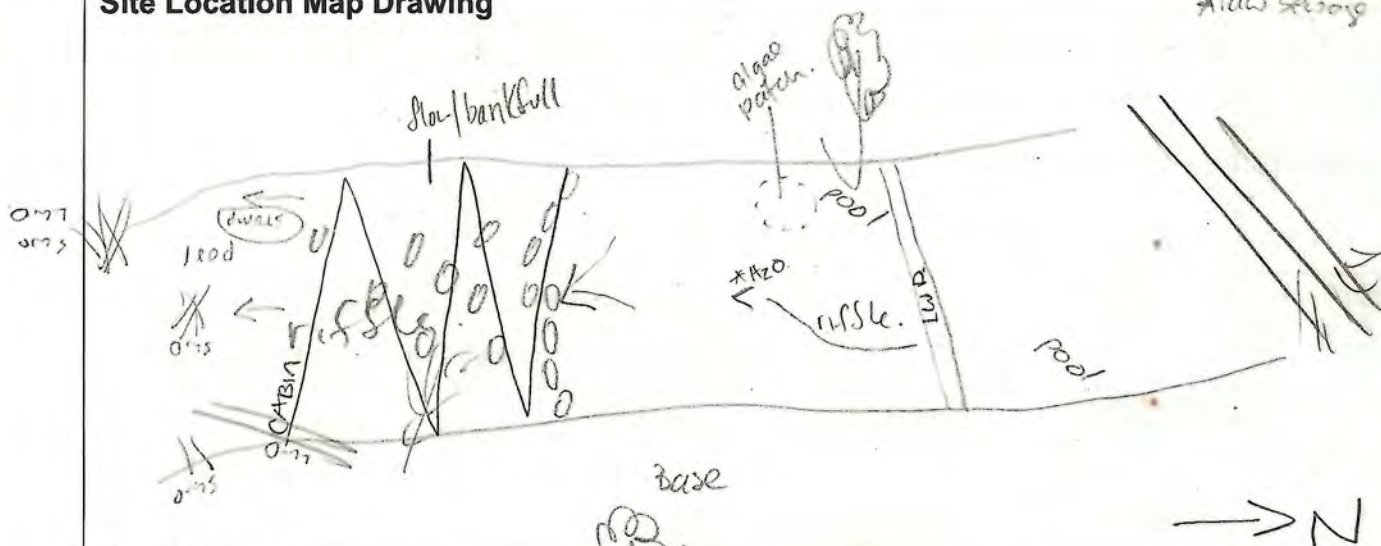
10U 0500194 5549249

Latitude: _____ N Longitude: - _____ W (DMS or DD)

Elevation: 624 (fast or masl) GPS Datum: ☐ GRS80 (NAD83/WGS84) ☐ Other:

Site Location Map Drawing

Draw Group Sells



Note: Indicate north

Field Crew: MS. H/L Site Code: JOR-DS-1031
 Sampling Date: (DD/MM/YYYY) 31/07/2019

Photos

☒ Field Sheet ☒ Upstream ☒ Downstream ☒ *3 Across Site ☐ Aerial View
☐ Substrate (exposed) ☒ Substrate (aquatic) ☐ Other _____

REACH DATA (represents 6 times bankfull width)

1. Habitat Types: (check those present)

☒ Riffle ☐ Rapids ☒ Straight run ☒ Pool/Back Eddy

2. Canopy Coverage: (stand in middle of stream and look up, check one)

☐ 0 % ☐ 1-25 % ☐ 26-50 % ☒ 51-75 % ☐ 76-100 %

3. Macrophyte Coverage: (not algae or moss, check one)

☒ 0 % ☐ 1-25 % ☐ 26-50 % ☐ 51-75 % ☐ 76-100 %

4. Streamside Vegetation: (check those present)

☒ ferns/grasses ☒ shrubs ☒ deciduous trees ☒ coniferous trees

5. Dominant Streamside Vegetation: (check one)

☐ ferns/grasses ☒ shrubs ☐ deciduous trees ☐ coniferous trees

6. Periphyton Coverage on Substrate: (benthic algae, not moss, check one)

- ☒ 1 - Rocks are not slippery, no obvious colour (thin layer < 0.5 mm thick)
☒ 2 - Rocks are slightly slippery, yellow-brown to light green colour (0.5-1 mm thick)
☐ 3 - Rocks have a noticeable slippery feel (footing is slippery), with patches of thicker green to brown algae (1-5 mm thick)
☐ 4 - Rocks are very slippery (algae can be removed with thumbnail), numerous large clumps of green to dark brown algae (5 mm -20 mm thick)
☐ 5 - Rocks are mostly obscured by algal mat, extensive green, brown to black algal mass may have long strands (> 20 mm thick)

Note: 1 through 5 represent categories entered into the CABIN database.

BENTHIC MACROINVERTEBRATE DATA

Habitat sampled: (check one) ☒ riffle ☐ rapids ☐ straight run

| | |
|---------------------------------|-------|
| 400 µm mesh Kick Net | |
| Person sampling | MS. |
| Sampling time (i.e. 3 min.) | 3 min |
| No. of sample jars | 1 |
| Typical depth in kick area (cm) | 24 cm |

Preservative used: Ethanol

Sampled sieved on site using "Bucket Swirling Method":

☐ YES ☒ NO

If YES, debris collected for QAQC ☐

Note: Indicate if a sampling method other than the recommended 400 µm mesh kick net is used.

Field Crew: _____ Site Code: JOK -Sampling Date: (DD/MM/YYYY) 31/07/2019**WATER CHEMISTRY DATA** Time: 1230 (24 hr clock) Time zone: PSTAir Temp: 12 (°C) Water Temp: 17.4 (°C) pH: 7.66Specific Conductance: 78.4 (µs/cm) DO: 9.44 (mg/L) Turbidity: - (NTU)Cond. 67.1 µs/cm 98.6 %

Check if water samples were collected for the following analyses:

- ☐ TSS (Total Suspended Solids) 707.3 mg/lg
- ☐ Nitrogen (i.e. Total, Nitrate, Nitrite, Dissolved, and/or Ammonia) 50.7 TDS
- ☐ Phosphorus (Total, Ortho, and/or Dissolved) 0.04 salinity (ppt)
- ☐ Major Ions (i.e. Alkalinity, Hardness, Chloride, and/or Sulphate) ☐ Other _____

Note: Determining alkalinity is recommended, as are other analyses, but not required for CABIN assessments.

CHANNEL DATA**Slope** - Indicate how slope was measured: (check one)☐ **Calculated from map**

Scale: _____ (Note: small scale map recommended if field measurement is not possible - i.e. 1:20,000).

contour interval (vertical distance) _____ (m),

distance between contour intervals (horizontal distance) _____ (m)

slope = vertical distance/horizontal distance = _____

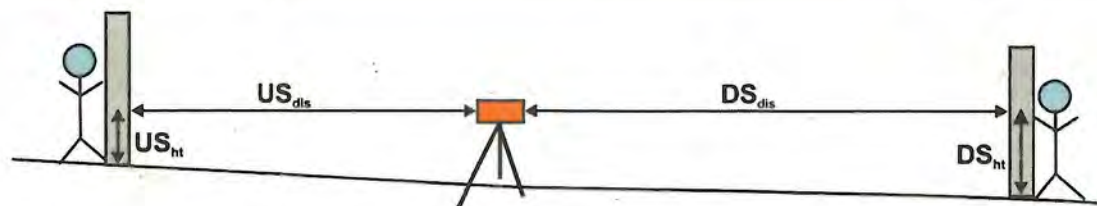
OR☒ **Measured in field**

Circle device used and fill out table according to device:

a. Survey Equipment b. Hand Level & Measuring Tape

Clinometer 3' @ 15m.

| Measurements | Upstream (U/S) | Downstream(D/S) | Calculation |
|-----------------------------------|-------------------------------------|-------------------------------------|--|
| ^a Top Hairline (T) | | | |
| ^a Mid Hairline (ht) OR | | | |
| ^b Height of rod | | | |
| ^a Bottom Hairline (B) | | | |
| ^b Distance (dis) OR | | | US _{dis} +DS _{dis} = |
| ^a T-B x 100 | ^a US _{dis} =T-B | ^a DS _{dis} =T-B | |
| Change in height (Δht) | | | DS _{ht} -US _{ht} = |
| Slope (Δht/total dis) | | | |



Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

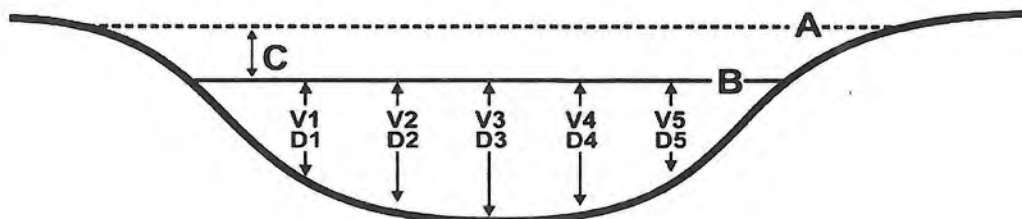
Widths and Depth

Location at site: middle of kick site (Indicate where in sample reach, ex. d/s of kick area)

A - Bankfull Width: 4.2 (m)

B - Wetted Stream Width: 4.4 (m) under cut

C - Bankfull-Wetted Depth (height from water surface to Bankfull): 14 cm (cm)



Note:

Wetted widths > 5 m, measure a minimum of 5-6 equidistant locations;

Wetted widths < 5 m, measure 3-4 equidistant locations.

Velocity and Depth

Check appropriate velocity measuring device and fill out the appropriate section in chart below. Distance from shore and depth are required regardless of method:

☒ **Velocity Head Rod (or ruler):** Velocity Equation (m/s) = $\sqrt{[2(\Delta D/100) * 9.81]}$

☐ **Rotary meters:** Gurley/Price/Mini-Price/Propeller (Refer to specific meter conversion chart for calculation)

☐ **Direct velocity measurements:** ☐ Marsh-McBirney ☐ Sontek or ☐ Other _____

opposite base:

| | 1 | 2 | 3 | 4 | 5 | 6 | AVG |
|---|------|-----|-----|------|------|----------------|-----|
| Distance from Shore (m) | 0.5 | 1.0 | 1.5 | 2.0 | 2.5 | 3.0 | 2.5 |
| Depth (D) (cm) | | | | | | | |
| Velocity Head Rod (ruler) | | | | | | | |
| Flowing water Depth (D ₁) (cm) | 12.5 | 15 | 24 | 31.5 | 25.5 | behind boulder | |
| Depth of Stagnation (D ₂) (cm) | 15 | 19 | 27 | 34 | 26.0 | ~ | |
| Change in depth ($\Delta D = D_2 - D_1$) (cm) | | | | | | | |
| Rotary meter | | | | | | | |
| Revolutions | | | | | | | |
| Time (minimum 40 seconds) | | | | | | | |
| Direct Measurement or calculation | | | | | | | |
| Velocity (V) (m/s) | | | | | | | |

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SUBSTRATE DATA**Surrounding/Interstitial Material**

Circle the substrate size category for the surrounding material.

| Substrate Size Class | Category |
|------------------------------------|----------|
| Organic Cover | 0 |
| < 0.1 cm (fine sand, silt or clay) | 1 |
| 0.1-0.2 cm (coarse sand) | 2 |
| 0.2-1.6 cm (gravel) | 3 |
| 1.6-3.2 cm (small pebble) | 4 |
| 3.2-6.4 cm (large pebble) | 5 |
| 6.4-12.8 cm (small cobble) | 6 |
| 12.8-25.6 cm (cobble) | 7 |
| > 25.6 cm (boulder) | 8 |
| Bedrock | 9 |

100 Pebble Count & Substrate Embeddedness

- Measure the intermediate axis (100 rocks) and embeddedness (10 rocks) of substrate in the stream bed.
- Indicate B for bedrock, S for sand/silt/clay (particles < 0.2 cm) and O for organic material.
- Embeddedness categories (E): Completely embedded = 1, 3/4 embedded, 1/2 embedded, 1/4 embedded, unembedded = 0

| Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E |
|---------------|------|---------------|------|---------------|------|---------------|-----|
| 1 | 6.0 | 26 | 14.0 | 51 | 16 | 76 | 23 |
| 2 | 7.0 | 27 | 23 | 52 | 8.5 | 77 | 19 |
| 3 | 13.0 | 28 | 3.5 | 53 | 22 | 78 | 10 |
| 4 | 19.0 | 29 | 35.0 | 54 | 16 | 79 | 37 |
| 5 | 1.0 | 30 | 23 | 55 | 7.5 | 80 | 6.5 |
| 6 | 12.0 | 31 | 15 | 56 | 28 | 81 | 13 |
| 7 | 8.0 | 32 | 7.5 | 57 | 17 | 82 | 28 |
| 8 | 9.5 | 33 | 34 | 58 | 19 | 83 | 9.0 |
| 9 | 2.0 | 34 | 29 | 59 | 18 | 84 | 14 |
| 10 | 18.0 | 35 | 33 | 60 | 10 | 85 | 15 |
| 11 | 6.0 | 36 | 15 | 61 | 18 | 86 | 8.0 |
| 12 | 15.0 | 37 | 15.5 | 62 | 30 | 87 | 20 |
| 13 | 5.5 | 38 | 7.0 | 63 | 21 | 88 | 10 |
| 14 | 1.5 | 39 | 30 | 64 | 17 | 89 | 22 |
| 15 | 4.5 | 40 | 4.0 | 65 | 11 | 90 | 17 |
| 16 | 6.0 | 41 | 35.0 | 66 | 11 | 91 | 19 |
| 17 | 15.5 | 42 | 23 | 67 | 18 | 92 | 8.5 |
| 18 | 7.0 | 43 | 18 | 68 | 4.0 | 93 | 2.5 |
| 19 | 8.5 | 44 | 8 | 69 | 13 | 94 | 10 |
| 20 | 8.0 | 45 | 15 | 70 | 9 | 95 | 11 |
| 21 | 26.0 | 46 | 22 | 71 | 27 | 96 | 3.5 |
| 22 | 33 | 47 | 16 | 72 | 13 | 97 | 9.0 |
| 23 | 11 | 48 | 27 | 73 | 13 | 98 | 14 |
| 24 | 21.5 | 49 | 23 | 74 | 9.5 | 99 | 20 |
| 25 | 9.0 | 50 | 19 | 75 | 10.0 | 100 | 9 |

Note: The Wolman D50 (i.e. median diameter), Wolman Dg (i.e. geometric mean diameter) and the % composition of the substrate classes will be calculated automatically in the CABIN database using the 100 pebble data. All 100 pebbles must be measured in order for the CABIN database tool to perform substrate calculations.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SITE INSPECTION

Site Inspected by: _____

Communication Information

☐ Itinerary left with contact person (include contact numbers)

Contact Person: _____ Time checked-in: _____

Form of communication: ☐ radio ☐ cell ☐ satellite ☐ hotel/pay phone ☐ SPOT

Phone number: () _____

Vehicle Safety

☐ Safety equipment (first aid, fire extinguisher, blanket, emergency kit in vehicle)

☐ Equipment and chemicals safely secured for transport

☐ Vehicle parked in safe location; pylons, hazard light, reflective vests if necessary

Notes:

Shore & Wading Safety

☐ Wading Task Hazard Analysis read by all field staff

☐ Wading Safe Work Procedures read by all field staff

☐ Instream hazards identified (i.e. log jams, deep pools, slippery rocks)

☐ PFD worn

☐ Appropriate footwear, waders, wading belt

☐ Belay used

Notes:

Field Crew: MS. J Kozikowska

Site Code: EGD-AQ11

Sampling Date: (DD/MM/YYYY) 30/07/2019 (1050am to 1240pm)

☐ Occupational Health & Safety: Site Inspection Sheet completed

PRIMARY SITE DATA

CABIN Study Name: _____ Local Basin Name: River of Golden Dreams

River/Stream Name: River of Golden Dreams Stream Order: (map scale 1:50,000) _____
(US site)

Select one: ☒ Test Site ☐ Potential Reference Site

Geographical Description/Notes:

valley
Lorimer Parking Lot - trail to south ~ 20m - through tree line
onto gravel bar

Surrounding Land Use: (check those present)

☒ Forest ☐ Field/Pasture
☐ Logging ☐ Mining

☐ Agriculture
☐ Commercial/Industrial

Information Source: _____

☒ Residential/Urban
☒ Other Outdoor Trails

Dominant Surrounding Land Use: (check one)

☐ Forest ☐ Field/Pasture
☐ Logging ☐ Mining

☐ Agriculture
☐ Commercial/Industrial

Information Source: _____

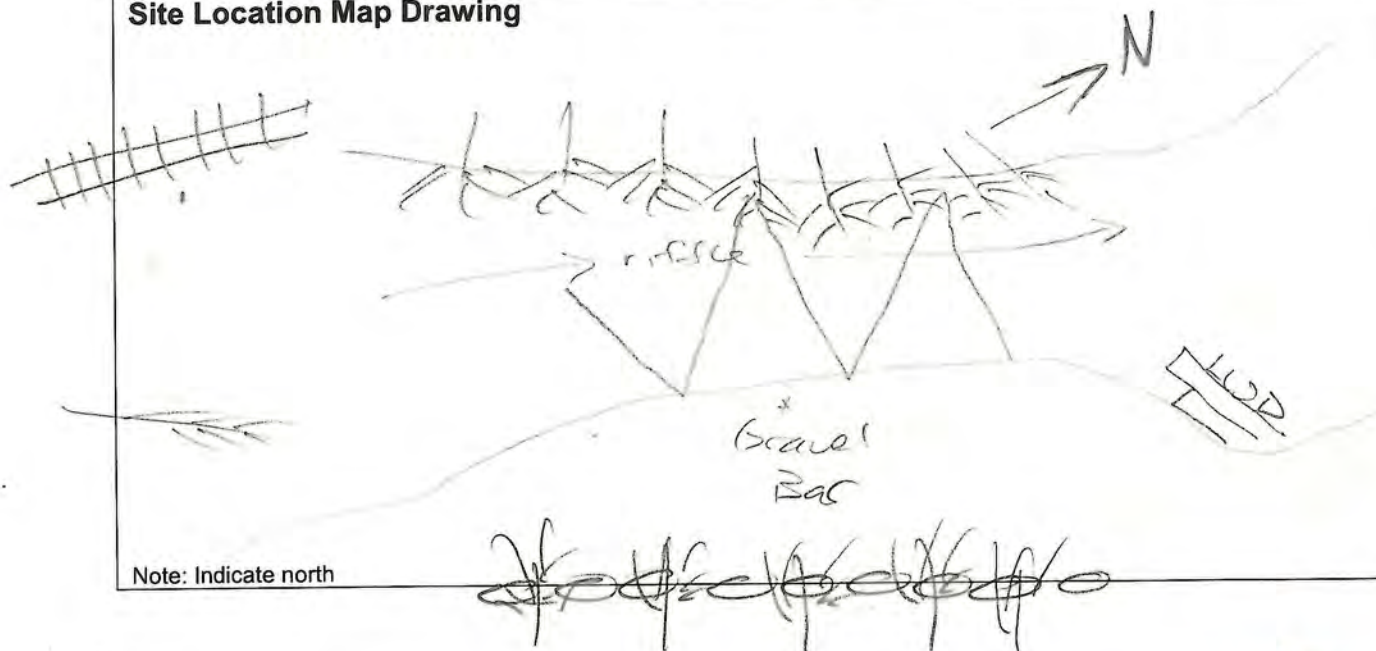
☒ Residential/Urban
☐ Other _____

Location Data 100 0502003 5552764

Latitude: _____ N Longitude: - _____ W (DMS or DD)

Elevation: 647 (fast or masl) GPS Datum: ☐ GRS80 (NAD83/WGS84) ☐ Other: _____

Site Location Map Drawing



Field Crew: _____

Site Code: RGD-AQ11Sampling Date: (DD/MM/YYYY) 30/07/2019**Photos**☒ Field Sheet☒ Upstream☒ Downstream☒ Across Site☐ Aerial View☐ Substrate (exposed)☒ Substrate (aquatic)☐ Other+ 2 Looking d/s w/ kayakers**REACH DATA** (represents 6 times bankfull width)

1. Habitat Types: (check those present)

☒ Riffle☐ Rapids☐ Straight run☒ Pool/Back Eddy

2. Canopy Coverage: (stand in middle of stream and look up, check one)

☐ 0 %☒ 1-25 %☐ 26-50 %☐ 51-75 %☐ 76-100 %

3. Macrophyte Coverage: (not algae or moss, check one)

☒ 0 %☐ 1-25 %☐ 26-50 %☐ 51-75 %☐ 76-100 %

4. Streamside Vegetation: (check those present)

☐ ferns/grasses☒ shrubs☒ deciduous trees☒ coniferous trees

5. Dominant Streamside Vegetation: (check one)

☐ ferns/grasses☒ shrubs☐ deciduous trees☐ coniferous trees

6. Periphyton Coverage on Substrate: (benthic algae, not moss, check one)

☒ 1 - Rocks are not slippery, no obvious colour (thin layer < 0.5 mm thick)☐ 2 - Rocks are slightly slippery, yellow-brown to light green colour (0.5-1 mm thick)☐ 3 - Rocks have a noticeable slippery feel (footing is slippery), with patches of thicker green to brown algae (1-5 mm thick)☐ 4 - Rocks are very slippery (algae can be removed with thumbnail), numerous large clumps of green to dark brown algae (5 mm -20 mm thick)☐ 5 - Rocks are mostly obscured by algal mat, extensive green, brown to black algal mass may have long strands (> 20 mm thick)

Note: 1 through 5 represent categories entered into the CABIN database.

BENTHIC MACROINVERTEBRATE DATAHabitat sampled: (check one) ☐ riffle ☐ rapids ☐ straight run

| | |
|---------------------------------|-------------------|
| 400 µm mesh Kick Net | |
| Person sampling | <u>M.S.</u> |
| Sampling time (i.e. 3 min.) | <u>3 min.</u> |
| No. of sample jars | <u>1 + 2 QAQC</u> |
| Typical depth in kick area (cm) | <u>15 cm</u> |

Preservative used: Ethanol

Sampled sieved on site using "Bucket Swirling Method":

☒ YES ☐ NOIf YES, debris collected for QAQC ☒ 2 jars+ 3 sculpin in sample

Note: Indicate if a sampling method other than the recommended 400 µm mesh kick net is used.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

WATER CHEMISTRY DATA Time: 12.15 (24 hr clock) Time zone: _____

Air Temp: 12 (°C) Water Temp: 12.8 (°C) pH: 6.77

Specific Conductance: 44.3 (µs/cm) DO: 92.7% (mg/L) Turbidity: _____ (NTU)

conductivity 33.3 µs/cm

9.81

TDS - 29.25

Check if water samples were collected for the following analyses:

- ☐ TSS (Total Suspended Solids)
- ☐ Nitrogen (i.e. Total, Nitrate, Nitrite, Dissolved, and/or Ammonia)
- ☐ Phosphorus (Total, Ortho, and/or Dissolved)
- ☐ Major Ions (i.e. Alkalinity, Hardness, Chloride, and/or Sulphate) ☐ Other _____

MnHg - 705.7

Note: Determining alkalinity is recommended, as are other analyses, but not required for CABIN assessments.

CHANNEL DATA

Slope - Indicate how slope was measured: (check one)

- ☐ **Calculated from map**

Scale: _____ (Note: small scale map recommended if field measurement is not possible - i.e. 1:20,000),
contour interval (vertical distance) _____ (m),
distance between contour intervals (horizontal distance) _____ (m)
slope = vertical distance/horizontal distance = _____

OR

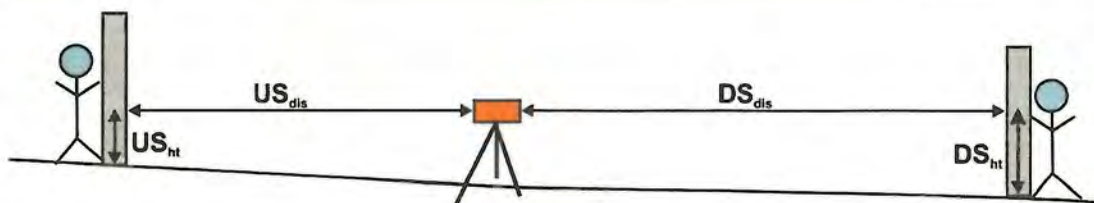
- ☐ **Measured in field**

Circle device used and fill out table according to device:

a. Survey Equipment b. Hand Level & Measuring Tape

Clinometer 11. Over 19m

| Measurements | Upstream (U/S) | Downstream(D/S) | Calculation |
|-----------------------------------|-------------------------------------|-------------------------------------|--|
| ^a Top Hairline (T) | | | |
| ^a Mid Hairline (ht) OR | | | |
| ^b Height of rod | | | |
| ^a Bottom Hairline (B) | | | |
| ^b Distance (dis) OR | | | US _{dis} +DS _{dis} = |
| ^a T-B x 100 | ^a US _{dis} =T-B | ^a DS _{dis} =T-B | |
| Change in height (Δht) | | | DS _{ht} -US _{ht} = |
| Slope (Δht/total dis) | | | |



Field Crew: _____ Site Code: _____

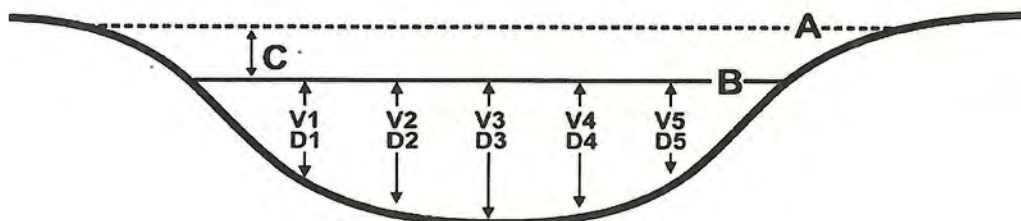
Sampling Date: (DD/MM/YYYY) _____

Widths and Depth

Location at site: middle of kick area (Indicate where in sample reach, ex. d/s of kick area)

A - Bankfull Width: 17.30 (m) B - Wetted Stream Width: 6.60 (m)

C - Bankfull-Wetted Depth (height from water surface to Bankfull): 33 (cm)



Note:

Wetted widths > 5 m, measure a minimum of 5-6 equidistant locations;
Wetted widths < 5 m, measure 3-4 equidistant locations.

Velocity and Depth

Check appropriate velocity measuring device and fill out the appropriate section in chart below. Distance from shore and depth are required regardless of method:

☒ **Velocity Head Rod (or ruler):** Velocity Equation (m/s) = $\sqrt{2(\Delta D/100) * 9.81}$

☐ **Rotary meters:** Gurley/Price/Mini-Price/Propeller (Refer to specific meter conversion chart for calculation)

☐ **Direct velocity measurements:** ☐ Marsh-McBirney ☐ Sontek or ☐ Other _____

Gravel Bar. →

| | 1 | 2 | 3 | 4 | 5 | 6 | AVG |
|---|------|-----|------|-----|-----|------|-----|
| Distance from Shore (m) | 0.5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | |
| Depth (D) (cm) | 9.5 | 23 | 31 | 31 | 29 | 17 | |
| Velocity Head Rod (ruler) | | | | | | | |
| Flowing water Depth (D ₁) (cm) | 9.5 | 23 | 31 | 31 | 29 | 17 | |
| Depth of Stagnation (D ₂) (cm) | 11.5 | 26 | 36.5 | 38 | 33 | 16.5 | |
| Change in depth ($\Delta D = D_2 - D_1$) (cm) | | | | | | | |
| Rotary meter | | | | | | | |
| Revolutions | | | | | | | |
| Time (minimum 40 seconds) | | | | | | | |
| Direct Measurement or calculation | | | | | | | |
| Velocity (V) (m/s) | | | | | | | |

Field Crew: _____

Site Code: RGD-AQ11Sampling Date: (DD/MM/YYYY) 30/07/2019**SUBSTRATE DATA****Surrounding/Interstitial Material**

Circle the substrate size category for the surrounding material.

| Substrate Size Class | Category |
|------------------------------------|----------|
| Organic Cover | 0 |
| < 0.1 cm (fine sand, silt or clay) | 1 |
| 0.1-0.2 cm (coarse sand) | 2 |
| 0.2-1.6 cm (gravel) | 3 |
| 1.6-3.2 cm (small pebble) | 4 |
| 3.2-6.4 cm (large pebble) | 5 |
| 6.4-12.8 cm (small cobble) | 6 |
| 12.8-25.6 cm (cobble) | 7 |
| > 25.6 cm (boulder) | 8 |
| Bedrock | 9 |

100 Pebble Count & Substrate Embeddedness

- Measure the intermediate axis (100 rocks) and embeddedness (10 rocks) of substrate in the stream bed.
- Indicate B for bedrock, S for sand/silt/clay (particles < 0.2 cm) and O for organic material.
- Embeddedness categories (E): Completely embedded = 1, 3/4 embedded, 1/2 embedded, 1/4 embedded, unembedded = 0

| Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E | Diameter (cm) | E |
|---------------|-----|---------------|-----|---------------|-----|---------------|-----|
| 1 | 2.2 | 26 | 4.4 | 51 | 3.6 | 76 | 3.5 |
| 2 | 2.5 | 27 | 5.4 | 52 | 4.0 | 77 | 1.5 |
| 3 | 2.1 | 28 | 6.0 | 53 | 3.0 | 78 | 1.0 |
| 4 | 1.5 | 29 | 6.0 | 54 | 3.0 | 79 | 1.1 |
| 5 | 1.7 | 30 | 4.6 | 55 | 2.2 | 80 | 3.5 |
| 6 | 2.3 | 31 | 7.4 | 56 | 3.1 | 81 | 3.0 |
| 7 | 5.8 | 32 | 5.2 | 57 | 4.4 | 82 | 2.8 |
| 8 | 5.5 | 33 | 6.2 | 58 | 2.1 | 83 | 1.1 |
| 9 | 3.6 | 34 | 2.3 | 59 | 2.9 | 84 | 1.5 |
| 10 | 1.6 | 35 | 6.9 | 60 | 4.2 | 85 | 5.8 |
| 11 | 3.5 | 36 | 2.2 | 61 | 3.6 | 86 | 3.0 |
| 12 | 4.5 | 37 | 2.3 | 62 | 4.5 | 87 | 1.4 |
| 13 | 4.5 | 38 | 4.3 | 63 | 3.1 | 88 | 0.6 |
| 14 | 1.5 | 39 | 1.6 | 64 | 5.5 | 89 | 4.0 |
| 15 | 8 | 40 | 2.8 | 65 | 1.5 | 90 | 1.4 |
| 16 | 4.1 | 41 | 6.2 | 66 | 1.2 | 91 | 1.0 |
| 17 | 3.4 | 42 | 3.3 | 67 | 6.1 | 92 | 2.0 |
| 18 | 3.8 | 43 | 1.9 | 68 | 2.5 | 93 | 0.7 |
| 19 | 3.0 | 44 | 1.1 | 69 | 4.2 | 94 | 2.3 |
| 20 | 2.5 | 45 | 4.5 | 70 | 5.1 | 95 | 3.8 |
| 21 | 3.2 | 46 | 5.4 | 71 | 1.8 | 96 | 3.0 |
| 22 | 2.0 | 47 | 4.0 | 72 | 3.0 | 97 | 3.0 |
| 23 | 3.7 | 48 | 3.0 | 73 | 2.9 | 98 | 5.0 |
| 24 | 1.3 | 49 | 0.7 | 74 | 1.9 | 99 | 1.2 |
| 25 | 3.2 | 50 | 2.0 | 75 | 3.2 | 100 | 2.3 |

Note: The Wolman D50 (i.e. median diameter), Wolman Dg (i.e. geometric mean diameter) and the % composition of the substrate classes will be calculated automatically in the CABIN database using the 100 pebble data. All 100 pebbles must be measured in order for the CABIN database tool to perform substrate calculations.

Field Crew: _____ Site Code: _____
Sampling Date: (DD/MM/YYYY) _____

SITE INSPECTION

Site Inspected by: _____

Communication Information

☐ Itinerary left with contact person (include contact numbers)

Contact Person: _____ Time checked-in: _____

Form of communication: ☐ radio ☐ cell ☐ satellite ☐ hotel/pay phone ☐ SPOT

Phone number: () _____

Vehicle Safety

☐ Safety equipment (first aid, fire extinguisher, blanket, emergency kit in vehicle)

☐ Equipment and chemicals safely secured for transport

☐ Vehicle parked in safe location; pylons, hazard light, reflective vests if necessary

Notes:

Shore & Wading Safety

☐ Wading Task Hazard Analysis read by all field staff

☐ Wading Safe Work Procedures read by all field staff

☐ Instream hazards identified (i.e. log jams, deep pools, slippery rocks)

☐ PFD worn

☐ Appropriate footwear, waders, wading belt

☐ Belay used

Notes:

Field Crew: MS, H.W Site Code: PGD-DS-AQ 12
Sampling Date: (DD/MM/YYYY) 31/07/2019

☐ Occupational Health & Safety: Site Inspection Sheet completed

PRIMARY SITE DATA

CABIN Study Name: _____ Local Basin Name: River of Golden Dreams (D/S)

River/Stream Name: PGD Stream Order: (map scale 1:50,000) _____

Select one: ☒ Test Site ☐ Potential Reference Site

Geographical Description/Notes:

Park @ Nicholas Golf Course - trail / Boat Pull out location.

Surrounding Land Use: (check those present)

☐ Forest ☐ Field/Pasture ☐ Agriculture ☒ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☐ Other Forest being developed Golf Course.

Information Source: _____

Dominant Surrounding Land Use: (check one)

☐ Forest ☐ Field/Pasture ☐ Agriculture ☒ Residential/Urban
☐ Logging ☐ Mining ☐ Commercial/Industrial ☐ Other _____

Information Source: _____

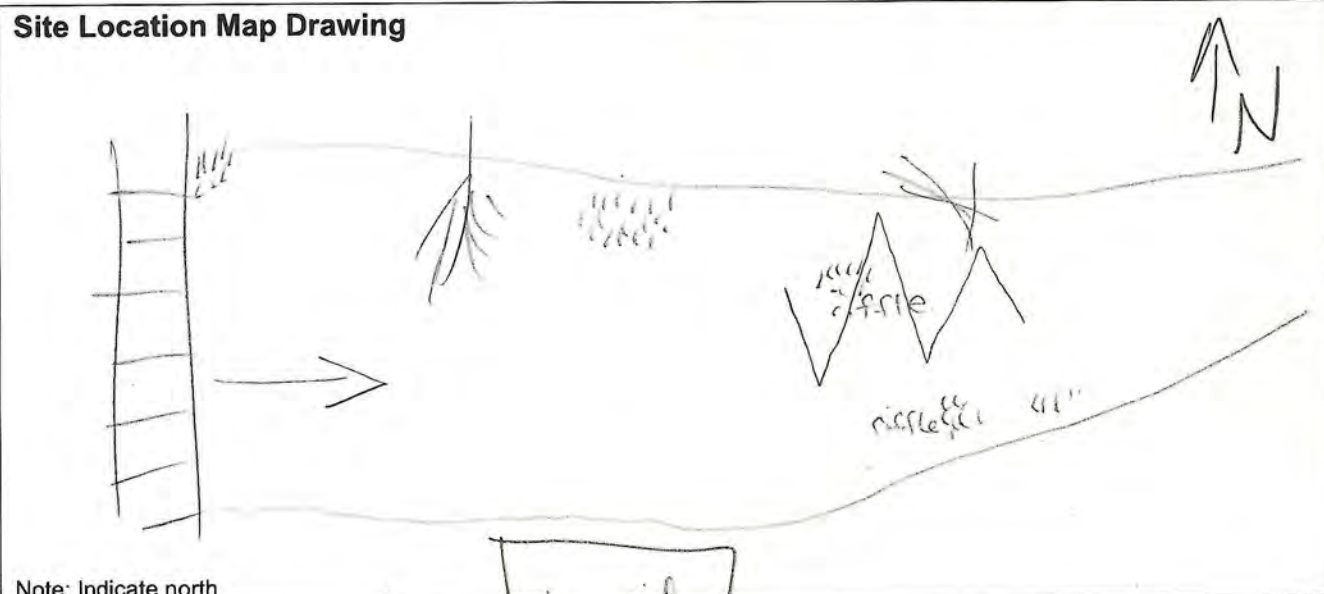
Location Data

10 U 05 03 027 5554676

Latitude: _____ N Longitude: - _____ W (DMS or DD)

Elevation: 614 (ft or masl) GPS Datum: ☐ GRS80 (NAD83/WGS84) ☐ Other: _____

Site Location Map Drawing



Field Crew: MS + HWSite Code: RGD-DS-AQ12Sampling Date: (DD/MM/YYYY) 31/07/2019**Photos**

- ☒ Field Sheet ☒ Upstream ☒ Downstream ☒ Across Site ☐ Aerial View
☐ Substrate (exposed) ☒ Substrate (aquatic) ☐ Other _____

REACH DATA (represents 6 times bankfull width)

1. Habitat Types: (check those present)

- ☒ Riffle ☐ Rapids ☐ Straight run ☒ Pool/Back Eddy

2. Canopy Coverage: (stand in middle of stream and look up, check one)

- ☐ 0 % ☒ 1-25 % ☐ 26-50 % ☐ 51-75 % ☐ 76-100 %

3. Macrophyte Coverage: (not algae or moss, check one)

- ☐ 0 % ☒ 1-25 % ☐ 26-50 % ☐ 51-75 % ☐ 76-100 %

4. Streamside Vegetation: (check those present)

- ☒ ferns/grasses ☒ shrubs ☒ deciduous trees ☐ coniferous trees

5. Dominant Streamside Vegetation: (check one)

- ☐ ferns/grasses ☐ shrubs ☒ deciduous trees ☐ coniferous trees

6. Periphyton Coverage on Substrate: (benthic algae, not moss, check one)

- ☐ 1 - Rocks are not slippery, no obvious colour (thin layer < 0.5 mm thick)
☐ 2 - Rocks are slightly slippery, yellow-brown to light green colour (0.5-1 mm thick)
☐ 3 - Rocks have a noticeable slippery feel (footing is slippery), with patches of thicker green to brown algae (1-5 mm thick)
☒ 4 - Rocks are very slippery (algae can be removed with thumbnail), numerous large clumps of green to dark brown algae (5 mm -20 mm thick)
☐ 5 - Rocks are mostly obscured by algal mat, extensive green, brown to black algal mass may have long strands (> 20 mm thick)
- random rocks w/ clumps of moss

Note: 1 through 5 represent categories entered into the CABIN database.

BENTHIC MACROINVERTEBRATE DATAHabitat sampled: (check one) ☒ riffle ☐ rapids ☐ straight run

| | |
|---------------------------------|-------|
| 400 µm mesh Kick Net | |
| Person sampling | MS |
| Sampling time (i.e. 3 min.) | 3 min |
| No. of sample jars | 1 |
| Typical depth in kick area (cm) | |

Preservative used: Ethanol

Sampled sieved on site using "Bucket Swirling Method":

☒ YES ☐ NOIf YES, debris collected for QAQC ☐

Note: Indicate if a sampling method other than the recommended 400 µm mesh kick net is used.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) 31/07/2019

WATER CHEMISTRY DATA Time: 1005 (24 hr clock) Time zone: _____

Air Temp: -17 (°C) Water Temp: 13.1 (°C) pH: 7.56 pHmV = -41.03

Specific Conductance: 78.4 (µs/cm) DO: 9.93 (mg/L) Turbidity: — (NTU)

Conductivity 60.6 µs/cm 94.3 %

Check if water samples were collected for the following analyses:

- ☐ TSS (Total Suspended Solids)
- ☐ Nitrogen (i.e. Total, Nitrate, Nitrite, Dissolved, and/or Ammonia)
- ☐ Phosphorus (Total, Ortho, and/or Dissolved)
- ☐ Major Ions (i.e. Alkalinity, Hardness, Chloride, and/or Sulphate) ☐ Other _____

707.4 mmHg
TDS 50.7
Salinity 0.04 ppt

Note: Determining alkalinity is recommended, as are other analyses, but not required for CABIN assessments.

CHANNEL DATA

Slope - Indicate how slope was measured: (check one)

☐ **Calculated from map**

Scale: _____ (Note: small scale map recommended if field measurement is not possible - i.e. 1:20,000).
contour interval (vertical distance) _____ (m),
distance between contour intervals (horizontal distance) _____ (m)
slope = vertical distance/horizontal distance = _____

OR

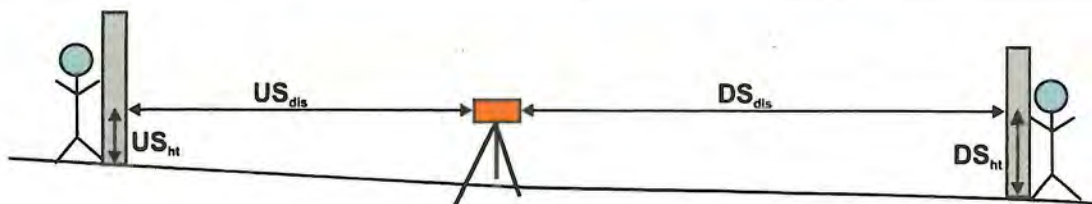
☒ **Measured in field**

Circle device used and fill out table according to device:

a. Survey Equipment b. Hand Level & Measuring Tape

Clinometer 0% over 25m

| Measurements | Upstream (U/S) | Downstream(D/S) | Calculation |
|-----------------------------------|-------------------------------------|-------------------------------------|--|
| ^a Top Hairline (T) | | | |
| ^a Mid Hairline (ht) OR | | | |
| ^b Height of rod | | | |
| ^a Bottom Hairline (B) | | | |
| ^b Distance (dis) OR | | | US _{dis} +DS _{dis} = |
| ^a T-B x 100 | ^a US _{dis} =T-B | ^a DS _{dis} =T-B | |
| Change in height (Δht) | | | DS _{ht} -US _{ht} = |
| Slope (Δht/total dis) | | | |



Field Crew: _____ Site Code: _____

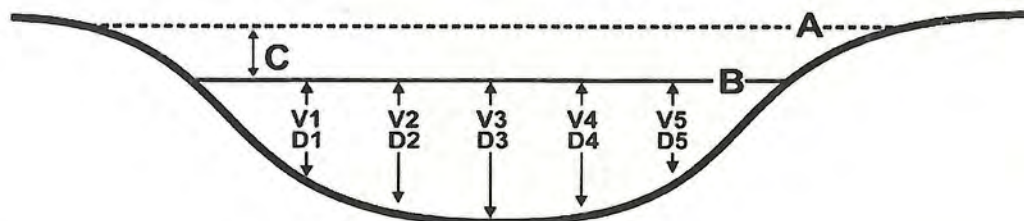
Sampling Date: (DD/MM/YYYY) _____

Widths and Depth

Location at site: across kicknet site (Indicate where in sample reach, ex. d/s of kick area)

A - Bankfull Width: 15.6 (m) B - Wetted Stream Width: 15.5 (m)

C - Bankfull-Wetted Depth (height from water surface to Bankfull): 7cm (cm)



Note:

Wetted widths > 5 m, measure a minimum of 5-6 equidistant locations;

Wetted widths < 5 m, measure 3-4 equidistant locations.

Velocity and Depth

Check appropriate velocity measuring device and fill out the appropriate section in chart below. Distance from shore and depth are required regardless of method:

☒ **Velocity Head Rod (or ruler):** Velocity Equation (m/s) = $\sqrt{[2(\Delta D/100) * 9.81]}$

☐ **Rotary meters:** Gurley/Price/Mini-Price/Propeller (Refer to specific meter conversion chart for calculation)

☐ **Direct velocity measurements:** ☐ Marsh-McBirney ☐ Sontek or ☐ Other _____

Base →

Behind veg patch.

| | 1 | 2 | 3 | 4 | 5 | 6 | AVG |
|---|-------------|-------------|-------------|-------------|-------------|-------------|-----|
| Distance from Shore (m) | <u>2m.</u> | <u>4</u> | <u>6</u> | <u>8</u> | <u>10</u> | <u>12m.</u> | |
| Depth (D) (cm) | | | | | | | |
| Velocity Head Rod (ruler) | | | | | | | |
| Flowing water Depth (D ₁) (cm) | <u>54.5</u> | <u>39.0</u> | <u>33.0</u> | <u>24.5</u> | <u>25.5</u> | <u>27.0</u> | |
| Depth of Stagnation (D ₂) (cm) | <u>54.5</u> | <u>40.0</u> | <u>34.5</u> | <u>24.5</u> | <u>26.0</u> | <u>27.0</u> | |
| Change in depth ($\Delta D = D_2 - D_1$) (cm) | | | | | | | |
| Rotary meter | | | | | | | |
| Revolutions | | | | | | | |
| Time (minimum 40 seconds) | | | | | | | |
| Direct Measurement or calculation | | | | | | | |
| Velocity (V) (m/s) | | | | | | | |

Site Code: _____

SUBSTRATE DATA

Circle the substrate size category for the surrounding material.

| Substrate Size Class | Category |
|------------------------------------|----------|
| Organic Cover | 0 |
| < 0.1 cm (fine sand, silt or clay) | 1 |
| 0.1-0.2 cm (coarse sand) | 2 |
| 0.2-1.6 cm (gravel) | 3 |
| 1.6-3.2 cm (small pebble) | 4 |
| 3.2-6.4 cm (large pebble) | 5 |
| 6.4-12.8 cm (small cobble) | 6 |
| 12.8-25.6 cm (cobble) | 7 |
| > 25.6 cm (boulder) | 8 |
| Bedrock | 9 |

- Measure the intermediate axis (100 rocks) and embeddedness (10 rocks) of substrate in the stream bed.
- Indicate B for bedrock, S for sand/silt/clay (particles < 0.2 cm) and O for organic material.
- Embeddedness categories (E): Completely embedded = 1, 3/4 embedded, 1/2 embedded, 1/4 embedded, unembedded = 0

| | Diameter (cm) | E | | Diameter (cm) | E | | Diameter (cm) | E | | Diameter (cm) | E |
|----|---------------|-----|----|---------------|-----|----|---------------|-----|-----|---------------|-----|
| 1 | 2.1 | | 26 | 3.8 | | 51 | 2.4 | | 76 | 1.5 | |
| 2 | 4.2 | | 27 | S | | 52 | 2.6 | | 77 | 2.8 | |
| 3 | 0.5 | | 28 | 4.0 | | 53 | S | | 78 | 1.7 | |
| 4 | 3.5 | | 29 | 2.2 | | 54 | S | | 79 | 4.0 | |
| 5 | 3.8 | | 30 | 2.6 | 3/4 | 55 | S | | 80 | 1.6 | 1/4 |
| 6 | S | | 31 | 2.2 | | 56 | 0.6 | | 81 | 1.2 | |
| 7 | 2.1 | | 32 | 2.3 | | 57 | 1.4 | | 82 | 3.0 | |
| 8 | 3.4 | | 33 | 3.8 | | 58 | 1.0 | | 83 | 2.5 | |
| 9 | 4.1 | | 34 | 0.9 | | 59 | S | | 84 | 1.6 | |
| 10 | 3.4 | 3/4 | 35 | S | | 60 | S | | 85 | 3.4 | |
| 11 | 1.5 | | 36 | S | | 61 | 3.1 | 3/4 | 86 | 1.9 | |
| 12 | 2.4 | | 37 | 0.7 | | 62 | 4.5 | | 87 | 2.0 | |
| 13 | 1.1 | | 38 | S | | 63 | 1.1 | | 88 | 2.6 | |
| 14 | 1.8 | | 39 | 1.1 | | 64 | 3.5 | | 89 | 3.0 | |
| 15 | 2.0 | | 40 | 1.6 | 3/4 | 65 | 3.0 | | 90 | 3.4 | 3/4 |
| 16 | 0.9 | | 41 | 0.8 | | 66 | 3.8 | | 91 | 2.2 | |
| 17 | 0.7 | | 42 | 2.7 | | 67 | 2.7 | | 92 | S | |
| 18 | 2.3 | | 43 | 3.6 | | 68 | 0.9 | | 93 | 1.4 | |
| 19 | 1.2 | | 44 | 1.2 | | 69 | 1.5 | | 94 | 0.6 | |
| 20 | 2.7 | 1 | 45 | 0.6 | | 70 | 1.2 | 1/4 | 95 | 1.3 | |
| 21 | 2.8 | | 46 | 3.5 | | 71 | 4.2 | | 96 | 1.3 | |
| 22 | 0.6 | | 47 | 2.6 | | 72 | 0.2 | | 97 | 0.5 | |
| 23 | 1.3 | | 48 | C | | 73 | 2.4 | | 98 | 2.0 | |
| 24 | 2.2 | | 49 | 4.5 | | 74 | 1.9 | | 99 | 2.1 | |
| 25 | 0.9 | | 50 | 2.2 | 3/4 | 75 | 2.0 | | 100 | 2.6 | 3/4 |

Note: The Wolman D50 (i.e. median diameter), Wolman Dg (i.e. geometric mean diameter) and the % composition of the substrate classes will be calculated automatically in the CABIN database using the 100 pebble data. All 100 pebbles must be measured in order for the CABIN database tool to perform substrate calculations.

Field Crew: _____ Site Code: _____

Sampling Date: (DD/MM/YYYY) _____

SITE INSPECTION

Site Inspected by: _____

Communication Information

☐ Itinerary left with contact person (include contact numbers)

Contact Person: _____ Time checked-in: _____

Form of communication: ☐ radio ☐ cell ☐ satellite ☐ hotel/pay phone ☐ SPOT

Phone number: () _____

Vehicle Safety

☐ Safety equipment (first aid, fire extinguisher, blanket, emergency kit in vehicle)

☐ Equipment and chemicals safely secured for transport

☐ Vehicle parked in safe location; pylons, hazard light, reflective vests if necessary

Notes:

Shore & Wading Safety

☐ Wading Task Hazard Analysis read by all field staff

☐ Wading Safe Work Procedures read by all field staff

☐ Instream hazards identified (i.e. log jams, deep pools, slippery rocks)

☐ PFD worn

☐ Appropriate footwear, waders, wading belt

☐ Belay used

Notes:

Appendix D

Fish Sampling Datasheets & Biological Characteristics

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|--------|-----------------|-----------------|------------|---------|--------|----------------|------------|---------------------|
| CRB-01 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | TR | EF | 136 | 23.3 | Photo #98-106 |
| CRB-02 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | TR | EF | 116 | 13.1 | Photo 107-108 |
| CRB-03 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 65 | 2.4 | 13 anal fin rays |
| CRB-04 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 77 | 5.7 | 13 anal fin rays |
| CRB-05 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 64 | 2.4 | |
| CRB-06 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 88 | 8.3 | |
| CRB-07 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | SB | EF | 53 | 1.8 | Photo 109-113 |
| CRB-08 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 64 | 2.8 | |
| CRB-09 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 62 | 2.6 | |
| CRB-10 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 68 | 3.7 | |
| CRB-11 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 77 | 4.6 | |
| CRB-12 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 57 | 1.7 | |
| CRB-13 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 56 | 2 | |
| CRB-14 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 49 | 1 | |
| CRB-15 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 43 | 0.7 | |
| CRB-16 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 60 | 2.4 | |
| CRB-17 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 72 | 4.1 | |
| CRB-18 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 59 | 2 | |
| CRB-19 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 42 | 0.6 | |
| CRB-20 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 54 | 1.5 | |
| CRB-21 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 58 | 2 | |
| CRB-22 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 55 | 1.8 | |
| CRB-23 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 50 | 1.4 | |
| CRB-24 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 58 | 1.8 | |
| CRB-25 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 50 | 1.3 | |
| CRB-26 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 47 | 1 | |
| CRB-27 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 55 | 1.8 | |
| CRB-28 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 47 | 0.9 | |
| CRB-29 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | CC | EF | 45 | 0.8 | |
| CRB-30 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | TR | EF | 42 | 0.6 | |
| CRB-31 | Crabapple Creek | Crabapple Creek | Aug-1-2019 | TR | EF | 28 | 0.1 | rounded up on scale |
| CRB-32 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | TR | MT | 107 | 16.1 | Photo on phone |

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|-----------|------------------------|-----------------|------------|---------|--------|----------------|------------|----------|
| CRB-33 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | TR | MT | 94 | 8.1 | |
| CRB-34 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | TR | MT | 94 | 8.1 | |
| CRB-35 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | TR | MT | 72 | 3.6 | |
| CRB-36 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | SB | MT | 65 | 3.3 | |
| CRB-37 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | SB | MT | 55 | 2 | |
| CRB-38 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | SB | MT | 52 | 1.9 | |
| CRB-39 | Crabapple Creek | Crabapple Creek | Aug-2-2019 | SB | MT | 70 | 3.2 | |
| RGD-DS-01 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.8 | |
| RGD-DS-02 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 46 | 1.5 | |
| RGD-DS-03 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 43 | 0.7 | |
| RGD-DS-04 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 40 | 0.6 | |
| RGD-DS-05 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.7 | |
| RGD-DS-06 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 43 | 0.7 | |
| RGD-DS-07 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 43 | 0.8 | |
| RGD-DS-08 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.9 | |
| RGD-DS-09 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 1 | |
| RGD-DS-10 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 43 | 0.7 | |
| RGD-DS-11 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 40 | 0.7 | |
| RGD-DS-12 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 40 | 0.7 | |
| RGD-DS-13 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 0.9 | |
| RGD-DS-14 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 54 | 1.6 | |
| RGD-DS-15 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 46 | 1 | |
| RGD-DS-16 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 40 | 0.7 | |
| RGD-DS-17 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.9 | |
| RGD-DS-18 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 40 | 0.8 | |
| RGD-DS-19 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 37 | 0.5 | |
| RGD-DS-20 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 37 | 0.4 | |
| RGD-DS-21 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 48 | 0.9 | |
| RGD-DS-22 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 0.9 | |
| RGD-DS-23 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 0.9 | |
| RGD-DS-24 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 49 | 1 | |
| RGD-DS-25 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 0.9 | |

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|-----------|------------------------|-------------|------------|---------|--------|----------------|------------|----------|
| RGD-DS-26 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 55 | 1.7 | |
| RGD-DS-27 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.8 | |
| RGD-DS-28 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.7 | |
| RGD-DS-29 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 51 | 1.4 | |
| RGD-DS-30 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 1 | |
| RGD-DS-31 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 1 | |
| RGD-DS-32 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 50 | 1.2 | |
| RGD-DS-33 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 39 | 0.6 | |
| RGD-DS-34 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 45 | 0.9 | |
| RGD-DS-35 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 1 | |
| RGD-DS-36 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 33 | 0.3 | |
| RGD-DS-37 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 37 | 0.5 | |
| RGD-DS-38 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.7 | |
| RGD-DS-39 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 35 | 0.5 | |
| RGD-DS-40 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 39 | 0.6 | |
| RGD-DS-41 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.8 | |
| RGD-DS-42 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.6 | |
| RGD-DS-43 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.9 | |
| RGD-DS-44 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.8 | |
| RGD-DS-45 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 39 | 0.5 | |
| RGD-DS-46 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.9 | |
| RGD-DS-47 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 36 | 0.4 | |
| RGD-DS-48 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.8 | |
| RGD-DS-49 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.7 | |
| RGD-DS-50 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.7 | |
| RGD-DS-51 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 44 | 0.7 | |
| RGD-DS-52 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 42 | 0.7 | |
| RGD-DS-53 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 41 | 0.6 | |
| RGD-DS-54 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 37 | 0.5 | |
| RGD-DS-55 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | SB | MT | 43 | 0.7 | |
| RGD-DS-56 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 60 | 2.1 | |
| RGD-DS-57 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 55 | 1.7 | |

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|-----------|------------------------|-------------|-------------|---------|--------|----------------|------------|--|
| RGD-DS-58 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 63 | 2.8 | |
| RGD-DS-59 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 65 | 2.8 | |
| RGD-DS-60 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 58 | 3.3 | 13 anal fin rays |
| RGD-DS-61 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | CC | MT | 57 | 2 | |
| RGD-DS-62 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | TR | MT | 73 | 3.6 | |
| RGD-DS-63 | River of Golden Dreams | RGD-DS-AQ12 | Aug-1-2019 | TR | MT | 84 | 5.8 | |
| RGD-01 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | SB | MT | 59 | 1.7 | |
| RGD-02 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | SB | MT | 50 | 1.1 | |
| RGD-03 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | SB | MT | 61 | 2.6 | |
| RGD-04 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | SB | MT | 48 | 1 | |
| RGD-05 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | SB | MT | 50 | 1.7 | |
| RGD-06 | River of Golden Dreams | RGD-AQ11 | Aug-2-2019 | CC | MT | 80 | 6.2 | |
| 21M-01 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 81 | | 7 Picture 77-78 18-19 cadual rays, 15 |
| 21M-02 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 76 | 11.1 | anal rays, pic 79-86 |
| 21M-03 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 70 | 4.8 | |
| 21M-04 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 67 | 3.7 | Picture 87 |
| 21M-05 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 49 | 1.8 | |
| 21M-06 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 61 | 3.1 | |
| 21M-07 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 48 | 2 | |
| 21M-08 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 67 | 4.1 | |
| 21M-09 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 55 | 2.2 | |
| 21M-10 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 45 | 0.8 | |
| 21M-11 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 61 | 2.6 | |
| 21M-12 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 50 | 1.5 | |
| 21M-13 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 60 | 2.3 | Picture 88-89 |
| 21M-14 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 57 | 2.3 | |
| 21M-15 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 42 | 1.3 | |
| 21M-16 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 47 | 1.2 | |
| 21M-17 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 45 | 1.1 | |
| 21M-18 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 31 | 0.2 | |
| 21M-19 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 43 | 1 | |

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|--------|---------------|-----------|-------------|---------|--------|----------------|------------|---------------------|
| 21M-20 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 45 | 1.1 | |
| 21M-21 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 49 | 2 | |
| 21M-22 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 56 | 2.8 | |
| 21M-23 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 51 | 1.2 | |
| 21M-24 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 29 | 0.2 | |
| 21M-25 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 27 | 0.1 | rounded up on scale |
| 21M-26 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 39 | 0.4 | |
| 21M-27 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 31 | 0.1 | |
| 21M-28 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 43 | 1 | |
| 21M-29 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 47 | 1.6 | |
| 21M-30 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 46 | 1.3 | |
| 21M-31 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 53 | 1.1 | |
| 21M-32 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 47 | 1.3 | |
| 21M-33 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 56 | 1.6 | |
| 21M-34 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 42 | 0.5 | |
| 21M-35 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 29 | 0.1 | |
| 21M-36 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 31 | 0.2 | |
| 21M-37 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 38 | 0.1 | |
| 21M-38 | 21 Mile Creek | 21 Mile | Jul-31-2019 | TR | EF | 30 | 0.2 | |
| 21M-39 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 4.1 | 1.2 | |
| 21M-40 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 44 | 1.5 | |
| 21M-41 | 21 Mile Creek | 21 Mile | Jul-31-2019 | CC | EF | 43 | 0.7 | |
| 21M-42 | 21 Mile Creek | 21 Mile | Aug-2-2019 | TR | MT | 87 | 7 | |
| 21M-43 | 21 Mile Creek | 21 Mile | Aug-2-2019 | TR | MT | 81 | 6.3 | |
| 21M-44 | 21 Mile Creek | 21 Mile | Aug-2-2019 | TR | MT | 75 | 4.5 | |
| 21M-45 | 21 Mile Creek | 21 Mile | Aug-2-2019 | TR | MT | 67 | 3.8 | mortality |
| 21M-46 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 51 | 1.8 | |
| 21M-47 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 52 | 1.7 | |
| 21M-48 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 52 | 1.7 | |
| 21M-49 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 46 | 0.6 | |
| 21M-50 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 46 | 0.7 | |

| ID | Location | Site Name | Date | Species | Method | Legnth (mm) | weight (g) | Comments |
|--------|---------------|--------------|------------|---------|--------|----------------|------------|---------------------------|
| 21M-51 | 21 Mile Creek | 21 Mile | Aug-2-2019 | SB | MT | 42 | na | mortality (innards eaten) |
| JOR-01 | Jordan Creek | Jordan Creek | Aug-1-2019 | TR | MT | 104 | 13.2 | pic 116-117 |
| JOR-02 | Jordan Creek | Jordan Creek | Aug-1-2019 | TR | MT | 86 | 7.5 | |
| JOR-03 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 49 | 1.5 | |
| JOR-04 | Jordan Creek | Jordan Creek | Aug-1-2019 | CC | MT | 94 | 11.2 | 13 anal fin rays |
| JOR-05 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 48 | 1.5 | |
| JOR-06 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 57 | 2 | |
| JOR-07 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 51 | 1.4 | |
| JOR-08 | Jordan Creek | Jordan Creek | Aug-1-2019 | TR | MT | 47 | 1.3 | |
| JOR-09 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 46 | 1.6 | |
| JOR-10 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 48 | 1.2 | Mort |
| JOR-11 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 56 | 1.7 | Mort |
| JOR-12 | Jordan Creek | Jordan Creek | Aug-1-2019 | TR | MT | 79 | 7.2 | Mort |
| JOR-13 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 53 | 1.8 | Mort |
| JOR-14 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 51 | 1.5 | Mort |
| JOR-15 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | MT | 47 | 1.1 | Mort |
| JOR-16 | Jordan Creek | Jordan Creek | Aug-1-2019 | TR | EF | 128 | 21.6 | pic 116-125 |
| JOR-17 | Jordan Creek | Jordan Creek | Aug-1-2019 | CC | EF | 83 | 7.4 | |
| JOR-18 | Jordan Creek | Jordan Creek | Aug-1-2019 | CC | EF | 74 | 4.8 | |
| JOR-19 | Jordan Creek | Jordan Creek | Aug-1-2019 | CC | EF | 58 | 2.2 | |
| JOR-20 | Jordan Creek | Jordan Creek | Aug-1-2019 | SB | EF | 59 | 2 | pic 126 |
| JOR-21 | Jordan Creek | Jordan Creek | Aug-1-2019 | CC | EF | 62 | 2.3 | |

Appendix E

Site Data for Coastal Tailed Frog Surveys

| Valley Side | Site | Date | Easting | Northing | Elev. (m) | Weather | Water Temp. (°C) | Air Temp. (°C) | Channel Width (m) | Wetted Width (m) | Discharge | Mean Depth (cm) | Stream Disturbance | Stream Morph. |
|-------------|-------------------------------------|------------|---------|----------|-----------|----------|------------------|----------------|-------------------|------------------|-----------|-----------------|--------------------|---------------|
| East | Archibald Creek - 1 | 2019-09-04 | 502387 | 5550606 | 695 | Sunny | 11.4 | 18.0 | 4.0 | 2.4 | Med | 12 | Med. | Step Pool |
| East | Archibald Creek - 2 | 2019-09-04 | 502854 | 5550298 | 835 | Sunny | 11.2 | 15.0 | 2.7 | 2.2 | Med | 18 | Med. | Step Pool |
| East | Archibald Creek - 3 | 2019-09-04 | 503310 | 5549422 | 1026 | Sunny | 9.4 | 17.0 | 2.2 | 0.9 | Med | 8 | Low | Step Pool (|
| East | Blackcomb Creek @ Yummy Numby | 2019-09-06 | 505211 | 5552576 | 762 | Sunny | 8.0 | 11.0 | 8.4 | 5.0 | Med | 16 | Low | Cascade (5 |
| West | FJ Unnamed | 2019-09-05 | 496157 | 5548481 | 699 | Cloudy | 11.0 | 18.0 | 8.0 | 2.0 | Med | 12 | High | Step Pool |
| West | FJ West Creek - 1 (South Flank) | 2019-09-05 | 496383 | 5548374 | 648 | Cloudy | 11.2 | 18.0 | 4.1 | 1.5 | Med | 12 | High | Step Pool |
| West | FJ West Creek - 3 (Into the Mystic) | 2019-09-03 | 496022 | 5549522 | 1119 | Sunny | 11.3 | 14.0 | 2.2 | 1.2 | Med | 10 | Low | Cascade (5 |
| West | Sproatt Creek - 1 (Danimal South) | 2019-09-03 | 499063 | 5549434 | 692 | Lt. Rain | 12.9 | 16.0 | 6.6 | 2.2 | Med | 9 | Low | Riffle (Step |
| West | Sproatt Creek - 2 (Don't Look Back) | 2019-09-03 | 498996 | 5549662 | 790 | Lt. Rain | 12.3 | 17.0 | 7.8 | 1.5 | Med | 10 | High | Riffle (Step |
| West | Sproatt Creek - 3 (Flank Trail) | 2019-09-03 | 498483 | 5550455 | 996 | Sunny | 12.0 | 15.0 | 5.0 | 0.8 | Med | 12 | High | Step Pool |
| West | Van West - 1 (Flank Trail) | 2019-09-05 | 497563 | 5549038 | 706 | Sunny | 12.5 | 16.0 | 5.1 | 1.3 | Low | 12 | High | Step Pool |
| West | Van West - 3 (Into the Mystic) | 2019-09-03 | 497125 | `` | 1036 | Sunny | 11.7 | 14.5 | 4.2 | 1.5 | Low | 12 | Med. | Step Pool |
| East | Whistler Creek - 1 | 2019-09-06 | 501041 | 5549045 | 692 | Sunny | 11.0 | 11.0 | 6.2 | 6.0 | Med | 12 | Med. | Step Pool |
| East | Whistler Creek - 2 | 2019-09-05 | 501417 | 5548276 | 879 | Sunny | 10.0 | 11.0 | 5.1 | 2.1 | Med | 18 | Low | Riffle (Step |
| East | Whistler Creek - 3 | 2019-09-05 | 501649 | 5547961 | 972 | Sunny | 10.2 | 11.0 | 4.1 | 2.3 | Med | 15 | Low | Step Pool |

| Valley Side | Site | Date | Easting | Northing | Rock Size | Rock Shape | Slope (%) | Crown Closure | Tree Comp. | Struct. Stage |
|----------------|-------------------------------------|------------|---------|----------|-------------------|------------|-----------|------------------|---------------|------------------|
| East | Archibald Creek - 1 | 2019-09-04 | 502387 | 5550606 | Bedrock (Boulder) | Subrounded | 17 | 75 | Decid. | Pole/Sapl. |
| East | Archibald Creek - 2 | 2019-09-04 | 502854 | 5550298 | Cobble (Boulder) | Subangular | 18 | 80 | Mixed | YF |
| East | Archibald Creek - 3 | 2019-09-04 | 503310 | 5549422 | Cobble (Boulder) | Subangular | 12 | 95 | Conif. | YF |
| East | Blackcomb Creek @ Yummy Numby | 2019-09-06 | 505211 | 5552576 | Boulder (Cobble) | Subrounded | 15 | 60 | Conif. | OF |
| West | FJ Unnamed | 2019-09-05 | 496157 | 5548481 | Boulder (Cobble) | Subangular | 10 | 25 | Mixed | YF/MF |
| West | FJ West Creek - 1 (South Flank) | 2019-09-05 | 496383 | 5548374 | Cobble (Bedrock) | Subangular | 14 | 80 | Mixed | YF |
| West | FJ West Creek - 3 (Into the Mystic) | 2019-09-03 | 496022 | 5549522 | Bedrock (Cobble) | Subrounded | 14 | 30 | Conif. | OF |
| West | Sproatt Creek - 1 (Danimal South) | 2019-09-03 | 499063 | 5549434 | Boulder (Cobble) | Subangular | 25 | 30 | Mixed | Shrub/MF |
| West | Sproatt Creek - 2 (Don't Look Back) | 2019-09-03 | 498996 | 5549662 | Boulder (Cobble) | Subrounded | 32 | 50 | Conif. | OF |
| West | Sproatt Creek - 3 (Flank Trail) | 2019-09-03 | 498483 | 5550455 | Boulder (Bedrock) | Subrounded | 24 | 40 | Conif. | MF |
| West | Van West - 1 (Flank Trail) | 2019-09-05 | 497563 | 5549038 | Boulder (Bedrock) | Subangular | 18 | 95 | Conif. | YF |
| West | Van West - 3 (Into the Mystic) | 2019-09-03 | 497125 | `` | Cobble (Boulder) | Subangular | 25 | 50 | Conif. | OF |
| East | Whistler Creek - 1 | 2019-09-06 | 501041 | 5549045 | Cobble (Boulder) | Subangular | 14 | 5 | Decid. | Shrub |
| East | Whistler Creek - 2 | 2019-09-05 | 501417 | 5548276 | Cobble (Boulder) | Subangular | 14 | 10 | Conif. | OF |
| East | Whistler Creek - 3 | 2019-09-05 | 501649 | 5547961 | Cobble (Bedrock) | Subangular | 25 | 40 | Conif. | OF |

Appendix F

Capture Results Coastal Tailed Frogs

| Valley Side | Site | Elev. (m) | Tadpoles detected | | | | | Mets/ Adults | Water Temp. | Survey Area |
|-------------|-------------------------------------|-----------|-------------------|----|----|-------|-----------|--------------|-------------|-------------|
| | | | T1 | T2 | T3 | Total | per 100m2 | | (°C) | (m2) |
| East | Archibald Creek - 1 | 695 | 0 | 4 | 1 | 5 | 15.4 | 0 | 11.4 | 32.5 |
| East | Archibald Creek - 2 | 835 | 0 | 0 | 1 | 1 | 6.9 | 0 | 11.2 | 14.5 |
| East | Archibald Creek - 3 | 1026 | 8 | 0 | 0 | 8 | 59.3 | 0 | 9.4 | 13.5 |
| East | Blackcomb Creek @ Yummy Numby | 762 | 0 | 0 | 0 | 0 | 0.0 | 0 | 8.0 | 23.5 |
| West | FJ Unnamed | 699 | 0 | 1 | 0 | 1 | 4.3 | 0 | 11.0 | 23.5 |
| West | FJ West Creek - 1 (South Flank) | 648 | 0 | 0 | 0 | 0 | 0.0 | 0 | 11.2 | 15.5 |
| West | FJ West Creek - 3 (Into the Mystic) | 1119 | 1 | 0 | 0 | 1 | 6.1 | 0 | 11.3 | 16.5 |
| West | Sproatt Creek - 1 (Danimal South) | 692 | 0 | 0 | 1 | 1 | 8.3 | 0 | 12.9 | 12.0 |
| West | Sproatt Creek - 2 (Don't Look Back) | 790 | 1 | 0 | 0 | 1 | 8.0 | 0 | 12.3 | 12.5 |
| West | Sproatt Creek - 3 (Flank Trail) | 996 | 4 | 0 | 5 | 9 | 38.3 | 0 | 12.0 | 23.5 |
| West | Van West - 1 (Flank Trail) | 706 | 0 | 0 | 0 | 0 | 0.0 | 0 | 12.5 | 12.0 |
| West | Van West - 3 (Into the Mystic) | 1036 | 2 | 1 | 3 | 6 | 46.2 | 0 | 11.7 | 13.0 |
| East | Whistler Creek - 1 | 692 | 1 | 1 | 8 | 10 | 45.0 | 0 | 11.0 | 22.2 |
| East | Whistler Creek - 2 | 879 | 9 | 5 | 1 | 15 | 90.9 | 0 | 10.0 | 16.5 |
| East | Whistler Creek - 3 | 972 | 0 | 2 | 0 | 2 | 15.2 | 0 | 10.2 | 13.2 |

Length (mm) of tadpoles captured by site

| Age Class / Cohort (Malt et al. 2014a, b) | T1 | | T2 | | T3 | Total tadpoles | Total metamorphs & adults |
|---|--------------|-----------------------------------|-------------------------------|----------------------|-----------------------|----------------|---------------------------|
| | No hind legs | Bulge only, hind legs not defined | Hind legs visible but covered | Hind feet protruding | Hind knees protruding | | |
| Archibald Creek - 1 | | | 44 | 45 | 50 | 5 | |
| | | | | 45 | | | |
| | | | | 45 | | | |
| Archibald Creek - 2 | | | | | 50 | 1 | |
| Archibald Creek - 3 | 30 | | | | | 8 | |
| | 30 | | | | | | |
| | 30 | | | | | | |
| | 30 | | | | | | |
| | 26 | | | | | | |
| | 26 | | | | | | |
| | 26 | | | | | | |
| Blackcomb Creek @ Yummy Numby | | | | | | 0 | |
| FJ Unnamed | | | 50 | | | 1 | |
| FJ West Creek - 1 (South Flank) | | | | | | 0 | |
| FJ West Creek - 3 (Into the Mystic) | | 35 | | | | 1 | |
| Sproatt Creek - 1 (Danimal South) | | | | | 60 | 1 | |
| Sproatt Creek - 2 (Don't Look Back) | | 40 | | | | 1 | |
| Sproatt Creek - 3 (Flank Trail) | 35 | 36 | | | 48 | 17 | |
| | 35 | | | | 51 | | |
| | 35 | | | | 49 | | |
| | | | | | 51 | | |

| Age Class / Cohort (Malt et al. 2014a, b) | T1 | | T2 | | T3 | Total tadpoles | Total metamorphs & adults |
|---|--------------|-----------------------------------|-------------------------------|----------------------|-----------------------|----------------|---------------------------|
| Developmental Stage | No hind legs | Bulge only, hind legs not defined | Hind legs visible but covered | Hind feet protruding | Hind knees protruding | | |
| | | | | | 50 | | |
| Van West - 1 (Flank Trail) | | | | | | 0 | |
| Van West - 3 (Into the Mystic) | 35 | 38 | | 45 | 48 | 6 | |
| | | | | | 47 | | |
| | | | | | 50 | | |
| Whistler Creek - 1 | | 36 | 43 | | 51 | 10 | |
| | | | | | 45 | | |
| | | | | | 45 | | |
| | | | | | 58 | | |
| | | | | | 47 | | |
| | | | | | 47 | | |
| | | | | | 47 | | |
| Whistler Creek - 2 | 30 | 36 | | 42 | 50 | 15 | |
| | 30 | 40 | | 45 | | | |
| | 30 | 44 | | 45 | | | |
| | 43 | | | 48 | | | |
| | 39 | | | 43 | | | |
| | 35 | | | | | | |
| Whistler Creek - 3 | | | | 44 | | 2 | |
| | | | | 50 | | | |

Notes: Numbers in red indicate tadpoles that were seen but not captured - lengths are therefore estimates.
Surveyors included Bob Brett (all sites), Jagoda Kozikowska (13 sites), and Hillary Williamson (2 sites)

Appendix G

RMOW EMP 2019 Goose
Hawks

| Location | Date | Easting | Northing | Elev. (m) | Record | Observer(s) | Source | Notes |
|-------------------------------------|------------|---------|----------|--------------|--------|---------------------------|-------------|---|
| Blackcomb Alpine | 2000-03-14 | 507070 | 5549311 | 1867 | Visual | B Max Götz | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2001-03-03 | 501773 | 5552539 | 643 | Visual | B Max Götz | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2007-06-02 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2008-02-02 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Blackcomb Alpine | 2009-02-14 | 507070 | 5549311 | 1867 | Visual | Peter Dunwiddie | eBird | Only info via eBird |
| Whistler Village and vicinity | 2009-08-22 | 503156 | 5551541 | 683 | Visual | Daniel Airola | eBird | Only info via eBird |
| Whistler Golf Club | 2011-08-06 | 502208 | 5551354 | 684 | Visual | Christopher Di Corrado | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2011-08-15 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Fitzsimmons Fan & Nicklaus North GC | 2011-11-02 | 503656 | 5554556 | 636 | Visual | Chris Dale | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2011-11-05 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2012-02-13 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2012-05-05 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2013-03-02 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2013-03-14 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2013-05-04 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2014-08-02 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2014-12-06 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Lost Lake and vicinity | 2015-03-15 | 504636 | 5552716 | 687 | Visual | Cole Gaerber | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2015-07-04 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Blackcomb Alpine | 2016-03-12 | 507070 | 5549311 | 1867 | Visual | Nina Rach | eBird | Only info via eBird |
| Valley Trail to Rainbow Beach | 2016-05-07 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | Only info via eBird |
| Millar's Pond | 2016-06-06 | 499601 | 5548228 | 727 | Visual | Bob Brett | RMOW EMP | heard alarm calls (video) |
| Millar's Pond | 2016-06-09 | 499601 | 5548228 | 727 | Visual | Bob Brett | RMOW EMP | video of single goshawk defending nest |
| Callaghan Valley Road | 2016-06-10 | 490798 | 5549818 | 679 | Visual | BBS Team | eBird | watched NOGO on nest from 15:00 to 15:20 with George Clulow, Rob Lyske, David, and David Aldcrof |
| Millar's Pond | 2016-06-12 | 499601 | 5548228 | 727 | Visual | Bob Brett | RMOW EMP | photo of bird on nest |
| Valley Trail to Rainbow Beach | 2016-07-02 | 501773 | 5552539 | 643 | Visual | C. Dale, H. Baines et al. | eBird | |
| Musical Bumps Trail | 2016-09-12 | 504873 | 5543244 | 1907 | Visual | Bob Brett | WBP | photo of goshawk flying overhead |
| Millar's Pond | 2016-09-20 | 499601 | 5548228 | 727 | Visual | Bob Brett | RMOW EMP | Glimpsed and heard a goshawk flying near nest. |
| Millar's Pond | 2016-09-20 | 499601 | 5548228 | 727 | Visual | Bob Brett | RMOW EMP | Glimpsed (poor view of bird flying uphill), uphill of road (inside stand). |
| Whistler Village and vicinity | 2016-11-30 | 503156 | 5551541 | 683 | Visual | Daniel Tinoco | eBird | |
| 5302 Alta Lake Rd. | 2017-06-21 | 500162 | 5550088 | 690 | Visual | C Palmer | eBird | |
| Decker Trail | 2017-07-31 | 508618 | 5546519 | 1918 | Visual | Dan Wilson | WBP | Est. date. Confirmed by George Clulow & Rob Lyske. |

| Location | Date | Easting | Northing | Elev. (m) | Record | Observer(s) | Source | Notes |
|--------------------------------------|------------|---------|----------|--------------|------------|---------------------------|-------------|---|
| Bayshores, ~500m e of active nest | 2017-10-09 | 500005 | 5543876 | 671 | Visual | Dave McPeake | WBP | video of goshawk on house's deck |
| Westside Road (unspecified) | 2018-04-14 | 499982 | 5550268 | 742 | Visual | Christa Vandenberg | RMOW EMP | via H. Beresford; Conf. by George Clulow & Rob Lyske. |
| Alta Lake Road n. of Wildlife Refuge | 2018-05-01 | 501524 | 5553719 | 685 | Visual | Bob Brett | WBP | Backlit bird perched on power line. Large imm. or adult. |
| Lost Lake and vicinity | 2018-06-09 | 504636 | 5552716 | 687 | Visual | Mike Farnworth | eBird | |
| Callaghan Valley Road | 2018-06-15 | 490798 | 5549818 | 679 | Visual | BBS Team | WNS | |
| Kadenwood 2018 FireSmart site | 2018-10-02 | 500291 | 5548095 | 756 | Visual | Bob Brett | WBP | Moderate confidence, blurred view in forest |
| Kadenwood 2018 FireSmart site | 2018-10-10 | 500386 | 5548095 | 870 | Visual | Leo Coudrau | WBP | High confidence, good view flying through forest |
| Near Emerald Forest south gravel pit | 2019-01-05 | 501730 | 5552795 | 644 | Visual | C. Dale, H. Baines et al. | WNS | Single bird in area likely hunting abundant Pine Siskins |
| Lost Lake (beach area) | 2019-04-25 | 504629 | 5552704 | 694 | Visual (?) | Jagoda Kozikowska | verbal | Jagoda saw a bird in the a.m. from the beach she thinks was a goshawk. |
| Chateau Golf Course, n. of hole #8 | 2019-05-25 | 504431 | 5553657 | 739 | Visual (?) | Dan Nash | WBP | Dan is pretty sure he saw a goshawk flying. |
| Whistler Olympic Park | 2019-06-22 | 491761 | 5554069 | 851 | Visual | Paul Maury | eBird | 2 sightings, same location, on eBird, May 22 & 25 |
| Baxter Creek, Rainbow Housing | 2019-07-01 | 503086 | 5556357 | 725 | Visual (?) | Scott Aitken | verbal | Saw a goshawk drinking from Baxter Creek (date is approximate). |
| Kadenwood Drive | 2019-07-01 | 500168 | 5548864 | 633 | Visual (?) | Arthur De Jong | verbal | Twice saw a mid-sized, grey raptor flying low overhead of road. |
| Lost Lake disc golf, hole 21 | 2019-07-14 | 503973 | 5553968 | 693 | Visual (?) | Bob Brett | personal | Saw a backlit bird shaped like a goshawk. |
| Sarajevo Drive, Creekside | 2019-08-01 | 500615 | 5548650 | 741 | Visual | Unknown (via Shawn Mason) | WNS | "She told me there is a trail off Sarajevo that she jogs on through the forest." |
| Kill Me Thrill Me vicinity | 2019-08-06 | 506279 | 5557196 | 634 | Visual (?) | Dan Raymond | email | "I think I saw a goshawk yesterday above kill me thrill me. It was soaring around making shrill screeches, then seemed to dive at a raven (then away)." |
| Powderwood condos, Whistler Road | 2019-12-14 | 501356 | 5549526 | 732 | Visual | Elizabeth Barrett | email | Liz saw the bird in her backyard |

| Location | Date | Easting | Northing | Elev. (m) | Record | Observer(s) | Source | Notes |
|------------------------------------|------------|---------|----------|--------------|----------|------------------------------|-------------|--|
| Brew Creek | 2015-07-24 | 490637 | 5545029 | 829 | Audible | T. Tripp, C. Churchland | MFLNRO | "3 ot 4 very clear juvenile begging calls." UTM is on east side of Brandywine Creek, i.e., far from 2011 active nest. |
| Comfortably Numb w. of Wedge Creek | 2019-07-20 | 506935 | 5555480 | 829 | Auditory | Trystan Willmott, Bob Brett | Brett 2020 | 1, possible 2, juveniles responded to begging call near 2014 nest |
| Comfortably Numb w. of Wedge Creek | 2014-06-30 | 506935 | 5555480 | 829 | Nest | Pablo Jost, Naomi Sands | BC MOE | Active nest with 3 juveniles |
| Comfortably Numb @ Jeff's Trail | 2015-07-24 | 506387 | 5555458 | 823 | Nest | T. Tripp, C. Churchland | MFLNRO | Active nest found in existing territory. |
| Millar's Pond | 2016-05-20 | 499601 | 5548228 | 727 | Nest | Brent Matsuda | RMOW EMP | found active nest in 73cm Douglas-fir, ~14 above ground, SE (uphill) aspect, in branch crotch; found three feathers, one was sent for DNA test |
| Millar's Pond | 2016-06-12 | 499601 | 5548228 | 727 | Nest | B. Brett, G. Clulow & others | WNS | watched active nest from 15:00-15:20 |
| Millar's Pond | 2017-06-03 | 499601 | 5548228 | 727 | Nest | B. Matsuda & Mike Tootchin | WBP | One goshawk responded to playback. 3 fledglings (M. Wilson, pers. comm.) |

Appendix H

Timing and Duration of Ice on
Alta Lake, 1942-1976 and
2001-2019

| Winter | Ice-On | | Ice-Off | | Days Frozen |
|---|------------|-----------|------------|-----------|-------------|
| | Date | Day Count | Date | Day Count | |
| 1942/43 | 1942-12-04 | 338 | 1943-04-19 | 109 | 136 |
| 1943/44 | 1943-12-15 | 349 | 1944-04-13 | 104 | 120 |
| 1944/45 | 1944-12-15 | 350 | 1945-04-27 | 117 | 133 |
| 1945/46 | 1945-11-08 | 312 | 1946-04-20 | 110 | 163 |
| 1946/47 | 1946-11-20 | 324 | 1947-04-13 | 103 | 144 |
| 1947/48 | 1947-12-11 | 345 | 1948-05-07 | 128 | 148 |
| 1948/49 | 1948-12-18 | 353 | 1949-04-19 | 109 | 122 |
| 1949/50 | 1949-12-14 | 348 | 1950-04-24 | 114 | 131 |
| 1950/51 | 1950-12-02 | 336 | 1951-04-19 | 109 | 138 |
| 1951/52 | 1951-12-13 | 347 | 1952-05-21 | 142 | 160 |
| 1952/53 | 1952-12-22 | 357 | 1953-05-08 | 128 | 137 |
| 1953/54 | 1954-01-10 | 375 | 1954-05-05 | 125 | 115 |
| 1954/55 | 1954-12-26 | 360 | 1955-05-07 | 127 | 132 |
| 1955/56 | 1955-12-18 | 352 | No Data | N/A | N/A |
| 1956/57 | 1956-12-01 | 336 | 1957-04-23 | 113 | 143 |
| 1957/58 | 1957-12-26 | 360 | 1958-04-08 | 98 | 103 |
| 1958/59 | 1958-11-26 | 330 | 1959-04-23 | 113 | 148 |
| 1959/60 | 1959-12-05 | 339 | 1960-04-16 | 107 | 133 |
| 1960/61 | 1960-12-10 | 345 | 1961-04-10 | 100 | 121 |
| 1961/62 | 1961-12-01 | 335 | 1962-04-09 | 99 | 129 |
| 1962/63 | No Data | N/A | 1963-03-23 | 82 | N/A |
| 1963/64 | 1963-12-13 | 347 | 1964-04-24 | 115 | 133 |
| 1964/65 | 1964-12-11 | 346 | 1965-04-22 | 112 | 132 |
| 1965/66 | 1965-12-12 | 346 | 1966-04-21 | 111 | 130 |
| 1966/67 | No Data | N/A | 1967-04-30 | 120 | N/A |
| 1967/68 | 1967-12-12 | 346 | 1968-04-27 | 118 | 137 |
| 1968/69 | 1968-12-05 | 340 | 1969-05-07 | 127 | 153 |
| 1969/70 | 1970-01-15 | 380 | 1970-04-06 | 96 | 81 |
| 1970/71 | 1970-12-04 | 338 | 1971-05-06 | 126 | 153 |
| 1971/72 | 1971-12-14 | 348 | 1972-05-02 | 123 | 140 |
| 1972/73 | 1972-12-28 | 363 | 1973-04-11 | 101 | 104 |
| 1973/74 | 1973-11-24 | 328 | 1974-04-28 | 118 | 155 |
| 1974/75 | No Data | N/A | No Data | N/A | N/A |
| 1975/76 | 1975-12-12 | 346 | No Data | N/A | N/A |
| Note: Data was not recorded between the fall 1975 freeze-up and the spring 2002 thaw. | | | | | |
| 2001/02 | No Data | N/A | 2002-04-14 | 104 | N/A |
| 2002/03 | No Data | N/A | 2003-03-17 | 76 | N/A |
| 2003/04 | No Data | N/A | 2004-03-25 | 85 | N/A |
| 2004/05 | No Data | N/A | No Data | N/A | N/A |
| 2005/06 | 2006-01-06 | 371 | 2006-03-08 | 67 | 61 |
| 2006/07 | 2006-11-30 | 334 | 2007-04-10 | 100 | 131 |
| 2007/08 | 2007-12-10 | 344 | 2008-04-29 | 120 | 141 |
| 2008/09 | 2008-12-20 | 355 | 2009-04-28 | 118 | 129 |

| Winter | Ice-On | | Ice-Off | | Days Frozen |
|---------|------------|-----------|------------|-----------|-------------|
| | Date | Day Count | Date | Day Count | |
| 2009/10 | 2009-12-08 | 342 | 2010-03-28 | 87 | 110 |
| 2010/11 | 2010-12-04 | 338 | 2011-04-23 | 113 | 140 |
| 2011/12 | No Data | N/A | 2012-04-23 | 114 | N/A |
| 2012/13 | 2012-12-16 | 351 | 2013-04-03 | 93 | 108 |
| 2013/14 | 2013-12-21 | 355 | 2014-04-14 | 104 | 114 |
| 2014/15 | 2014-12-26 | 360 | 2015-02-20 | 51 | 56 |
| 2015/16 | 2015-12-24 | 358 | 2016-03-16 | 76 | 83 |
| 2016/17 | No Data | N/A | 2017-04-24 | 114 | N/A |
| 2017/18 | No Data | N/A | 2018-04-10 | 100 | N/A |
| 2018/19 | 2019-01-01 | 366 | 2019-04-12 | 102 | 101 |